



Arklow Bank Wind Park 2

Environmental Impact Assessment Report

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Arklow Bank Wind Park 2: Underwater Noise Assessment (**Revised March 2026**)

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Executive Summary

Subacoustech Environmental, on behalf of GoBe Consultants, has undertaken a study in order to assess the potential underwater noise and its effects during impact piling at Arklow Bank Wind Park 2 (ABWP2). This study follows from the previous underwater noise modelling undertaken as part of the Environmental Impact Assessment Report (EIAR). This revision takes into account the Further Information Request [doc no. ABP-319864-24].

Impact piling is required at ABWP2 to install foundations for wind turbine generators (WTG) and offshore platforms (OSP). Modelling of underwater noise generated by impact piling was undertaken at five representative locations, with the loudest levels predicted at the Central location for the installation of WTG foundations, primarily due to the deeper water at that location.

The modelling results were analysed in terms of relevant noise metrics and criteria to assess the effects of impact piling noise on marine mammals and fish, which have been used to aid biological assessments. For marine mammals the largest auditory injury (including permanent threshold shift (PTS)) onset ranges were predicted for LF cetaceans, which includes minke whale, with maximum impact ranges out to 7.1 km. For fish, the largest recoverable injury ranges were predicted to be 6.0 km for a stationary receptor, reducing to less than 50 m when assuming a moving receptor.

As part of the assessment, a Mitigated piling ramp-up scenario has been modelled to show the potential effects of modifying the durations, strike rate and blow energies on the Precautionary ramp-up scenario. These modifications resulted in maximum PTS ranges of 870 m for LF cetaceans, and maximum fish recoverable injury ranges of 5.5 km for a stationary receptor.

Additionally, the effect of physical mitigation, in the form of a low-noise hammer (or other system capable of reducing noise levels by 4 dB), has also been considered on the Precautionary ramp-up scenario. When the noise reduction from a noise abatement system is considered, PTS onset ranges for LF cetaceans are reduced to 720 m, and recoverable injury ranges for stationary fish are expected to be reduced to 3.4 km.

A review was undertaken of various underwater noise requirements or limits for offshore wind farm installation in use in the European Union (EU), as requested in the Further Information Request item 9 a ii. Of those considered, ABWP2 believes the Danish requirements, as noted in the GOMOREUS report (Tougaard *et al.*, 2025) for the Marine Institute, offer the best combination of consideration of the varying sensitivities of different species groups expected to be present in Irish waters, the site specific characteristics in comparison to Danish waters (primarily the depths and varying bathymetry, and as it includes the total noise produced by an impact piling event. The Danish requirements introduce the concept of r_{safe} , which is the distance within which a receptor species (typically marine mammals, although in principle this could include any species with available impact threshold) would receive sufficient noise exposure to exceed a PTS threshold. ABWP2 intends to use an r_{safe} target of 1 km, and would ensure marine mammals would not be present in this zone at the start of piling. This is typically through the use of Marine Mammal Observers and/or Acoustic Deterrent Devices. Modelling shows that this target can be achieved using either the low noise hammer (or other technique offering 4 dB noise reduction) with Precautionary ramp-up scenario, and the Mitigated ramp up scenario, without additional noise abatement.

Noise sources other than impact piling were considered using higher-level methodologies, and these included cable laying, dredging, drilling, rock placement, trenching, vessel noise, and operational WTG noise. All these sources were predicted to have a much smaller impact compared to impact piling noise.

Noise from unexploded ordnance (UXO) clearance showed there is a risk of PTS onset out to 990 m with use of the expected low-order UXO clearance technique. This considered the unweighted peak criteria for the very

high-frequency (VHF) cetaceans hearing group, which includes harbour porpoise. In the event that a high-order detonation does occur, the maximum PTS onset range is predicted to be 14 km from detonation of the largest considered device (800 kg + donor charge), using the same VHF cetacean criteria. It should be noted that this is likely to be highly precautionary as the impact range is based on a precautionary criterion and calculation methodology that does not account for any smoothing of the pulse over long ranges, which would reduce the pulse peak and other characteristics of the sound that cause injury. High-order detonation is not a proposed UXO clearance methodology.

It should be stressed that, due to the nature of modelling, while the results present specific ranges at which each impact threshold is met, the ranges should be taken as indicative and worst case in determining where environmental effects may occur in receptors during the proposed operations.

The outputs of this modelling have been used to inform analysis of the impacts of underwater noise on marine mammals and fish in their respective reports.

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Acronyms

Acronym	Definition
ABWP2	Arklow Bank Wind Park 2
ADD	Acoustic Deterrent Device
BGS	British Geological Survey
EIA	Environmental Impact Assessment
EIAR	Environmental Impact Assessment Report
EMODnet	European Marine Observation and Data Network
FPSO	Floating Production Storage and Offloading
GIS	Geographic Information System
HF	High-Frequency Cetaceans (Marine mammal hearing group from Southall <i>et al.</i> , 2019)
INSPIRE	Impulse Noise Sound Propagation and Range Estimator (Subacoustech Environmental's noise model for estimating impact piling noise)
LF	Low-Frequency Cetaceans (Marine mammal hearing group from Southall <i>et al.</i> , 2019)
MTD	Marine Technology Directorate
NMFS	National Marine Fisheries Service
NOAA	National Oceanic and Atmospheric Administration
NPL	National Physical Laboratory
OSP	Offshore Platform
PCW	Phocid Carnivores in Water (Marine mammal hearing group from Southall <i>et al.</i> , 2019)
PPV	Peak Particle Velocity
PTS	Permanent Threshold Shift
RMS	Root Mean Square
SE	Sound Exposure
SEL	Sound Exposure Level
SEL _{cum}	Cumulative Sound Exposure Level
SEL _{ss}	Single Strike Sound Exposure Level
SPL	Sound Pressure Level
SPL _{peak}	Peak Sound Pressure Level
SPL _{peak-to-peak}	Peak-to-peak Sound Pressure Level
SPL _{RMS}	Root Mean Square Sound Pressure Level
TNT	Trinitrotoluene (explosive)
TTS	Temporary Threshold Shift
UXO	Unexploded Ordnance
VHF	Very High-Frequency Cetaceans (Marine mammal hearing group from Southall <i>et al.</i> , 2019)
WTG	Wind Turbine Generator

Technical Glossary Terminology

Term	Definition
Decibel (dB)	A customary scale commonly used (in various ways) for reporting levels of sound. A difference of 10 dB corresponds to a factor of 10 in sound power. The actual The dB represents a ratio/comparison of a sound measurement is compared to (e.g., sound pressure) over a fixed reference level and the “decibel” value is defined to be $10 \log_{10}(\text{actual/reference})$ where $(\text{actual/reference})$ is a power ratio. Because sound power is usually proportional to sound pressure squared, the decibel value for sound pressure is $20 \log_{10}(\text{actual/reference pressure})$. The standard reference for underwater sound is 1 micro pascal (µPa). The dB symbol is followed by a second symbol identifying the specific reference value (e.g., re 1 µPa).
Peak pressure	The highest pressure above or below ambient that is associated with a sound wave.
Peak-to-peak pressure	The sum of the highest positive and negative pressures that are associated with a sound wave.
Permanent Threshold Shift (PTS)	Noise threshold that represents the onset level of a permanent total or partial loss of impairment hearing caused by acoustic trauma. PTS results in irreversible damage to the sensory hair cells of the air ear, and thus a permanent reduction of hearing acuity.
Root Mean Square (RMS)	The square root of the arithmetic average of a set of squared instantaneous values. Used for presentation of an average sound pressure level.
r_{safe}	As per Danish guidance, the distance at which a marine mammal must be at the start of piling to avoid PTS noise exposure
Sound Exposure Level (SEL or L_{eq})	The constant sound level acting for one second, which has the same amount of acoustic energy, as indicated by the square of the sound pressure, as the original sound. It is the time-integrated, sound-pressure-squared level. SEL is typically used to compare transient sound events having different time durations, pressure levels, and temporal characteristics.
Sound Exposure Level, cumulative (SEL _{cum} or L_{Cum})	Single value for the collected, combined total of sound exposure over a specified time or multiple instances of a noise source.
Sound Exposure Level, single strike (SEL _{ss} or L_{Cum})	Calculation of the sound exposure level representative of a single noise impulse, typically a pile strike.
Sound Pressure Level (SPL or L_p)	The sound pressure level is an expression of sound pressure using the decibel (dB) scale; the standard frequency pressures of which are 1 µPa for water and 20 µPa for air.
Sound Pressure Level Peak (SPL _{peak} , $L_{\text{p,pt}}$ or $L_{\text{p,p}}$)	The highest (zero-peak) positive or negative sound pressure, in decibels.
Temporary Threshold Shift (TTS)	Onset threshold level for a temporary reduction of hearing acuity because of caused by exposure to sound over time. Exposure to high levels of sound over relatively short time periods could cause the same level of TTS as exposure to lower levels of sound over

	<p>longer time periods. The mechanisms underlying TTS are not well understood, but there may be some temporary damage to the sensory cells. The duration of TTS varies depending on the nature of the stimulus.</p>
<p>Unweighted sound level</p>	<p>Sound levels which are “raw” or have not been adjusted in any way, for example to account for the hearing ability of a species.</p>
<p>Weighted sound level</p>	<p>A sound level which has been adjusted with respect to a “auditory weighting function” or “weighting envelope” in the frequency domain, typically to make an unweighted level relevant to a particular species. Examples of this are the dB(A), where the overall sound level has been adjusted to account for the hearing ability of humans in air, or the filters used by Southall et al. (2019) for marine mammals.</p>

Acronyms

Acronym	Definition
ABP	An Bord Pleanála
ABWP2	Arklow Bank Wind Park 2
ACP	An Coimisiún Pleanála
ADD	Acoustic Deterrent Device
BGS	British Geological Survey
BSH	Bundesamt für Seeschifffahrt und Hydrographie
Cefas	Centre for the Environment, Fisheries and Aquaculture Science
DAHG	Department of the Arts, Heritage and the Gaeltacht, now DHLGH (see below)
Defra	Department of Environment, Food and Rural Affairs
DHLGH	Department of Housing, Local Government and Heritage
EIAR	Environmental Impact Assessment Report
EMODnet	European Marine Observation and Data Network
EPS	European Protected Species
EU	European Union
FPSO	Floating Production Storage and Offloading (vessel)
GES	Good Environmental Status
GIS	Geographic Information System
GOMOREUS	Guidance on Managing Offshore Renewable Energy Underwater Sound
HE	High Explosive
HF	High-Frequency Cetaceans
INSPIRE	Impulsive Noise Sound Propagation and Impact Range Estimator
ISO	International Organisation for Standardisation
JNCC	Joint Nature Conservation Committee
LF	Low-Frequency Cetaceans
LOBE	Level of Onset of Biologically Significant Effect (MSFD)
MMO	Marine Management Organisation
MSFD	Marine Strategy Framework Directive
MTD	Marine Technology Directorate
NAS	Noise Abatement System
NE	Natural England
NEQ	Net Explosive Quantity
NMFS	National Marine Fisheries Service

NDA&	National Oceanic and Atmospheric Administration
NPL	National Physical Laboratory
NPWS	National Parks and Wildlife Service
OSP	Offshore Platform
OWF	Offshore Wind Farm
PCW	Phocid Carnivores in Water
PPV	Peak Particle Velocity
PTS	Permanent Threshold Shift
RMS	Root Mean Square
SE	Sound Exposure
SEL ($L_{E,d}$)	Sound Exposure Level
SEL _{cum} ($L_{E,d,t}$)	Cumulative Sound Exposure Level
SEL _{iso} ($L_{E,iso}$)	Single Strike Sound Exposure Level
SPL	Sound Pressure Level
SPL _{peak} ($L_{p,pt}$)	Peak Sound Pressure Level
SPL _{rms} (L_p)	Root Mean Square Sound Pressure Level
TNT	Trinitrotoluene (explosive)
TTS	Temporary Threshold Shift
UK	United Kingdom
UXO	Unexploded Ordnance
VHF	Very High-Frequency Cetaceans
WTG	Wind Turbine Generator

Units

Unit	Definition
dBbl/min	Decibel (sound pressure) Blows per minute (piling strike rate)
GW dB	Gigawatt (power) Decibel (sound pressure)
Hz	Hertz (frequency)
kg	Kilogram (mass)
kJ Hz	Kilojoule (energy) Kilohertz (frequency)
kHz J	Kilohertz (frequency) Kilojoule (energy)
km	Kilometre (distance)
km ²	Square kilometres (area)
kn	Knot (speed)
kW	Kilowatt (power)
m	Metre (distance)
mm/s ms ⁻¹	Millimetres per second (particle velocity)
m/s ms ⁻¹	Metres per second (speed)
MW	Megawatt (power)
Pa	Pascal (pressure)
Pa ² s	Pascal squared seconds (acoustic energy)
s	Seconds (duration)
μPa	Micropascal (pressure)

1 Introduction

Arklow Bank Wind Park 2 (ABWP2) is a ~~proposed offshore windfarm in the southern Irish Sea. As part of the Environmental Impact Assessment Report (EIAR) process,~~ planned offshore wind farm situated on and around Arklow Bank in the Irish Sea, approximately 6 to 15 km to the east of Arklow in County Wicklow, Ireland.

Subacoustech Environmental Ltd. has undertaken detailed modelling and analysis in relation to ~~the effect of~~ underwater noise ~~on marine mammals and fish~~ from impact piling in order to install foundations for wind turbine generators (WTG) and offshore platforms (OSP).

Following from the initial underwater noise assessment undertaken as part of the EIAR (Appendix 11.1), this report is a revision of that assessment utilises the latest version of the modelling software along with updated input parameters in order to ascertain the potential effects of impact piling on marine mammals and fish during construction of ABWP2.

The array ~~Area for ABWP2~~ covers an area of 63.4 km² and ~~is situated 6 to 15 km from the Wicklow coast.~~ ~~The proposed development~~ has a proposed capacity of 800 ~~Megawatt (MW)~~, utilising either 47 or ~~56~~53 wind turbine generators (~~WTGs~~)WTG) depending on the final Design Option chosen. The location of ABWP2 is shown in Figure 1-1.

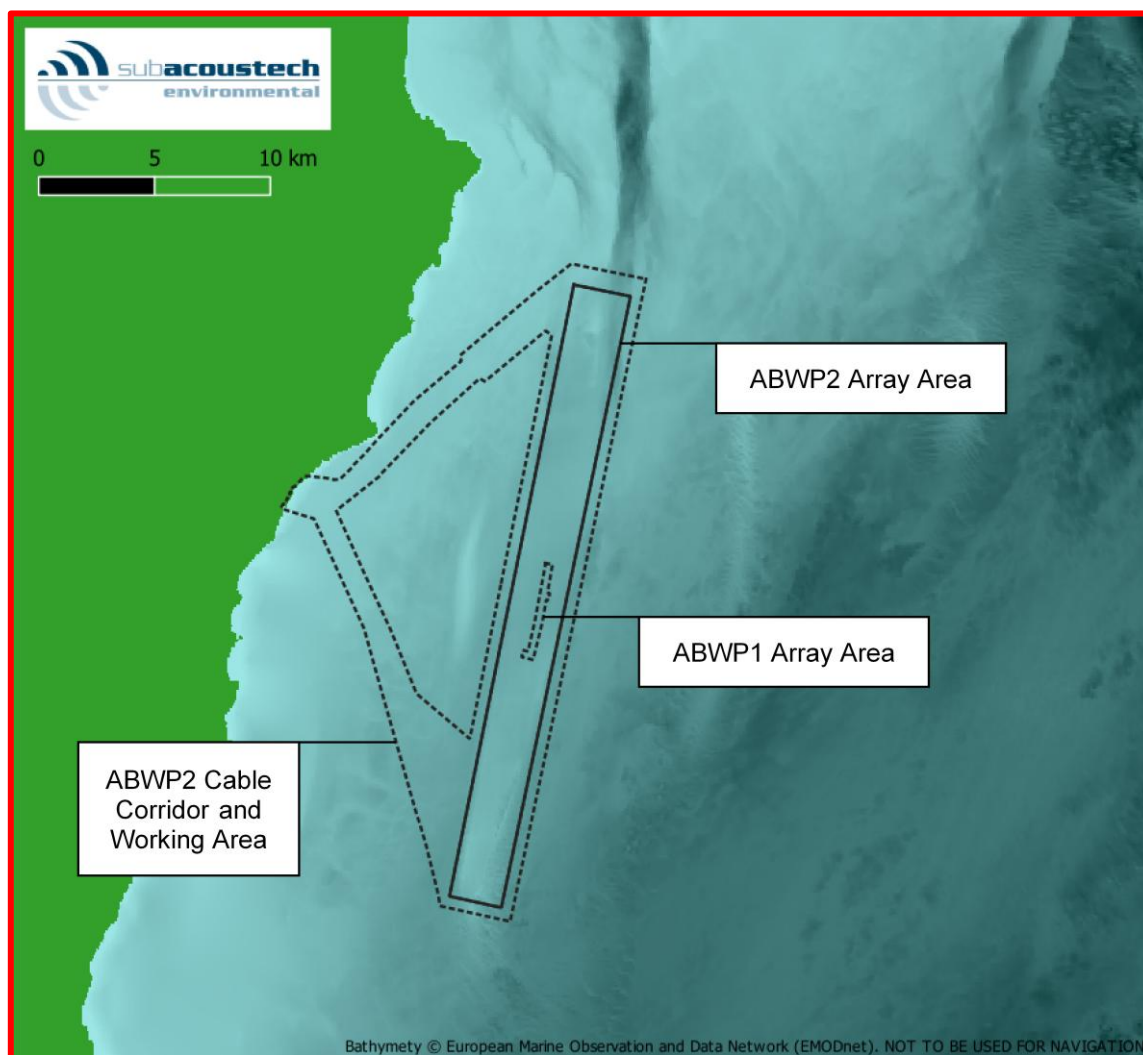


Figure 1-1 Overview map showing the ABWP2 Boundary, Cable Corridor and Working Area, and the surrounding bathymetry and coastline

This **technical** report presents a detailed assessment of the potential underwater noise **from impact piling activities** during **the construction and operation** of ABWP2, and includes the following:

- Background information covering the units for measuring and assessing underwater noise, and a review of the underwater noise metrics and criteria used to assess the possible environmental effects in marine receptors (section 2).
- Discussion of the approach, **confidence**, input parameters, and **assumptions and** assumptions for the detailed **impact piling** modelling undertaken (section 3).
- Presentation and interpretation of the detailed subsea noise modelling for impact piling with regards to its effect on marine mammals and fish, **including the use of a noise attenuation system (NAS)** (section 4).
- **Noise** Modelling of other noise sources expected around the construction and operation of ABWP2 including cable laying, dredging, drilling, rock placement, vessel movements, operational WTG noise, and unexploded ordnance (UXO) clearance (section 5); **and**.
- Summary and conclusions (section 6).

Further modelling results covering the noise from the first pile strike and for non-impulsive thresholds (see sections 2.2.1 and 2.3.1) are presented in [Appendix A](#).

This technical report is an update to the previous Arklow Bank Wind Park underwater noise report issued as part of the EIA in 2024, including updated modelling and consideration of Further Information Requests from ACP.

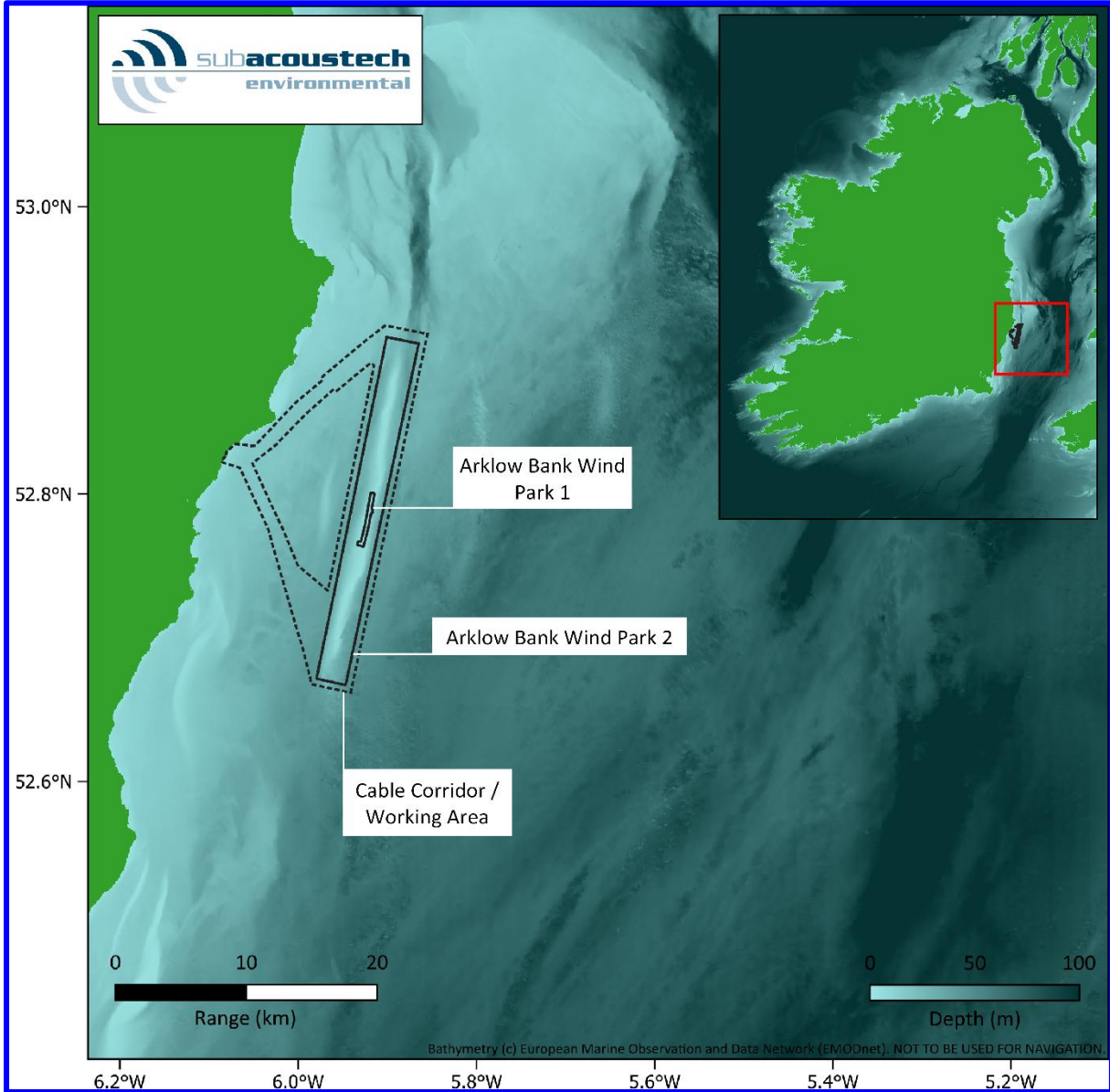


Figure 1-1: Overview map showing the ABWP2 boundary, its location next to the Irish coast, and the surrounding bathymetry.

2 ~~Background to~~ Underwater noise concepts

2.1 ~~Underwater noise~~

Sound travels much faster in water (approximately 1,500 m/sms^{-1}) than in air (340 m/sms^{-1}). ~~Since~~ as water is a relatively incompressible, ~~dense medium, the pressure associated with~~ and has a higher density than air. This affects the way in which sound measurements are expressed between the two mediums, which means that underwater sound ~~tends to be much higher than in air.~~ levels are not directly comparable to airborne sound levels. This is noted for context; this report does not contain or include any reference to airborne sound.

~~Underwater noise levels should not be confused with noise levels in air, which use a different scale.~~

2.1 ~~2.1.1~~ Units of measurement

Sound measurements underwater are usually expressed using the decibel (dB) scale, which is a logarithmic measure of sound. A logarithmic scale is used ~~because, rather than~~ this better reflects how sound is perceived. For example, equal increments of sound ~~having~~ pressure do not have an equal increase in ~~effect,~~ ~~typically each~~ the perceived sound. Instead, a doubling of sound ~~level~~ pressure will cause a roughly equal increase of “perceived loudness each time.”

Any quantity expressed in this dB scale is termed a “level.” For example, if the unit is sound pressure, ~~expressed on the dB scale,~~ it will be termed a “sound pressure level.” ~~on the dB scale.~~

The fundamental definition of the dB scale is given by:

$$\text{Level} = 10 \times \log_{10} \left(\frac{Q}{Q_{ref}} \right)$$

where Q is the quantity being expressed on the scale, and Q_{ref} is the reference quantity.

The dB scale represents a ratio. It is therefore used with a reference unit, which expresses the base from which the ratio is expressed. The reference quantity is conventionally smaller than the smallest value to be expressed on the scale so that any level quoted is positive. For example, a reference quantity of 20 μPa is used for sound in air since that is the lower threshold of human hearing.

When used with sound pressure, the pressure value is squared. So that variations in the units agree, the sound pressure must be specified as units of Root Mean Square (RMS) pressure squared. This is equivalent to expressing the sound as:

$$\text{Sound pressure level } (L_p) = 20 \times \log_{10} \left(\frac{P_{RMS}}{P_{ref}} \right)$$

For underwater sound, a unit of 1 μPa is typically used as the reference unit (P_{ref}); a Pascal is equal to the pressure exerted by one Newton over one square metre, one micropascal equals one millionth of this.

2.1.1 ~~2.1.2~~ Sound pressure level (SPL)

~~The~~ Sound Pressure Level (SPL or L_p) is ~~normally used to characterise noise of a continuous nature, such as drilling, boring, continuous wave sonar, or background sea and river noise levels. To calculate the SPL, the variation in sound pressure is measured over a specific period to determine the RMS level of the time-varying sound. The SPL can therefore be considered a measure of the average unweighted level of sound over the measurement period.~~ a measure of the pressure variation caused by sound waves,

expressed in decibels (dB), as seen in the equations above. Variations of L_p are used depending on the noise source being measured. Unless otherwise defined, all L_p noise levels in this report are referenced to 1 μPa .

2.1.1.1 Mean squared sound pressure

For continuous, non-impulsive noise sources such as drilling or vibropiling, an unweighted sound pressure level, averaged over a measurement period, known as a root mean squared (RMS) sound pressure level (SPL_{RMS} or $L_{p,\text{RMS}}$), can be used to represent the noise levels. The RMS period must be specified (e.g. $L_{p,\text{RMS},125\text{ms}}$), as the mean level can vary significantly depending on the measurement duration.

~~Where SPL is used to characterise transient pressure waves, such as that from impact piling, seismic airgun or underwater blasting, it is critical that the period over which the RMS level is calculated is quoted. For instance, in the case of a pile strike lasting a tenth of a second, the mean taken over a tenth of a second will be ten times higher than the mean averaged over one second. Often, transient sounds such as these are quantified using "peak" SPLs or Sound Exposure Levels (SELs).~~

~~Unless otherwise defined, all SPL noise levels in this report are referenced to 1 μPa .~~

2.1.1.2 ~~2.1.3~~ Peak sound pressure Level (SPL_{peak})

~~Peak SPLs are often used to characterise Transient sound from, impulsive sources, pressure waves such as percussive generated from impact piling-- are usually expressed using the level of the peak sound pressure (SPL_{peak} or $L_{p,\text{pk}}$). This is calculated using the maximum pressure variation of the pressure from positive to zero within the wave. This represents the maximum, representing the peak change in positive pressure (differential pressure from positive to zero) as the transient pressure-wave propagates.~~

~~A further variation of this is the peak-to-peak SPL ($\text{SPL}_{\text{peak-to-peak}}$) where the maximum variation of the pressure from positive to negative is considered. Where the wave is symmetrically distributed in positive and negative pressure, the peak-to-peak pressure will be twice the peak level, or 6 dB higher (see section 2.1).~~

2.1.2 ~~2.1.4~~ Sound exposure level (SEL)

When considering the noise from transient sources, the issue of the duration of the pressure wave is often addressed by measuring the total acoustic energy (energy flux density) of the wave. This form of analysis was used by Bebb and Wright (1953, 1954a, 1954b, 1955), and later by Rawlins (1987), to explain the apparent discrepancies in the biological effect of short and long-range blast waves on human divers. More recently, this form of analysis has been used to develop criteria for assessing injury ranges for fish and marine mammals from various noise sources (e.g., Popper *et al.*, 2014; Southall *et al.*, 2019; Southall *et al.*, 2007).

The $\text{SEL}_{L_{E,p}}$ sums the acoustic energy over a measurement period, and effectively takes account of both the SPL_{L_p} of the sound and the duration it is present in the acoustic environment. Sound Exposure (SE) is defined by the equation:

$$SE = \int_0^T p^2(t) dt$$

where p is the acoustic pressure in Pascals, T is the total duration of sound in seconds, and t is time in seconds. The SE is a measurement of acoustic energy and has units of Pascal squared seconds (Pa^2s).

To express the SE on a logarithmic scale by means of a dB, it must be compared with a reference acoustic energy ($P_{\text{ref}}^2 T_{\text{ref}}$) and a reference time (T_{ref}). The $\text{SEL}_{L_{E,p}}$ is then defined by:

$$L_{E,p} = 10 \times \log_{10} \left(\frac{\int_0^T p^2(t) dt}{P_{\text{ref}}^2 T_{\text{ref}}} \right)$$

By using a common reference pressure P_{ref} of 1 μPa for assessments of underwater noise, the SEL_{LE} and SPL_{Lp} can be compared using the expression:

$$L_{E,p} = L_p + 10 \times \log_{10} T$$

where the SPL_{Lp} is a measure of the average level of broadband noise and the $SEL_{LE,p}$ sums the cumulative broadband noise energy.

This means that, for continuous sounds of less than (i.e., fractions of) one second, the $SEL_{LE,p}$ will be lower than the SPL_{Lp} . For periods greater than one second, the $SEL_{LE,p}$ will be numerically greater than the SPL_{Lp} (i.e., for a continuous sound of 10 seconds duration, the $SEL_{LE,p}$ will be 10 dB higher than the SPL_{Lp} ; for a sound of 100 seconds duration the $SEL_{LE,p}$ will be 20 dB higher than the SPL_{Lp} , and so on).

Where a single impulse noise such as the soundwave from a pile strike is considered in isolation, this can be represented by a "single strike" $SEL_{LE,p}$ or SEL_{ss} . A cumulative $SEL_{LE,p}$, or SEL_{cum} , accounts for the exposure from multiple impulses or pile strikes over time, where the number of impulses replaces the T in the equation above, leading to:

$$Cumulative L_{E,p} = L_{E,p} + 10 \times \log_{10} X$$

Where $SEL_{LE,p}$ is the sound exposure level of one impulse and X is the total number of impulses or strikes.

Unless otherwise defined, all $SEL_{LE,p}$ noise levels in this report are referenced to 1 $\mu\text{Pa}^2\text{s}$.

2.2 Properties of sound

2.2.1 Impulsive and non-impulsive sound

Sound can be loosely categorised into two types: impulsive noise and non-impulsive noise. Non-impulsive noise can be defined as a steady-state noise which does not necessarily have a long duration (e.g., vibropiling, drilling). Impulsive noise can be defined as sound with a high peak sound pressure, short duration, fast rise-time and a broad frequency content at the source (e.g., impact piling, explosives, seismic airguns).

These differences are important when considering the potential for auditory injury, as impulsive noise is more injurious than non-impulsive noise (Henderson and Hamernik, 1986; Hastie *et al.*, 2019).

Due to the differences between impulsive and non-impulsive noise sources, different metrics are appropriate, for example:

- Impulsive noises: peak SPL ($L_{p,pk}$) and cumulative SEL ($L_{E,p,t}$).
- Non-impulsive noises: cumulative SEL ($L_{E,p,t}$) and SPL_{RMS} ($L_{p,RMS}$).

Objective categorisation of a noise as impulsive or non-impulsive is not always clear. This is particularly the case if sound is travelling over large distances. For example, as an impulsive sound propagates through an environment, the energy within the sound wave will scatter and dissipate, and it will become less impulsive with distance. This is important to consider regarding auditory injury and impact range calculations, as noise will become less injurious if it becomes less impulsive.

Research to define a range-dependent transition from impulsive to non-impulsive noise has been a significant field of study (see, for example, Martin *et al.*, 2020). Although the situation is complex, Hastie *et al.* (2019) concluded that an impulsive sound can be considered effectively non-impulsive at a range of 3.5 km from the source using some metrics. However, the recent study by Matei *et al.* (2024) concludes that there is still insufficient evidence to clearly define a transition point suitable for an assessment such as this. It is, however, reasonable to presume there is a fully impulsive region close to the source, and a fully non-impulsive region at some greater distance, and a transition region in between. The paper makes it clear that there is a substantial

reduction in impulsiveness within the first 5 km. However, due to the uncertainty in identifying a transition point, no presumption of a change has been made in this report, although it is reasonable to assume that the sound can be considered not fully impulsive where PTS onset ranges (see section 2.3.1 for marine mammals) are calculated above 5 km. Results in respect of both impulsive and non-impulsive criteria have been presented for impact piling noise in this report.

2.2.2 Particle motion

The motion of the particles that make up a medium is an important component of sound. Particle motion is present wherever there is sound, and it describes the back-and-forth movement of particles in water, which in the context of underwater noise, are caused by a sound wave passing through the water column. This back-and-forth movement means that, unlike sound pressure at a single point, particle motion always contains directional information (Hawkins and Popper, 2017). Regarding quantifying particle motion, it is usually defined in reference to the velocity of the particle (often as a peak particle velocity, PPV), but sometimes the related acceleration or displacement of the particle is used.

It has been identified by several researchers that many fish species (e.g., Popper and Hawkins, 2019; Nedelec *et al.*, 2016; Radford *et al.*, 2012), as well as marine invertebrates (see section 2.3.3) are sensitive to particle motion. However, sound pressure metrics are still preferred and more widely used than particle motion due to a lack of supporting data (Popper and Hawkins, 2018). There continue to be calls for additional research on the levels of and effects on marine receptors with respect to levels of particle motion.

2.3 ~~2.2~~ Analysis of environmental effects: Assessment criteria

Over the last 20 years it has become increasingly evident that noise from human activities in and around underwater environments can have an impact on the marine species in the area. The extent to which intense underwater sound might cause adverse impacts in species is dependent upon the incident sound level, source frequency, duration of exposure, and/or repetition rate of an impulsive sound (see, for example, Hastings and Popper, 2005). As a result, scientific interest in the hearing abilities of aquatic species has increased. Studies are primarily based on evidence from high level sources of underwater noise such as ~~blasting or~~ seismic airguns, impact piling, ~~and blasting~~ as these sources are likely to have the greatest immediate environmental impact and therefore the clearest observable effects, although interest in chronic noise exposure is increasing.

The impacts of underwater sound on marine species can be broadly summarised as follows:

- Physical traumatic injury and fatality;
- Auditory injury (either permanent or temporary); ~~and~~.
- ~~Disturbance~~ Behavioural responses.

The following sections discuss the underwater noise criteria used in this study with respect to species of marine mammals and fish that may be present around ~~ABWP2~~ the study area.

The main metrics and criteria that have been used in this study to aid assessment of environmental effects come from ~~three~~ two key papers covering underwater noise and its effects:

- Southall *et al.* (2019) marine mammal exposure criteria; ~~and~~.
- Popper *et al.* (2014) sound exposure guidelines for fishes and sea turtles.

~~At the time of writing these include relevant and authoritative criteria for assessing environmental effects for use in impact assessments.~~

2.3.1 ~~2.2.1~~ Marine mammals

The Southall *et al.* (2019) paper is the most used and recognised reference for marine mammal hearing thresholds at the time of writing this report. It provides identical ~~thresholds~~ threshold to those from the National Marine Fisheries Service (NMFS) (2018) guidance for marine mammals ~~(although it names~~. It should be noted that, despite the identical thresholds, the marine mammal ~~categories~~ hearing groups are described slightly differently). ~~It updates and supersedes the methodology from the previous in the~~ Southall *et al.* (2007/2019) paper to the NMFS (2018) guidance. Therefore, care should be taken when comparing results using the Southall *et al.* (2019) and NMFS (2018) criteria.

The Southall *et al.* (2019) guidance ~~groups~~ categorises marine mammals into ~~categories~~ groups of similar species and applies filters to the unweighted noise to approximate the hearing sensitivities of the receptor in question. The hearing groups given by Southall *et al.* (2019) are summarised in Table 2-1 and the relevant auditory weighting functions are shown in Figure 2-1. Further groups for sirenians and other marine carnivores in water are given, but these have not been included in this study as those species are not commonly found in ~~the Irish Sea~~ our study area.

It should be noted that despite Southall *et al.* (2019) referring to SPL_{peak} and cumulative SEL as SEL_{cum} , this report notation has since been updated (ISO 18405: 2017) and will be referred to as $L_{p,pk}$ instead of SPL_{peak} , and $L_{E,p,t}$ instead of SEL_{cum} in the rest of this report.

Table 2-1: Marine mammal hearing groups (from Southall et al., 2019).

Hearing group	Generalised hearing range	Example species
Low-frequency cetaceans (LF)	7 Hz to 35 kHz	Baleen whales (including minke whales)
High-frequency cetaceans (HF)	150 Hz to 160 kHz	Dolphins, toothed whales, beaked whales, bottlenose whales (including bottlenose dolphin)
Very high-frequency cetaceans (VHF)	275 Hz to 160 kHz	True porpoises (including harbour porpoise)
Phocid carnivores in water (PCW)	50 Hz to 86 kHz	True seals (including harbour seals)

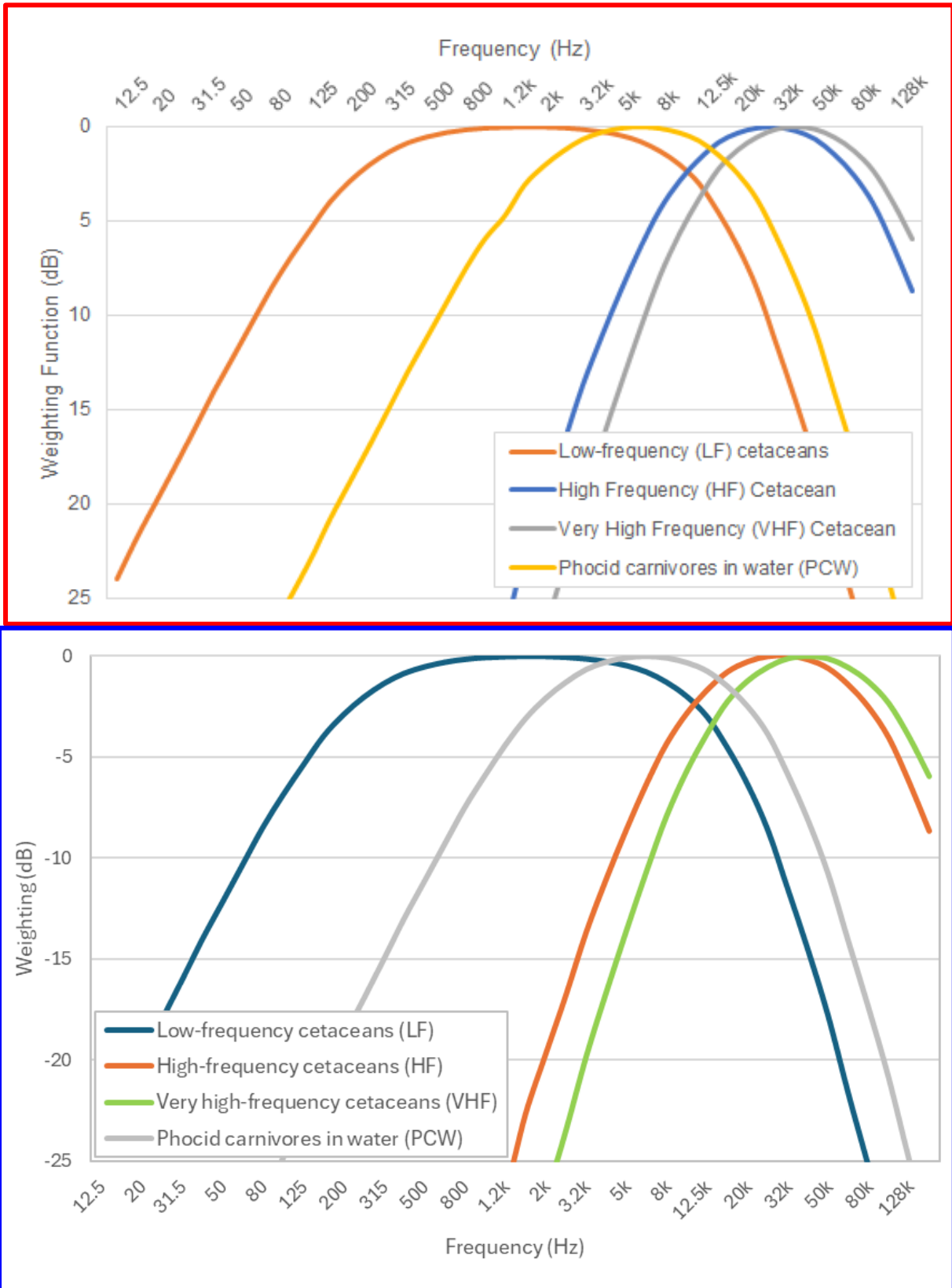


Figure 2-1: Auditory weighting functions for low-frequency cetaceans (LF), high-frequency cetaceans (HF), very high-frequency cetaceans (VHF), and phocid carnivores in water (PCW) (from Southall et al., 2019).

Southall *et al.* (2019) ~~also gives individual criteria based on~~ considers the nature of the sound in the context of whether ~~the noise source is considered~~ it is an impulsive or non-impulsive. ~~Southall *et al.* (2019) categorises impulsive noises as having high peak sound pressure, short duration, fast rise time and broad frequency content at source, and non-impulsive sources as steady-state noise. Explosives, impact piling and seismic airguns are considered impulsive noise sources and sonars, vibro-piling, drilling and other low-level continuous noises are considered non-impulsive. A non-impulsive noise does not necessarily have to have a long duration.~~ noise source (see section 2.2.1 for details). Although the use of impact ranges derived using the impulsive criteria are recommended for all but clearly defined non-impulsive sources, it should be recognised that where calculated ranges are beyond 5 km (see section 2.2.1; Matei *et al.*, 2024), the sound is expected to be beyond the fully impulsive region and the real impact range is likely to be somewhere between the impulsive and non-impulsive impact criteria. Therefore, if the modelled impact range of an impulsive noise has been predicted to be greater than 5 km, the non-impulsive impact range should also be considered. Both impulsive and non-impulsive criteria have been presented in this study.

~~Southall *et al.* (2019) presents single strike, unweighted peak criteria (SPL_{peak}) and cumulative weighted sound exposure criteria (SEL_{cum} , i.e., can include the accumulated exposure of multiple pulses) for both permanent threshold shift (PTS), where unrecoverable (but incremental) hearing damage may occur, and temporary threshold shift (TTS), where a temporary reduction in hearing sensitivity may occur in individual receptors. These dual criteria (SPL_{peak} and SEL_{cum}) are only used for impulsive noise: the criteria set giving the greatest calculated range is typically used as the relevant impact range.~~

~~As sound pulses propagate through the environment and dissipate, they also lose their most injurious characteristics (e.g., rapid pulse rise time and high peak sound pressure) and become more like a “non-pulse” at greater distances; Southall *et al.* (2019) briefly discusses this. Active research is currently underway into the identification of the distance at which the pulse can be considered effectively non-impulsive, and Hastie *et al.* (2019) have analysed a series of impulsive data to investigate it. Although the situation is complex, the paper reported that most of the signals crossed their threshold for rapid rise time and high peak sound pressure characteristics associated with impulsive noise at around 3.5 km from the source. Southall (2021) discusses this further and suggests that the impulsive characteristics can correspond with significant energy content of the pulse above 10 kHz. This will naturally change depending on the noise source and the environment over which it travels. To provide as much detail as possible, both impulsive and non-impulsive criteria from Southall *et al.* (2019) have been included in this study.~~

~~Although the use of impact ranges derived using the impulsive criteria are recommended for all but the clearly non-impulsive sources (such as drilling), it should be recognised that where calculated ranges are beyond 3.5 km, they would be expected to become increasingly less impulsive and harmful, and the impact range is therefore likely to be somewhere between the modelled impulsive and non-impulsive impact range. Where the impulsive impact range is significantly greater than 3.5 km, the non-impulsive range should be considered.~~

Table 2-2 and Table 2-3 present the ~~unweighted SPL_{peak} and weighted SEL_{cum}~~ impulsive and non-impulsive criteria ~~for marine mammals from~~ set out by Southall *et al.* (2019) ~~covering both impulsive and non-impulsive noise~~ for permanent threshold shift (PTS) and temporary threshold shift (TTS) in marine mammals used in this study.

Table 2-2 ~~Single strike SPL_{peak} : Unweighted $L_{p,pk}$ criteria for PTS and TTS in marine mammals (Southall et al., 2019).~~

Southall et al. (2019) Unweighted $L_{p,pk}$ (dB re 1 μ Pa)	Unweighted SPL_{peak} (dB re 1 μ Pa) impulsive	
	Impulsive	
	PTS	TTS
Low-frequency cetaceans (LF)	219	213
High-frequency high frequency cetaceans (HF)	230	224
Very high-frequency cetaceans (VHF)	202	196
Phocid carnivores in water (PCW)	218	212

Table 2-3 ~~Impulsive and non-impulsive SEL_{cum} : Weighted $L_{E,p,24h,wtd}$ criteria for PTS and TTS in marine mammals (Southall et al., 2019).~~

Southall et al. (2019) Weighted $L_{E,p,24h,wtd}$ (dB re 1 μ Pa ² s)	Weighted SEL_{cum} (dB re 1 μ Pa ² s)			
	Impulsive		Non-impulsive	
	PTS	TTS	PTS	TTS
Low-frequency cetaceans (LF)	183	168	199	179
High-frequency high frequency cetaceans (HF)	185	170	198	178
Very high-frequency cetaceans (VHF)	155	140	173	153
Phocid carnivores in water (PCW)	185	170	201	181

Where SEL_{cum} thresholds are required for marine mammals, a fleeing animal model has been used for marine mammals. This assumes that a receptor, when exposed to high noise levels, will swim away from the noise source. A constant fleeing speed of 3.25 m/s has been assumed for the low-frequency cetaceans (LF) group (Blix and Folkow, 1995), based on data for minke whale, and for other receptors, a constant rate of 1.5 m/s has been assumed for flee speed fleeing, which is a cruising speed (i.e., sustainable long-term) for a harbour porpoise (Otani et al., 2000). These are considered worst case assumptions as marine mammals are expected to be able to swim much faster under stress conditions (Kastelein et al. 2018), especially at the start of any noisy process when the receptor will be closest to the noise source.

The fleeing animal model, and assumptions related to it, are discussed in more detail in section 3.3.

Limited data is available for behavioural disturbance on species of marine mammal. To take this into account, the National Oceanic and Atmospheric Administration (NOAA) (2005) Level B (behavioural disturbance) harassment criterion for impulsive noise on marine mammals has been included to cover disturbance effects. This criterion is 160 dB unweighted Root Mean Square Sound Pressure Level (SPL_{RMS}) re 1 μ Pa ($L_{p,RMS}$) from a single strike.

In addition, the concept of r_{safe} , following the Danish regulations (Danish Energy Agency, 2023), has been utilised. This derives a distance from the pile that would represent a distance that, provided a marine mammal is beyond at the start of piling, it would be safe from the effects of PTS according to the Southall et al. (2019) guidance. This distance is known as “ r_{safe} ” and is effectively the distance that must be cleared before piling starts to avoid auditory injury for marine mammals. This is also one of the options presented in the Guidance On Managing Offshore Renewable Energy Underwater Sound (GOMOREUS) study for underwater sound in Irish waters (Tougaard et al., 2025) and is used as the basis of this assessment. The Danish, and other EU, guidance currently in use is discussed further in section 2.4, in accordance with the Further Information Request from An Coimisiún Pleanála (ACP) (RFI 9a ii).

2.3.2 ~~2.2.2~~ Fish

~~The large number of, and variation in, fish species leads to a greater challenge in production of a generic noise criterion, or range of criteria, for the assessment of noise impacts. Whereas previous studies applied broad criteria based on limited studies of fish that are not present in UK waters (e.g., McCauley et al., 2000) or measurement data not intended to be used as criteria (Hawkins et al., 2014), the publication of Popper et al. (2014) provides an authoritative summary of the latest research and guidelines for fish exposure to sound and uses categories for fish that are representative of the species present in UK and Irish waters.~~

The Popper et al. (2014) guidelines are recognised as a suitable reference for underwater noise impacts on marine fauna (aside from marine mammals) in UK and Irish waters. Popper et al. (2014) provides a summary of research and guidelines for fish (and other marine fauna) exposure to sound and uses categories that a representative of the species present around ABWP2.

~~The Popper et al. (2014) study groups species of fish by whether they possess a swim bladder, and whether it is involved in its hearing; groups for sea turtles and fish eggs and larvae are also included. The guidance also gives specific criteria (as both unweighted SPL_{Lpeak} and unweighted SEL_{Lcum} values) for a variety of noise sources. (It is recognised that these are related to sound pressure, whereas more recent documents (e.g., Popper and Hawkins, 2019) clearly state that many fish species are most sensitive to particle motion. This is discussed in section 2.3.3.)~~

guidelines present criteria dependent on the type of noise source, species of marine fauna and their hearing capabilities, and impact type. Noise sources considered in the guidance include explosions, pile driving, seismic airguns, sonar, and shipping and continuous noise. For this study, the criteria for ~~impact piling, pile driving, explosions, and continuous noise sources, and explosions~~ have been ~~considered; these are summarised in Table 2-4 Criteria for mortality and potential mortal injury, recoverable injury, and TTS in species of fish from impact piling noise (Popper et al., 2014) to Table 2-6 Criteria for potential mortal injury in species of fish from explosions (Popper et al., 2014) used.~~

For each sound source, the marine fauna are categorised into groups covering fish, sea turtles and eggs and larvae. Due to their diversity and quantity, fish are categorised further into three groups depending on their hearing capabilities, which can be indicated by whether they possess a swim bladder or not, and whether the swim bladder is involved in its hearing.

~~Table 2-4 Criteria for mortality and potential mortal injury, recoverable injury, and TTS in species of fish from impact piling noise (Popper et al., 2014)~~

Type of animal	Mortality and potential mortal injury	Impairment	
		Recoverable injury	TTS
Fish: no swim bladder	> 219 dB SEL _{Lcum} > 213 dB SPL _{Lpeak}	> 216 dB SEL _{Lcum} > 213 dB SPL _{Lpeak}	>> 186 dB SEL _{Lcum}
Fish: swim bladder is not involved in hearing	210 dB SEL _{Lcum} > 207 dB SPL _{Lpeak}	203 dB SEL _{Lcum} > 207 dB SPL _{Lpeak}	> 186 dB SEL _{Lcum}
Fish: swim bladder involved in hearing	207 dB SEL _{Lcum} > 207 dB SPL _{Lpeak}	203 dB SEL _{Lcum} > 207 dB SPL _{Lpeak}	186 dB SEL _{Lcum}
Sea turtles	> 210 dB SEL _{Lcum} > 207 dB SPL _{Lpeak}	See Table 2-7	
Eggs and larvae	> 210 dB SEL _{Lcum} > 207 dB SPL _{Lpeak}		

Popper et al. (2014) provides separate criteria, depending on the species and noise source, for various impacts associated with noise exposure. These are mortality and potential mortal injury, impairment (split into recoverable injury, TTS, and masking), and behavioural effects.

~~Table 2-5 Criteria for recoverable injury and TTS in species of fish from continuous noise sources (Popper et al., 2014)~~

Type of animal	Impairment	
	Recoverable injury	TTS
Fish: swim bladder involved in hearing	170 dB SPL _{RMS} for 48 hrs	158 dB SPL _{RMS} for 12 hours

~~Table 2-6 Criteria for potential mortal injury in species of fish from explosions (Popper et al., 2014)~~

Type of animal	Mortality and potential mortal injury
Fish: no swim bladder	229 – 234 dB SPL _{peak}
Fish: swim bladder is not involved in hearing	229 – 234 dB SPL _{peak}
Fish: swim bladder involved in hearing	229 – 234 dB SPL _{peak}
Sea turtles	229 – 234 dB SPL _{peak}
Eggs and larvae	> 13 mm/s peak velocity

Depending on the noise source, quantitative criteria are given in appropriate metrics ($L_{p,pk}$, $L_{E,p,24h}$, L_p), which can then be used as thresholds for the onsets of listed impacts. Where insufficient data ~~are~~ is available, Popper *et al.* (2014) also gives ~~qualitative criteria that summarise~~ a description of relative risk. This summarises the effect of the noise as having either a high, moderate, or low relative risk of an effect on an individual in either ~~the near-field~~ near (tens of metres), ~~intermediate-field~~ intermediate (hundreds of metres), or ~~far-field~~ far (thousands of metres). ~~These qualitative effects are reproduced in~~ Table 2-7 to Table 2-9. from the source.

Where $L_{E,p,t}$ thresholds are required for fish, both a stationary and a fleeing animal model has been used. Most species described by Popper *et al.* (2014) are likely to be able to move away from a sound that is loud enough to cause harm (Dahl *et al.*, 2015; Popper *et al.*, 2014). For those species that can swim away, a speed of 1.5 ms⁻¹ (based on Hirata, 1999) has been considered as a conservative speed at which to base a fleeing animal model. However, considering the diversity of species described by Popper *et al.* (2014), whether an animal flees or remains stationary in response to a loud noise will differ between species. It is recognised that there is limited evidence for fish fleeing from high level noise sources in the wild (Hubert *et al.*, 2024). The species that are likely to remain stationary are though more likely to be benthic species of species without a swim bladder, due to their reduced hearing capabilities making these species least sensitive to noise (Goertner *et al.*, 1994; Goertner *et al.*, 1978; Stephenson *et al.*, 2010; Halvorsen *et al.*, 2012).

Hubert *et al.* (2024) noted that pelagic fish did not clearly flee with exposure to sound, albeit at sound pressure levels far lower than piling noise, and did not rule out the possibility that a flee response could occur at higher levels. Despite this, only including results for a stationary animal as a worst-case scenario is likely to greatly overestimate the potential risk to fish species. As such, a combined approach is recommended, which considers impact ranges from both fleeing and stationary receptors.

The thresholds and relative risk descriptions given by Popper *et al.* (2014) used in this study are reproduced in ~~Table 2-7 Summary of the qualitative effects on species of fish from impact piling noise (4 to Table 2-9 Summary of the qualitative effects on species of fish from explosions (6.~~ Similar to the Southall *et al.* (2019) criteria in section 2.3.1, the Popper *et al.* (2014) criteria use the SPL_{peak} and SEL_{cum} notation, and this report will present the ISO 18405:2017 notation ($L_{p,pk}$ and $L_{E,p,t}$ respectively) for consistency.

Table 2-7 Summary of the qualitative effects on species of fish from impact piling noise (4: Recommended guidelines for pile driving according to Popper et al. (2014) for species of fish, sea turtles and eggs and larvae (N = near-field; I = intermediate-field; F = far-field).

Type of animal receptor	Mortality and potential mortal injury	Impairment			Behaviour
		Recoverable injury	TTS	Masking	
Fish: no swim bladder	> 219 dB $L_{E,p,24h}$ > 213 dB $L_{p,pk}$	See Table 2-4 Criteria for mortality and potential mortal injury, recoverable injury, and TTS in species of fish from impact piling noise (Popper et al., 2014) > 216 dB $L_{E,p,24h}$ > 213 dB $L_{p,pk}$	>> 186 dB $L_{E,p,24h}$	(N) Moderate (I) Low (F) Low	(N) High (I) Moderate (F) Low
Fish: swim bladder is not involved in hearing	210 dB $L_{E,p,24h}$ > 207 dB $L_{p,pk}$	203 dB $L_{E,p,24h}$ > 207 dB $L_{p,pk}$	> 186 dB $L_{E,p,24h}$	(N) Moderate (I) Low (F) Low	(N) High (I) Moderate (F) Low
Fish: swim bladder involved in hearing	207 dB $L_{E,p,24h}$ > 207 dB $L_{p,pk}$	203 dB $L_{E,p,24h}$ > 207 dB $L_{p,pk}$	186 dB $L_{E,p,24h}$	(N) High (I) High (F) Moderate	(N) High (I) High (F) Moderate
Sea turtles	> 210 dB $L_{E,p,24h}$ > 207 dB $L_{p,pk}$	(N) High (I) Low (F) Low	(N) High (I) Low (F) Low	(N) High (I) Moderate (F) Low	(N) High (I) Moderate (F) Low
Eggs and larvae	> 210 dB $L_{E,p,24h}$ > 207 dB $L_{p,pk}$	(N) Moderate (I) Low (F) Low	(N) Moderate (I) Low (F) Low	(N) Moderate (I) Low (F) Low	(N) Moderate (I) Low (F) Low

Table 2-8 Summary of the qualitative effects on fish from continuous noise from 5: Recommended guidelines for explosions according to Popper et al. (2014) for species of fish, sea turtles and eggs and larvae (N = near-field; I = intermediate-field; F = far-field).

Type of animal receptor	Mortality and potential mortal injury	Impairment			Behaviour
		Recoverable injury	TTS	Masking	
Fish: no swim bladder	(N) Low (I) Low (F) Low 229 – 234 dB $L_{p,pk}$	(N) Low (I) Low (F) Low	(N) Moderate (I) High (I) Moderate (F) Low	(N) High (I) High (F) Moderate N/A	(N) Moderate (I) High (I) Moderate (F) Low
Fish: swim bladder is not involved in hearing	(N) Low (I) Low (F) Low 229 – 234 dB $L_{p,pk}$	(N) Low (I) Low (F) Low	(N) Moderate (I) High (I) Moderate (F) Low	(N) High (I) High (F) Moderate N/A	(N) Moderate (I) High (I) Moderate (F) Low

Type of animal/receptor	Mortality and potential mortal injury	Impairment			Behaviour
		Recoverable injury	TTS	Masking	
Fish: swim bladder involved in hearing	(N) Low (I) Low (F) Low 229 – 234 dB $L_{p,pk}$	See Table 2-5 Criteria for recoverable injury and TTS in species of fish from continuous noise sources (Popper et al., 2014) (N) High (I) Low (F) Low	(N) High (I) Low (F) Low	(N) High (I) High (F) High N/A	(N) High (I) Moderate (F) Low
Sea turtles	(N) Low (I) Low (F) Low 229 – 234 dB $L_{p,pk}$	(N) Low High (I) Low (F) Low	(N) Moderate High (I) Low (F) Low	(N) High (I) High (F) Moderate N/A	(N) High (I) Moderate (F) Low
Eggs and larvae	(N) Low (I) Low (F) Low > 13 mms ⁻¹ peak velocity	(N) Low High (I) Low (F) Low	(N) Low High (I) Low (F) Low	(N) High (I) Moderate (F) Low N/A	(N) Moderate High (I) Moderate (F) Low

Table 2-9 ~~Summary of the qualitative effects on species of fish from explosions (6: Recommended guidelines for shipping and continuous noise according to Popper et al. (2014) for species of fish, sea turtles and eggs and larvae (N = near-field; I = intermediate-field; F = far-field).~~

Type of animal receptor	Mortality and potential mortal injury	Impairment				Behaviour
	Recoverable injury	Recoverable injury	TTS	TTS Masking	Masking	
Fish: no swim bladder	(N) High Low (I) Low (F) Low	(N) Low (I) Low (F) Low	(N) Moderate (I) Low (F) Low	(N) High (I) High (F) Moderate (F) Low	N/A	(N) High (I) Moderate (F) Low
Fish: swim bladder is not involved in hearing	(N) High Low (I) High Low (F) Low	(N) Low (I) Low (F) Low	(N) Moderate (I) Low (F) Low	(N) High (I) High (F) Moderate (F) Low	N/A	(N) High (I) Moderate (F) Low
Fish: swim bladder involved in hearing	(N) High Low (I) High Low (F) Low	170 dB $L_{p,48h}$	158 dB $L_{p,12h}$	(N) High (I) High (F) Lowhigh	N/A	(N) High (I) Moderate (F) Low
Sea turtles	(N) High Low (I) High Low (F) Low	(N) Low (I) Low (F) Low	(N) Moderate (I) Low (F) Low	(N) High (I) High (F) Low LowModerate	N/A	(N) High (I) Moderate (F) Low
Eggs and larvae	(N) High Low (I) Low (F) Low	(N) Low (I) Low (F) Low	(N) Low (I) Low (F) Low	(N) High (I) Low LowModerate (F) Low	N/A	(N) High (I) Moderate (F) Low

It is important to note that, despite the emerging evidence that fish are sensitive to particle motion (see section 2.2.2), the Popper *et al.* (2014) guidance defines noise impacts in terms of sound pressure or sound pressure-associated functions (i.e., $L_{E,p,t}$).

It has been suggested that the criteria set out by Popper *et al.* (2014) could have been derived from unmeasured particle motion, as well as sound pressure. Whilst this may be true, sound pressure remains the preferred metric in the criteria due to a lack of data surrounding particle motion (Popper and Hawkins, 2018), particularly regarding the ability to predict the consequences of the particle motion of a noise source, and the sensitivity of fish to a specific particle motion value. Therefore, as stated by Popper and Hawkins (2019): “since there is an immediate need for updated criteria and guidelines on potential effects of anthropogenic sound on fishes, we recommend, as do our colleagues in Sweden (Andersson *et al.*, 2017), that the criteria proposed by Popper *et al.* (2014) should be used.”

~~Both fleeing animal and stationary animal models have been used to cover the SEL_{cum} criteria for fish. It is recognised that there is limited evidence for fish fleeing from high level noise sources in the wild, and it would reasonably be expected that the reaction would differ between species. Most species are~~

~~likely to move away from a sound that is loud enough to cause harm (Dahl *et al.*, 2015; Popper *et al.*, 2014), some may seek protection in the sediment and others may dive deeper in the water column. For those species that flee, the speed chosen for this study of 1.5 m/s is relatively slow in relation to data from Hirata (1999) and thus is considered somewhat conservative.~~

~~Although it is feasible that some species will not flee, those that are likely to remain are thought more likely to be benthic species or species without a swim bladder; these are the least sensitive species. For example, from Popper *et al.* (2014): “There is evidence (e.g., Goertner *et al.*, 1994; Stephenson *et al.*, 2010; Halvorsen *et al.*, 2012) that little or no damage occurs to fish without a swim bladder except at very short ranges from an in-water explosive event. Goertner (1978) showed that the range from an explosive event over which damage may occur to a non-swim bladder fish is in the order of 100 times less than that for swim bladder fish.”~~

~~Stationary animal modelling has been included in this study, acknowledging the limited evidence for fish fleeing behaviour as a result of noise exposure, and other modelling for similar Environmental Impact Assessment (EIA) projects. However, basing the modelling on a stationary (zero flee speed) receptor is likely to greatly overestimate the potential risk to fish species, assuming that an individual would remain in the high noise level region of the water column for the whole duration of piling, especially when considering the precautionary nature of the parameters already built into the cumulative exposure calculations.~~

2.3.3 ~~2.2.3~~ Marine invertebrates

A review by Solé *et al.* (2023) highlights ~~that although the data on sound perception in invertebrates is scarce (see also ABPmer, 2023), there is~~ the increasing evidence that some types of anthropogenic noise can negatively impact a variety of marine invertebrate taxa. These impacts include changes in behaviour, physiology, and rate of mortality, as well as physical impairment, at the individual, population, or ecosystem level. Much of the damage from exposure to noise comes from vibration of the invertebrate body (André *et al.*, 2016) caused by the ~~passing~~ passage of sound.

Comparatively, the studies described by Solé *et al.* (2023) show a general inconsistency in the way noise impacts have been quantified for marine invertebrates. For example, Hubert *et al.* (2021) notes behavioural changes in blue mussels to 150 and 300 Hz tones, whereas Spiga *et al.* (2016) describes behavioural changes in the same species at ~~SEL_{ss} 153.5~~ $L_{E,p}$ (single pulse) 153.47 dB re 1 ~~$\mu\text{Pa}^2\text{s}$~~ μPa . These inconsistencies make it difficult to generate accurate thresholds for the onset of any impact for species. ~~Gastropod species⁴ showed effects ranging from changes in movement to cell damage and larva mortality, although numerical noise data at which these could occur is frequently not available.~~ A notable exception is the cephalopods group, in which, ~~in~~ several studies, mainly by Solé *et al.* (2019, 2013, 2018, 2013, 2019) and André *et al.* (2011) show a consistent threshold for auditory damage on various species of cephalopod at 157 dB re 1 ~~$\mu\text{Pa}^2\text{s}$~~ (SEL_{ss}) μPa . While further research is needed even on this group to ensure accurate thresholds which are satisfactory to regulators, the current state of research on cephalopods sets a goal for the research required for other marine invertebrate groups, if they are to be used usefully as impact thresholds.

The meta-analysis conducted by Solé *et al.* (2023) also reveals inconsistencies in the responses of taxonomically near species of marine invertebrates to the effect of anthropogenic noise. For example, Fields *et al.* (2019) demonstrates low mortality of zooplankton during seismic airguns, whereas for the same noise source, McCauley *et al.* (2017) showed mass mortality of krill larvae. Clearly, the effect of noise on one species may not necessarily be applicable on another species despite being taxonomically near, which again makes it difficult to

⁴ ~~It is understood the gastropod species whelk (family *Buccinum*) is of particular concern in the region of Arklow Bank. No specific data is available for this species.~~

generate a generalised impact threshold that can confidently be applied to different taxonomic groups of marine invertebrates.

In its current state, research on the effects of anthropogenic noise on marine invertebrates is emerging, but more slowly than for marine mammals and fish. At this time, this research is in too early a stage to be used to accurately generate impact thresholds which would be satisfactory to regulators. ~~However, it cannot be ignored that convincing evidence of noise impacts to marine invertebrates does exist.~~ The data available could potentially be referenced for some species but with caution, as there are still considerable gaps in the knowledge that would enable reliable conclusions for the impact of noise for most species.

2.4 EU regulations in marine mammals

ACP requested, in their Further Information Request (RFI 9a ii and 9I), consideration of underwater noise thresholds in use in other EU jurisdictions. As a result, this section details the current status of these thresholds, alongside the United Kingdom (UK).

2.4.1 ~~2.2.4 Particle motion~~ Ireland

In Ireland, the National Parks and Wildlife Service (NPWS), as part of Department of the Arts, Heritage and the Gaeltacht (DAHG, now Department of Housing, Local Government and Heritage), published guidance to manage the risk that anthropogenic sound sources pose to marine mammals in Irish waters (DAHG, 2014). This guidance presents a staged process towards managing risk but **does not enforce any specific process nor state explicit dB limits** to restrict noise impacts. Although it states that TTS “may constitute an injury” as it could “have negative effects on the ability to use natural sounds”, this is not a universally accepted definition (Tougaard *et al.*, 2025) and other countries do not consider these short-term effects an injury. However, it offers suggestions for mitigation that can be applied to projects, including pre-piling monitoring for marine mammals and a soft start and ramp up process. In section 4 of the guidance various management options for regulatory authorities to consider are detailed while approving any marine licences.

The recent publication of Ireland’s Marine Strategy Part 1, Assessment of the Marine Environment (Annex III), Department of Housing, Local Government and Heritage (DHLGH, 2025) also notes a threshold for Level of Onset of Biologically Significant Effect (LOBE) in reference to the Marine Strategy Framework Directive (MSFD) Descriptor 11 (Underwater Noise) of 176 dB SEL. This was the lowest reported sound level at which bottlenosed dolphins begin to experience temporary hearing loss when exposed to impulsive noise as presented in NMFS, (2018). Bottlenosed Dolphin was selected as a receptor organism for assessment of impulsive underwater noise since it is found throughout the Irish maritime area and is sensitive to mid-frequency sounds such as those generated by impulsive noise sources. It is important to note that the value of 176 dB SEL appears to have been selected erroneously: the results were derived from a table in NMFS (2018) that explicitly references continuous exposure to steady-state, non-impulsive, and therefore we do not recommend use of this threshold.

Additionally, the recent GOMOREUS study (Tougaard *et al.*, 2025), prepared for the Marine Institute, concludes with an overarching recommendation towards the Danish requirements (see section 2.4.2) or a modified single figure SEL threshold, similar to the German requirements (see section 2.4.3).

~~The criteria defined in the above section define the noise impacts on fishes in terms of sound pressure or sound pressure associated functions (i.e., SEL). It has been identified by researchers (e.g., Popper and Hawkins, 2019; Nedelec *et al.*, 2016; Radford *et al.*, 2012) that many species of fish, as well as invertebrates, actually detect particle motion rather than acoustic pressure. Particle motion describes the back-and-forth movement of a tiny theoretical ‘element’ of water, substrate or other media as a sound wave passes, rather than the pressure caused by the action of the force created by this movement. Particle motion is usually defined in reference to the velocity of the particle (often a peak particle velocity, PPV), but sometimes the related acceleration or displacement of the particle is used.~~

~~Note that species in the “Fish: swim bladder involved in hearing” category, the species most sensitive to noise, are sensitive to sound pressure.~~

~~Popper and Hawkins (2018) state that in derivation of the sound pressure based criteria in Popper *et al.* (2014) it may be the unmeasured particle motion detected by the fish, to which the fish were responding: there is a relationship between particle motion and sound pressure in a medium. This relationship is very difficult to define where the sound field is complex, such as close to the noise source or where there are multiple reflections of the sound wave in shallow water. Even these terms “shallow” and “close” do not have simple definitions.~~

~~The primary reason for the continuing use of sound pressure as the criteria, despite particle motion appearing to be the physical measure to which so many fish react or sense, is a lack of data (Popper and Hawkins, 2018) both in respect of predictions of the particle motion level as a consequence of a noise source such as piling, and a lack of knowledge of the sensitivity of a fish, or a wider category of fish, to a particle motion value. There continue to be calls for additional research on the levels of and effects with respect to levels of particle motion. Until sufficient data are available to enable revised thresholds based on the particle motion metric, Popper and Hawkins, 2019 states that “since there is an immediate need for updated criteria and guidelines on potential effects of anthropogenic sound on fishes, we recommend, as do our colleagues in Sweden (Andersson *et al.*, 2017), that the criteria proposed by Popper *et al.* (2014) should be used.”~~

~~2.4.2~~ ~~2.2.5 Seabed vibration~~Denmark

~~Vibration is commonly mentioned in association with noise and would typically be thought of as the movement of a solid surface (or substrate), which either radiates sound into an adjacent fluid medium (e.g. the vibration of the surface of a foundation pile transmitting sound into water) and propagates further, or affects a receptor directly connected to it.~~

~~High intensity sources that directly affect the seabed, such as impact piling, will also generate vibration and be transmitted through the substrate. This has the potential to affect benthic and demersal species (Roberts and Elliott, 2017; Popper and Hawkins, 2019). While the presence of vibration during piling is expected, an assessment of this has the same limitations as those for particle motion in fish, as little is known of the quantitative influence of the vibration source, or of the sensitivity of relevant species to it.~~

The Danish Energy Agency (2023) has published guidelines for managing underwater noise during the installation of impact or vibratory driven piles. They include a required prognosis scenario section which asks the developer to consider noise reduction methodologies during a planned construction case. The developer may choose to use NAS or not but must show that the planned activity complies with the relevant PTS thresholds highlighted in section 2 of their guidance. These thresholds follow those presented in Southall *et al.* (2019). Thresholds are species-specific, covering the six species of marine mammal common to Danish waters (grey and harbour seal, minke whale, pilot whale, white beaked dolphin, harbour porpoise, all of which are also found in Irish waters), and were derived from Southall *et al.* (2019) and reviewed in Tougaard (2021). The guidelines specify that **a radial distance from the piling, r_{safe} , should be presented over which there should be no animals of these species, and therefore no risk of PTS within this distance (i.e., modelled PTS range < r_{safe}),** and which can be mitigated through the use of an ADD (unless r_{safe} is low, < 200 m). In effect, the Danish guidelines use the same marine mammal thresholds as in the UK (Southall *et al.*, 2019) but use them to set a requirement for a PTS range that must be fully mitigated.

It is worth noting that although the Further Information Request from ACP states in section 9 a ii that PTS ranges of 200 m should not be exceeded to be compliant with Danish guidance, this is not strictly the case. In effect, the guidance requires that there is sufficient attenuation to reduce the PTS range enough that this range is mitigatable with an ADD. Indicative maximum ranges suitable for ADD effectiveness are 1 km (minimum) for minke whale (McGarry *et al.*, 2017) and 7.5 km for harbour porpoise (Brandt *et al.*, 2013).

2.4.3 Germany

Germany's regulation authority, Bundesamt für Seeschifffahrt und Hydrographie (BSH), was the first to establish a legal dB limit of **160 dB re 1 $\mu\text{Pa}^2\text{s}$ ($L_{E,p,ss}$) or 190 dB re 1 μPa ($L_{p,pk}$)** which must be complied with at a distance of 750 m from the pile installation location (Koschinski and Lüdemann, 2013; Müller and Zerbs, 2013). The criteria are derived from research into noise impacts on harbour porpoise (Lucke *et al.*, 2009). Since 2013, licences for pile driving are only granted if the anticipated piling noise is below this legal limit (Merchant and Robinson, 2020). These limitations have been adopted in marine licences for some projects in other jurisdictions, such as Taiwan, where potential Offshore Wind Farm (OWF) installations are a concern for the critically endangered Taiwanese white dolphin (*Sousa chinensis taiwanensis*; Liu *et al.*, 2018).

Merchant (2019) reviewed the German approach and concluded that it was the most effective noise management model available at the time. However, as pile sizes and hammer energies are increasing, the German sound limit is becoming increasingly more difficult to achieve and is not always met even with the deployment of multiple noise abatement systems (NAS) (Tetra Tech RPS Energy Limited and Seiche, 2024). The German requirements are also criticised in the GOMOREUS (Tougaard *et al.* 2025) study for the Marine Institute as not accounting for cumulative effects or any consideration of the sensitivity to noise of any species. They would also be much harder to meet in Ireland due to the deeper water, larger piles and hammers currently proposed, and the fact that the most common NAS utilised in German waters, bubble curtains, will not be suitable for many locations in Irish waters due to high currents.

2.4.4 Belgium

Within the European MSFD, underwater noise is addressed as a priority for implementation where it considers that 'good environmental status (GES) is achieved when introduction of energy, including underwater noise, is at levels that do not adversely affect the marine environment' (European Commission, 2017). To implement this, Belgium adopted an 'interim' criterion for impulsive noise of a 'maximum zero to peak noise level (L_{z-p}) (equivalent to $L_{p,pk}$ or SPL_{peak}) of **185 dB re 1 μPa at 750 m** from the source' (Anonymous, 2012a, as cited in Rumes *et al.*, 2016). It is understood that an $L_{E,p,ss}$ (SEL_{ss}) threshold of **162 dB re 1 $\mu\text{Pa}^2\text{s}$ at 750 m** is also proposed alongside the L_{z-p} limit. However, noise mitigation is only implemented if this noise limit is exceeded (Rumes *et al.*, 2016). Belgium also has seasonal piling restrictions where there is no piling from 1st January to 30th April to

protect sensitive life-history periods for key species such as harbour porpoise and seals (Rumes *et al.*, 2016; Verfuss *et al.* 2016).

2.4.5 Netherlands

The Netherlands requires mitigation for piling and geophysical survey activities on a **case-by-case basis** (Anonymous, 2012b, as cited in Rumes *et al.*, 2016). An example presented in Rumes *et al.* (2016) showed that they set a noise restriction at Borssele OWF at 160 to 172 dB re 1 $\mu\text{Pa}^2\text{s}$ at 750 m from the source. This project set up continuous noise monitoring and had a seasonal piling restriction between 1st January and 31st May to protect sensitive life-history periods for key species such as harbour porpoise and seals (Rumes *et al.*, 2016; Verfuss *et al.*, 2016).

There is no standard set of thresholds or criteria in the Netherlands.

2.4.6 United Kingdom

No decibel limit currently exists in the UK, although underwater noise may be one factor considered within the UK marine licensing system (Convention on Biological Diversity, 2016). The Department of Environment, Food and Rural Affairs (Defra) commissioned a feasibility assessment and pilot programme titled “A noise limit for offshore wind pile driving,” aimed at evaluating the feasibility of introducing a decibel-based noise limit for offshore wind pile driving in English and Welsh waters (Tetra Tech RPS Energy Limited and Seiche, 2024). As part of the study, modelling was conducted to assess the effectiveness of NAS. The findings concluded that whilst the German limit (160 dB re 1 $\mu\text{Pa}^2\text{s}$ $L_{E,p,ss}$ at 750 m) is technically achievable, it is situation dependent and not expected to be universally attainable for future projects in the UK, Germany and probably Ireland due to the increasing scale of new turbine designs. Therefore, a variable decibel limit would need to be applied to English and Welsh waters to allow developments to continue. It is also emphasised that a combination of NAS solutions is likely essential, especially for large, deep-water piles.

In January 2025, the Joint Nature Conservation Committee (JNCC), Natural England (NE) and the Centre for the Environment, Fisheries and Aquaculture Science (Cefas) published a joint position statement on the use of quieter piling methods and noise abatement systems when installing offshore wind turbine foundations (JNCC, 2025). This joint position statement states that the use of NAS or quieter installation methods is necessary to meet Test 2 of the European Protected Species (EPS) Licence application process, which requires applicants to demonstrate that there are no satisfactory alternatives.

Furthermore, the Marine Management Organisation (MMO) has published its policy paper “Reducing Marine Noise”, which outlines that developers are now expected to demonstrate best endeavours to reduce underwater noise. This includes the use of primary and/or secondary noise reduction measures, such as NAS (MMO, 2025).

2.4.7 Conclusion

Bearing in mind all of the details above, and in the absence of the revised underwater noise guidance for Ireland, it is considered that the Danish guidelines would be most appropriate and relevant to consider for the purpose of this assessment and in response to the Further Information Request from ACP, which requested consideration of defined thresholds in other European jurisdictions. While the German, Belgian and Dutch regulations are relatively straightforward to interpret, in that they use single-figure criteria to define a simple acceptable/unacceptable threshold, they are also based on older and limited studies focused on harbour porpoise (primarily Lucke *et al.*, 2009) that does not take into account the variety of species in Irish waters, their sensitivity to sound frequencies, cumulative exposures or the latest research on PTS, TTS and auditory injury in marine mammals. Such factors are, however, incorporated into the Danish guidance, which has the advantage of being easily updatable in the event that better data for PTS and TTS becomes available. The variable bathymetry around Ireland is also similar to the conditions around Denmark. As a result, the concepts and principles in this guidance are recommended for further consideration as part of this assessment and in response

to the RFI. This is also discussed in the recommendations in the Irish GOMOREUS report (Tougaard *et al.*, 2025), which discusses the applicability and consequences of the various noise thresholds in use in the different jurisdictions. The GOMOREUS study primarily investigates the balance between the use of German (or German-logic) thresholds and the Danish guidance, but criticises the lack of species-specific thresholds and weightings in the German framework.

Additionally, the TTS threshold for LOBE in Ireland's Marine Strategy Part 1 (2025) is not recommended and should be reviewed.

3 Modelling methodology

To estimate the underwater noise levels likely to arise during ~~the construction and operation of~~ impact piling operations at ABWP2, predictive noise modelling has been undertaken. The methods described in this section, and used within this report, meet the requirements set by the National Physical Laboratory (NPL) Good Practice Guide 133 for underwater noise measurement (Robinson *et al.*, 2014).

~~Of those considered, the noise source most important to consider is impact piling due to the noise level and duration it will be present (Bailey *et al.*, 2014). As such, the noise related to impact piling activities is the primary focus of this study.~~

The modelling of impact piling has been undertaken using the INSPIRE underwater noise model, ~~developed by Subacoustech Environmental~~ which has been widely used for wind farm assessments around the UK and Ireland. The INSPIRE model (currently version 5.26.0) is a semi-empirical underwater noise propagation model based around a combination of numerical modelling ~~(, a combined geometric and energy flow/hysteresis loss method)~~, and actual measured data. It is designed to calculate the propagation of noise in shallow (i.e., ~~less than generally around 100 m or less~~), mixed water; typical of the conditions around the UK and Ireland and ~~the UK, and well suited~~ well-suited for use in the Irish Sea. ~~The model has been tuned for accuracy using over 80 datasets of underwater noise propagation from monitoring around offshore piling activities.~~

The model INSPIRE provides estimates of unweighted ~~SPL_{peak}, SEL_{ss} and SEL_{cum}~~ $L_{p,pk}$, $L_{E,p,ss}$, $L_{p,RMS}$ and $L_{E,p,t}$ noise levels, as well as ~~various~~ other weighted noise metrics. Calculations are made along 180 ~~equally spaced radial~~ equally-spaced transects (one every two degrees). For each modelling run, a criterion level can be specified allowing a contour to be drawn, within which a given effect may occur. These results can then be plotted over digital bathymetry ~~data~~ so that impact ranges can be clearly visualised as necessary. INSPIRE also produces these contours as ~~Geographic Information System (GIS)~~ shapefiles.

INSPIRE considers a wide array of input parameters, including variations in bathymetry and source frequency to ensure accurate results are produced ~~specific to~~ for the location and nature of the piling operation. It should also be noted that the results should be considered conservative as maximum design parameters and ~~worst case~~ worst-case assumptions have been selected for:

- Piling hammer blow energies;
- Soft start, hammer energy ~~ramp up~~ ramp-up, and strike rate;
- Total duration of piling; and
- Receptor swim speeds.

This report presents results for remodelling of the underwater noise impacts from the previous (2024) version of Appendix 11.1 using the most up to date model and piling design parameters. More details are provided in this section.

Simpler modelling approaches have been used for noise sources other than impact piling that may be ~~present~~ presented during the construction and operation of ABWP2; these are discussed in section 5.

3.1 Modelling confidence

The INSPIRE model is semi-empirical, and as such, ~~a~~ validation process is inherently built into the development process. Whenever a new set of ~~good, reliable,~~ impact piling measurement data is gathered through offshore surveys, ~~either by Subacoustech or a published by a third party~~, it is compared against the outputted levels from INSPIRE and, if necessary, the model can be adjusted.

Currently ~~over 80, 120~~ separate impact piling noise datasets, primarily from ~~piling in the North and Irish Sea and North Sea Seas~~, have been used as part of the development for the latest version of INSPIRE, ~~and in each case~~. For $L_{E,p}$, an average ~~fit is used~~, or slightly above the average, fit to the data is used, meaning that for a given dataset some points in the measured dataset will be louder than the predicted level. When cumulative noise is considered this is necessary to reduce conservatism due to the variations in level for individual pile strikes, which can be as much as 5 to 10 dB (Bailey *et al.*, 2010). Calculating a cumulative SEL ($L_{E,p,t}$) based on every pulse being worst case would lead to an excessive prediction. For $L_{p,pk}$ however, a slightly more conservative fit to the data has been used to reduce the chance of underestimation. Designing a model to over-predict for all parameters would ultimately lead to an excessively precautionary and unrealistic model.

INSPIRE is designed to predict trends when increasing parameters beyond empirical data, and uses the measured data combined with standard acoustic theory to predict the effect of greater blow energies, larger piles and deeper water on the noise levels produced and propagated through the water.

~~In addition, INSPIRE is also~~ The largest pile diameter included in the analysis for development of INSPIRE v6.0 was 9.5 m in diameter, and the highest measured blow energy was 3,000 kJ. The model has been validated by comparing the noise levels outputted from the model with measurements and modelling undertaken by third parties, for example in Thompson *et al.* (2013) ~~and Thompson *et al.* (2025)~~. In Thompson *et al.* (2025), piles up to 10 m in diameter and blow energies up to 4,400 kJ were modelled using INSPIRE v5.2 in blind testing against measured data, and a good agreement was found in general, although INSPIRE was found to over-estimate the impact ranges in this earlier software version, especially closer to the pile (< 7 km) where the exposures to noise would have the greatest effect.

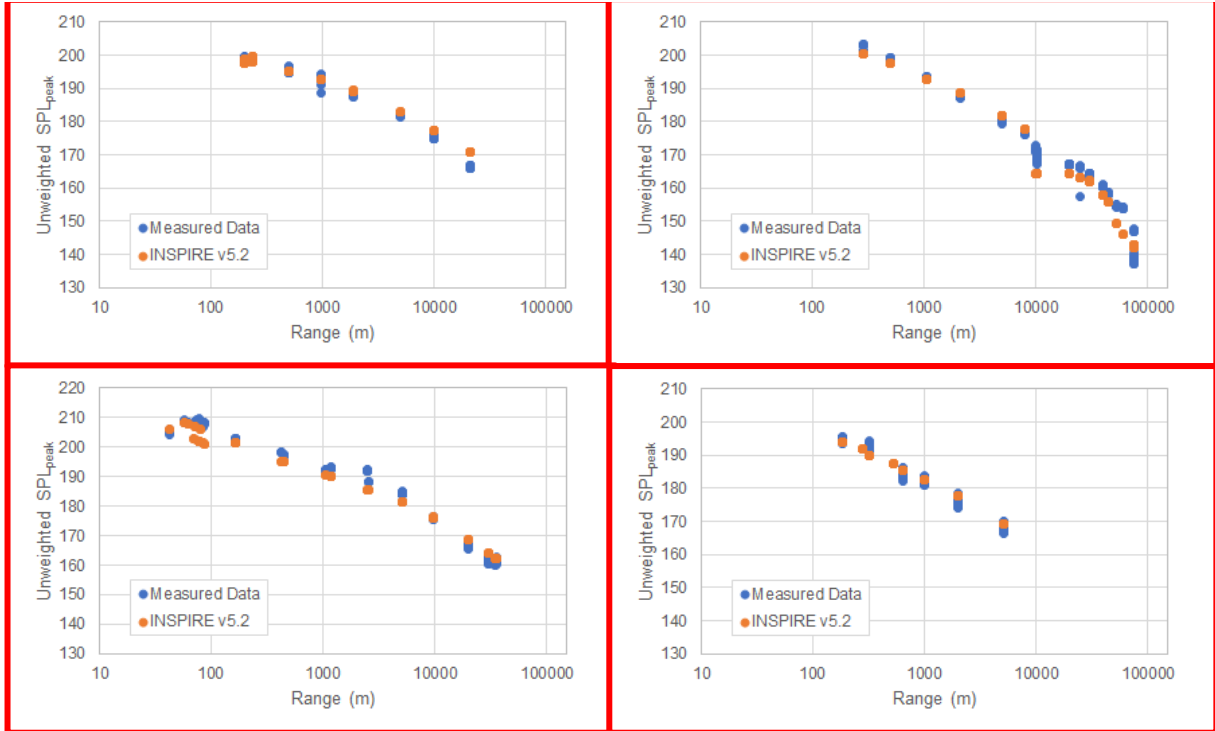
The ~~current~~ version of INSPIRE (~~version 5.2~~) used to remodel the impacts of underwater noise for ABWP2 (v6.0) in this report is the product of reanalysing all the impact piling noise in Subacoustech Environmental's measurement database ~~and any other data available~~, in preparation for the NMFS (2024) guidance, and cross-referencing it with blow energy data from piling logs, and relevant parameters such as pile diameter, length and water depth. This gives a database of single strike noise levels referenced to a specific blow energy at a specific range and environmental conditions, primarily water depth.

~~Previous iterations of the INSPIRE model have endeavoured to give a worst case estimate of underwater noise levels produced by various permutations of impact piling parameters. There is always some natural variability with underwater noise measurements, even when considering measurements of pile strikes under the same conditions (i.e., at the same blow energy, taken at the same range). For example, there can be variations in noise level of up to five or even 10 dB, as seen in Bailey *et al.* (2010) and the data shown in Figure 3-1. When modelling using the upper bounds of this range, in combination with other worst case parameter selections, conservatism can be compounded to create excessively overcautious predictions, especially when calculating SEL_{cum} . With this in mind, the current version of INSPIRE attempts to calculate closer to the average fit of the measured noise levels at all ranges.~~

Figure 3-1 ~~presents~~ and Figure 3-3 present a small selection of the measured impact piling noise data plotted against outputs from INSPIRE. The plots show data points from measured data (in blue) plotted alongside modelled data (in orange) ~~using from~~ INSPIRE ~~v5.2~~, matching the pile size, blow energy and position of the measured data. These show the fit to the data, with the modelled INSPIRE data points ~~sitting placed~~, more or less, in the middle (or slightly above, in the case of $L_{p,pk}$) of the measured noise levels at each range (this can also be seen in Figure 3-2 and Figure 3-4). When combined with the ~~worst case~~ precautionary assumptions in parameter selection, modelled results will remain precautionary.

The greatest deviations from the model tend to be at the greatest distances (> 10 km), where INSPIRE appears over-precautionary in many cases, but due to the lower relative levels the influence on the SEL_{cum} overall $L_{E,p,t}$ exposure will be minimal small.

Statistical analysis has been carried out to compare measured and modelled data to show the confidence present in INSPIRE v6.0. Figure 3-2 and Figure 3-4 show the distribution of the predicted levels against measured data with R^2 values of 0.81 for unweighted $L_{p,pk}$ and 0.89 for unweighted $L_{E,p,ss}$.



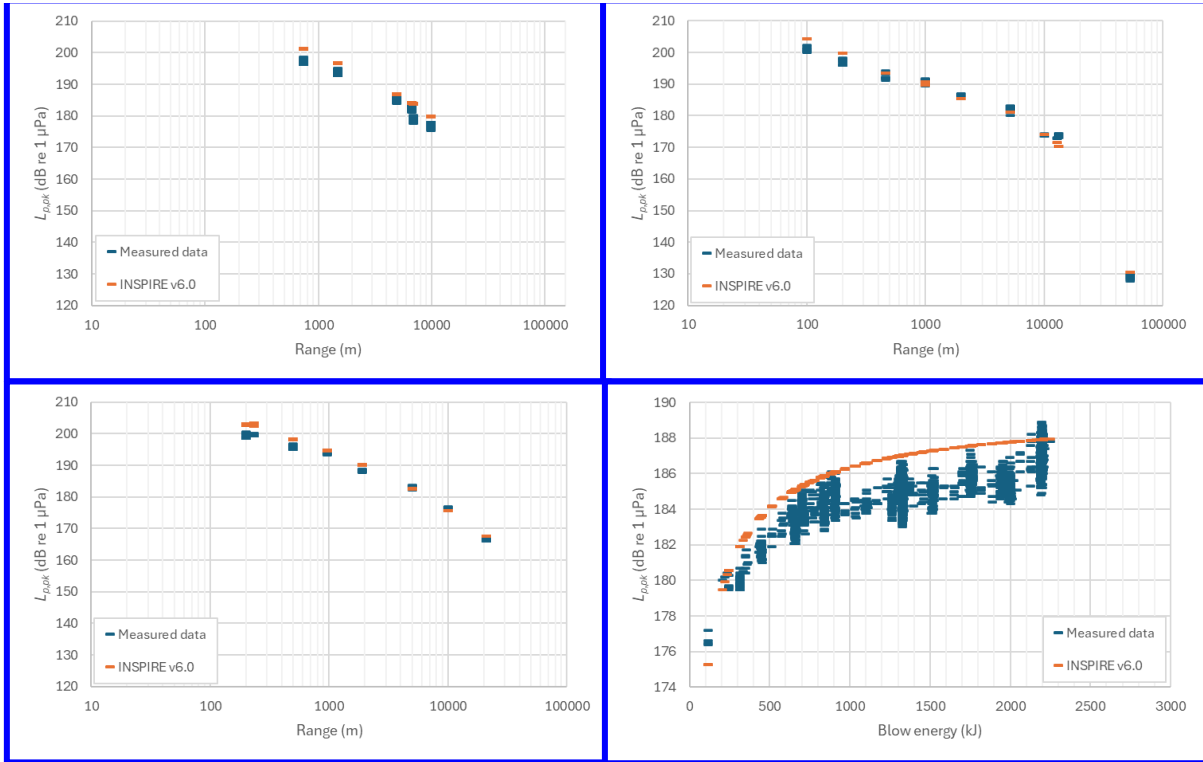


Figure 3-1: Comparison between example measured $L_{p,pk}$ impact piling data (blue points) and modelled data using INSPIRE version 5.2v6.0 (orange points)²¹.

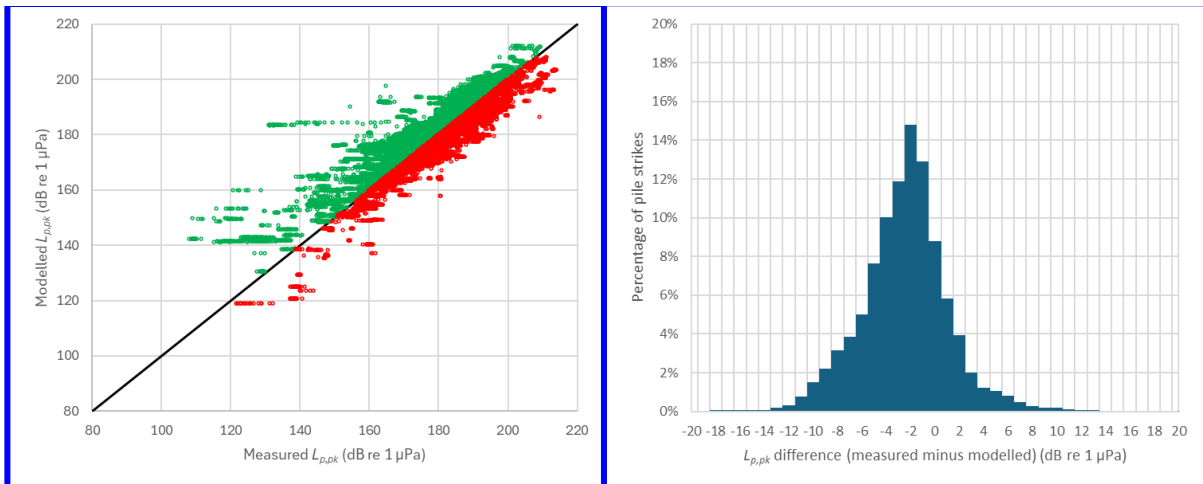


Figure 3-2: Distribution of measured impact piling data against modelled levels using INSPIRE v6.0 for unweighted $L_{p,pk}$ ($R^2 = 0.81$).

²¹ Top left: ~~6.08.6 m diameter pile, off the Suffolk coast~~ 2,500 kJ max hammer energy, North Sea, ~~2009~~2024; Top right: ~~4.86.0 m diameter pile, West of Barrow in Furness~~ 890 kJ max hammer energy, Irish Sea, 2010; Bottom left: ~~5.36.0 m diameter pile, off the North Welsh~~ 1,010 kJ max hammer energy, Suffolk Coast, ~~2012~~2009; Bottom right: ~~6.08.6 m diameter pile, off the coast of Cumbria, 2010~~ 3.0 km range, 2,250 kJ max hammer energy, North Sea, 2024.

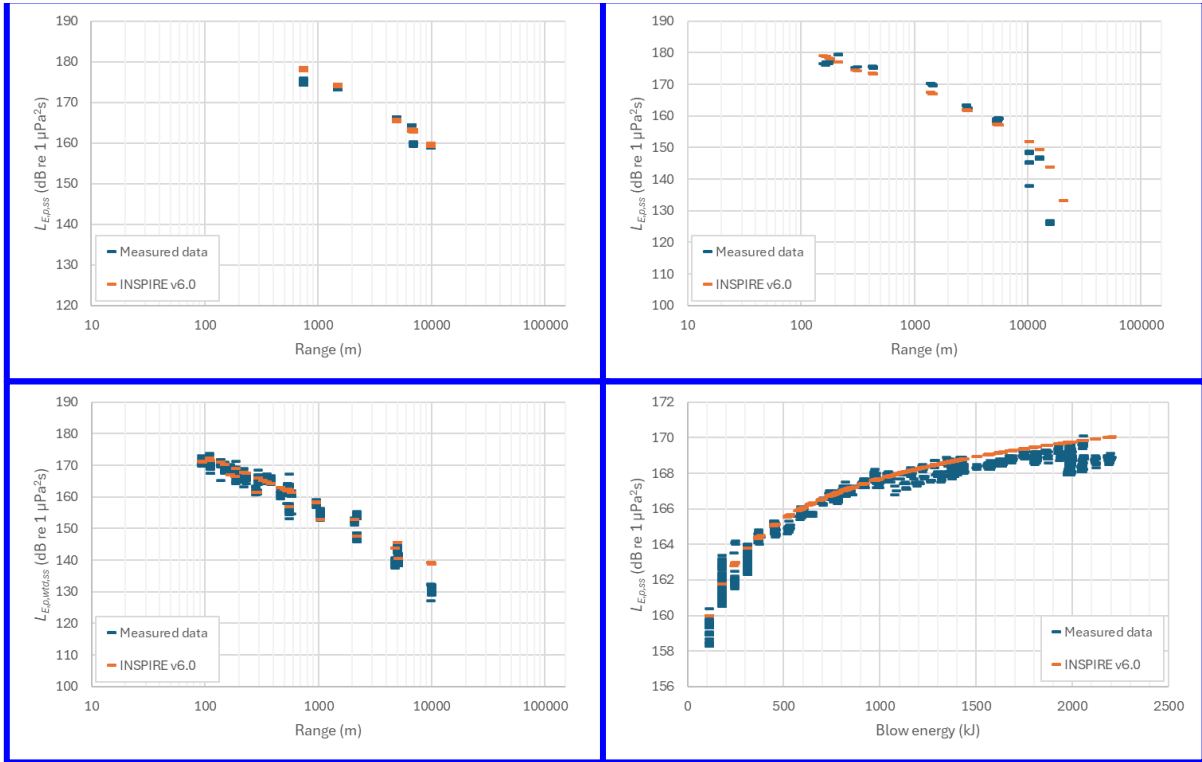
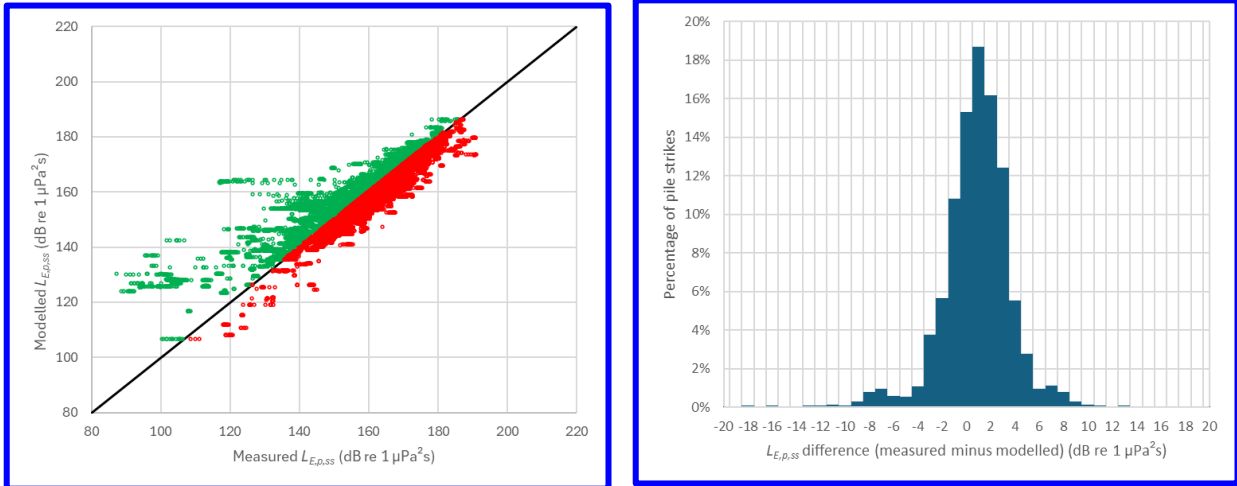


Figure 3-3: Comparison between example measured $L_{E,p,ss}$, impact piling data (blue) and modelled data using INSPIRE v6.0 (orange)².

Figure 3-4: Distribution of measured impact piling data against modelled levels using INSPIRE v6.0 for unweighted $L_{E,p,ss}$ ($R^2 = 0.89$).



² Top left: 8.6 m diameter pile, 2,500 kJ max hammer energy, North Sea, 2024; Top right: 4.7 m diameter pile, 1,340 kJ max hammer energy, North Wales Coast, 2012; Bottom left: 1.4 m diameter pile, 620 kJ max hammer energy, Lincolnshire Coast, 2011; Bottom right: 8.6 m diameter pile, 2.1 km range, 2,200 kJ max hammer energy, North Sea, 2024.

Additional validation has been undertaken using data presented by von Pein *et al.* (2022), which studied trends in noise level with changes in piling parameters using data primarily acquired in the North Sea and Baltic Sea. The data showed a strong correlation with blow energy, and a lower correlation with pile diameter, which Subacoustech agrees with, although the calculated correlation based on that data appears to have overestimated the trend. Figure 3-5 and Figure 3-6 are adapted from von Pein *et al.* (2022), replicating their results and overlaying with measured data from Subacoustech’s measurement database (selecting samples taken at the same reference distance) and results at equivalent datapoints using INSPIRE v6.0.

This shows a very good agreement with Subacoustech’s data (relating to blow energy). It should be noted that the upper and lower bounds for a correlation of noise level with pile diameter, based on the von Pein *et al.* (2022) data alone, could easily be close to horizontal; there is also no control for blow energy within the dataset, which is not constant. With the inclusion of Subacoustech’s data, there is little correlation at greater pile diameters, and it can be seen that the variations at a single pile diameter are largely controlled by changes in blow energy.

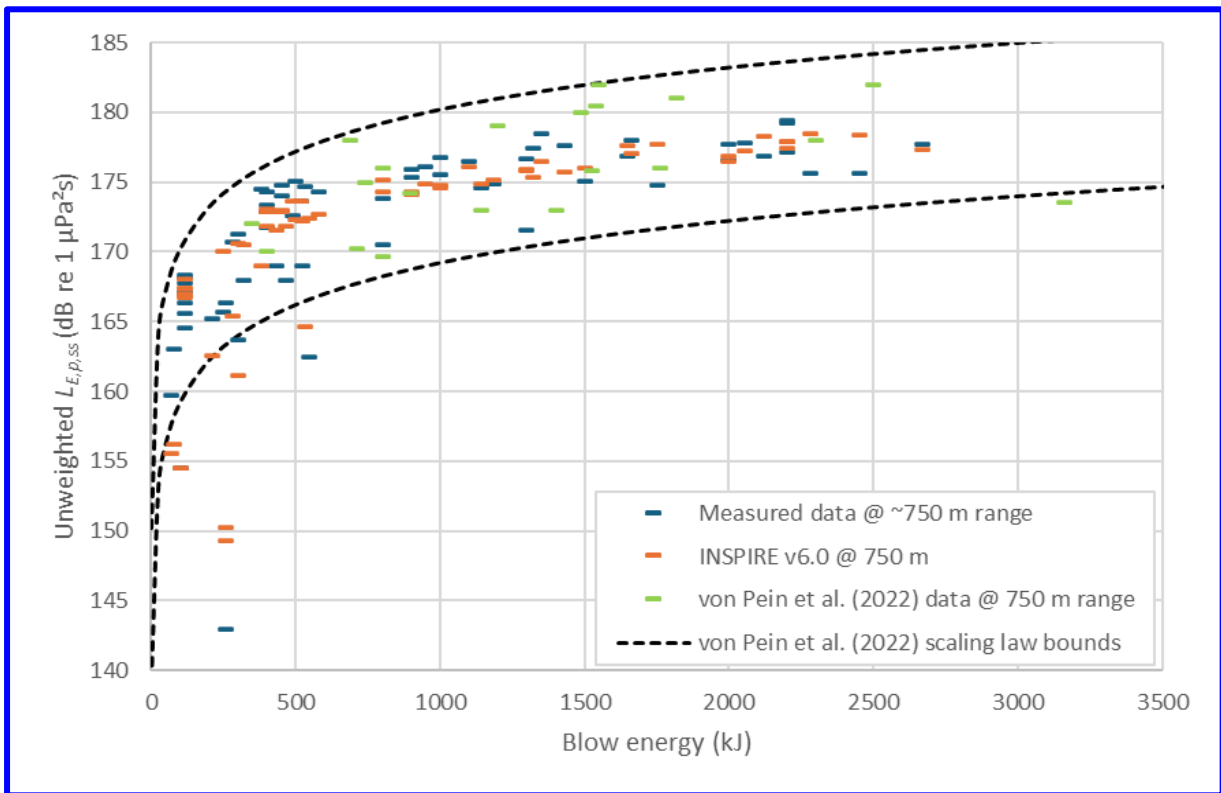


Figure 3-5: Data relating blow energy to noise level ($L_{E,p,ss}$) adapted from von Pein *et al.* (2022) (green) overlaid with Subacoustech measured data (blue) and INSPIRE v6.0 predictions (orange). Upper and lower scaling law bounds from von Pein (2022).

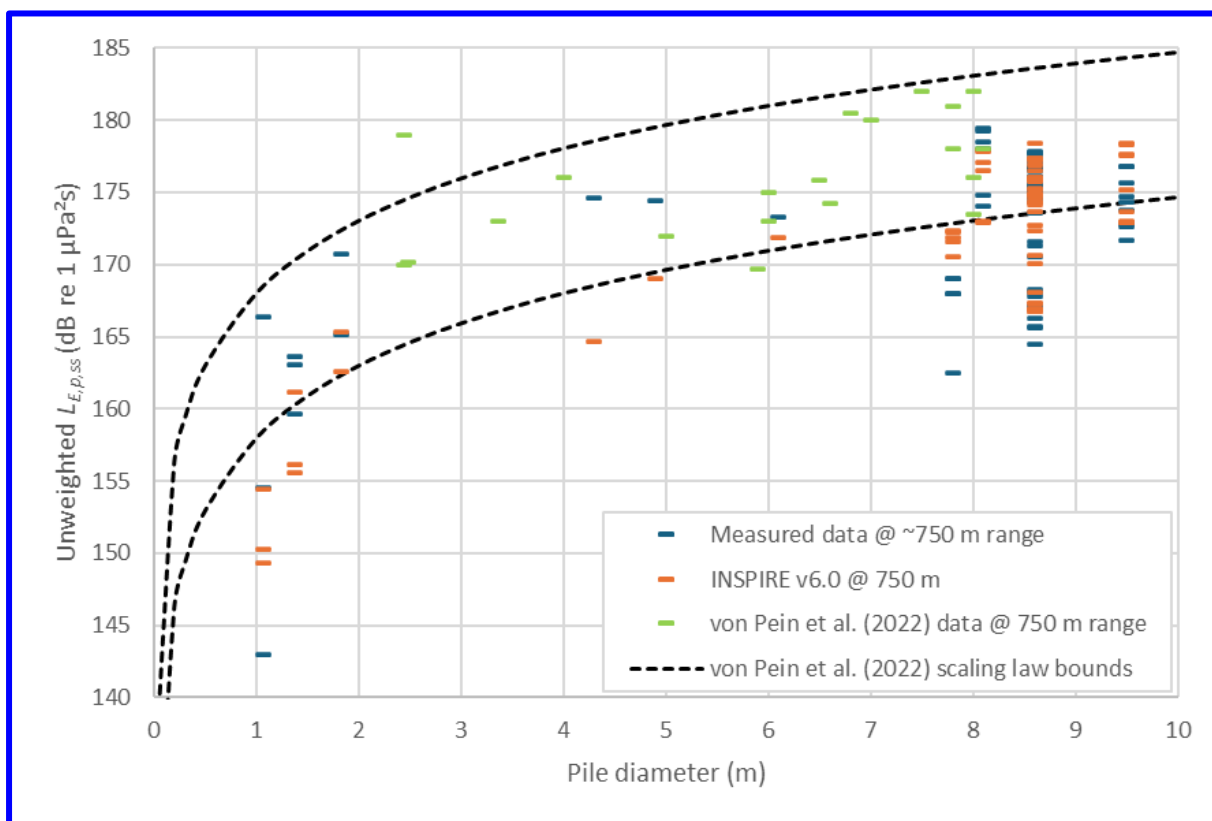


Figure 3-6: Data relating pile diameter to noise level ($L_{E,p,ss}$) adapted from von Pein (2022) (green) overlaid with Subacoustech measured data (blue) and INSPIRE v6.0 predictions (orange). Upper and lower scaling law bounds from von Pein (2022).

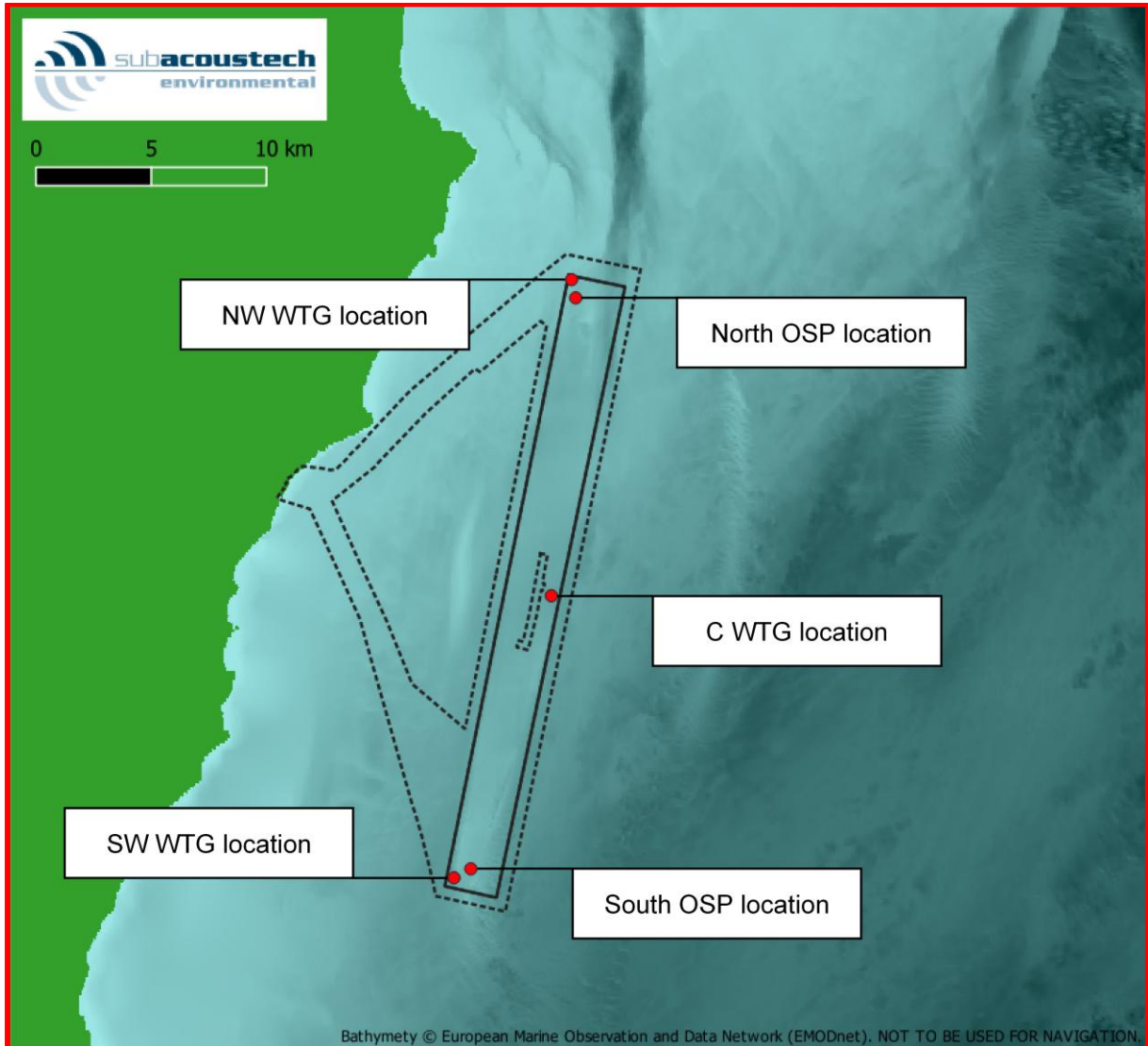
3.2 Modelling Input parameters

3.2.1 Modelling locations

Modelling for impact piling noise to install foundations for WTG and ~~Offshore Platform (OSP) foundation impact piling foundations~~ has been undertaken at ~~a total of~~ five representative locations covering the extents of ~~the ABWP2 site. The locations were chosen to give the greatest geographical spread to maximise the potential impact ranges to the north and the south of the site,~~ giving a spread of water depths, distances to shore, and bathymetry. Monopile foundations have been considered at three locations for WTG foundations and two locations for OSP foundations. These locations are summarised in Table 3-1 and illustrated in Figure 3-2 7, and have been selected from the Design Option 1 layout (53 WTG option).

Table 3-1: Summary of the underwater noise modelling locations used for this ~~study~~ assessment.

Modelling locations	Latitude	Longitude	Water depth
North West (NW) WTG location (WT03)	52.91636-52.91484°N	005.93213 005.91770°W	48.726.2 m
Centre (C) WTG central location (WT28)	52.79306-52.80250°N	005.93687 005.93113°W	35.630.7 m
South west location (SW) WTG (WT53)	52.68177-52.68030°N	005.99187 005.99125°W	30.326.4 m
North OSP location	52.90933-52.90693°N	005.92903 005.92530°W	24.422.0 m
South OSP location	52.68553-52.68627°N	005.98133 005.98945°W	26.321.1 m



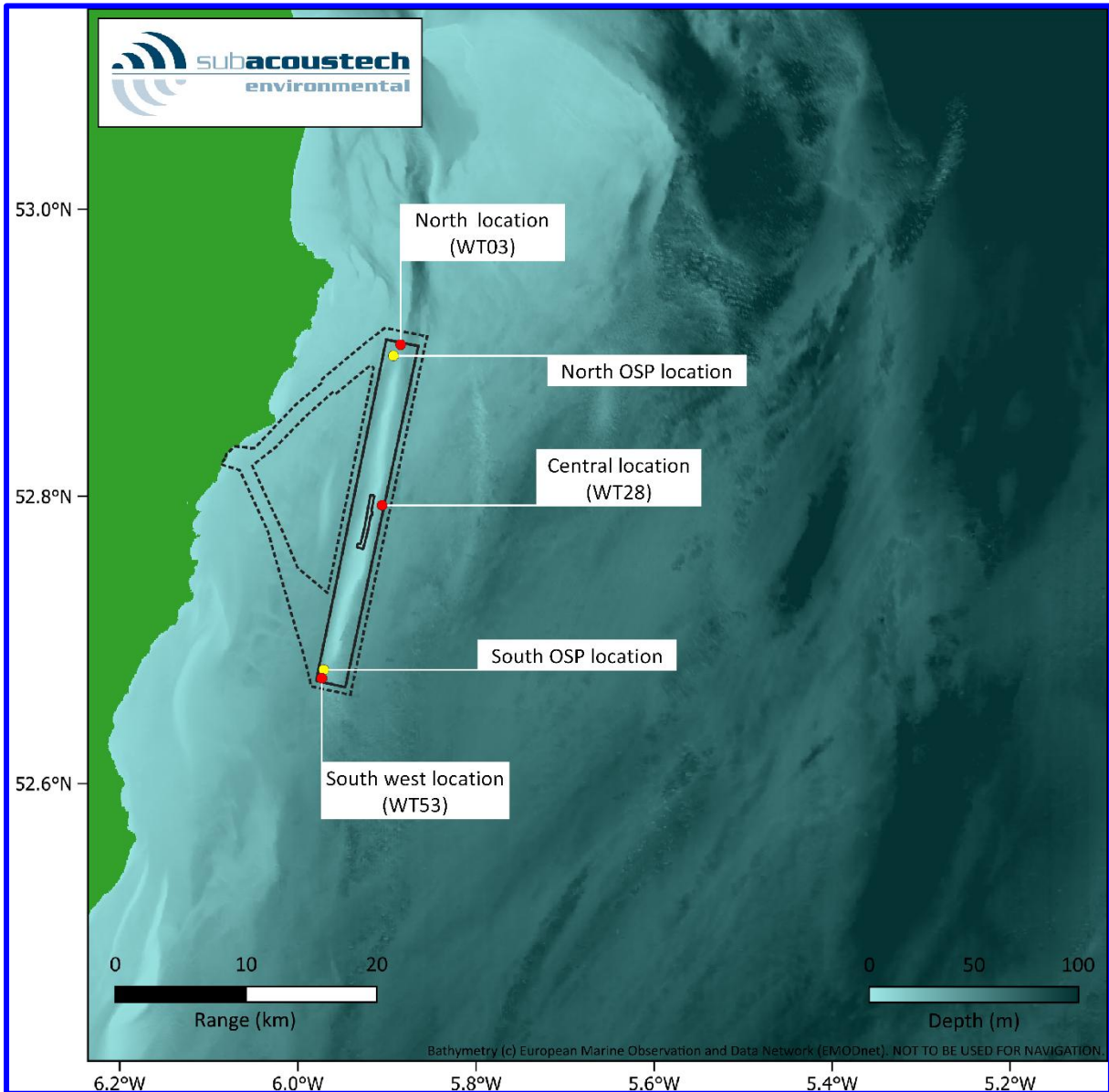


Figure 3-2-7: Approximate positions of the WTG and OSP modelling locations used at ABWP2.

3.2.2 Environmental conditions

With the inclusion of measured noise propagation data for similar offshore piling operations in UK and Irish waters, the INSPIRE model intrinsically accounts for various environmental conditions. This includes the differences that can occur with the temperature and salinity of the water throughout the day or year, as well as the sediment type in and around the site. Data from the British Geological Survey (BGS) show that the seabed in and around ABWP2 is made up of various combinations of sand and gravel over a bedrock of mudstone and sandstone.

Digital bathymetry from the European Marine Observation and Data Network (EMODnet, 2024) has been used for this modelling.

3.2.3 ~~3.2.2 WTG and OSP foundation and~~ *Impact piling parameters*

~~For each location, two foundation~~ Four piling scenarios have been considered ~~to cover the foundation options~~ for this study, including large (11 m diameter for WTG and 14 m diameter for OSP) monopiles and smaller (7 m diameter) monopiles. Maximum hammer energies have been identified in engineering estimates relevant to each piling location. ~~These are proposed for different locations within the site. In summary these are based on two monopile designs.~~ The largest diameter piles were chosen, smaller monopiles would have lower noise impacts.

- 11 m ~~and 7 m monopile~~ diameter monopiles for WTG foundations ~~at the NW and C WTG locations,~~ installed with a maximum blow energy of ~~4,000~~ 3,500 kJ;
- ~~14 m and 7 m monopile~~ diameter monopiles for OSP foundations ~~at the SW WTG location,~~ installed with a maximum blow energy of ~~6,600~~ 3,500 kJ;

In each case, two ramp-up scenarios have been modelled for these foundations, a Precautionary scenario using the largest piling energies and parameters likely to lead to the maximum impact ranges, and an Alternative scenario utilising lower energies for longer periods and slower strike rates. This Alternative scenario is designed to reduce $L_{E,p,t}$ impacts without the need for physical mitigation (section 3.2.4), to represent a more gradual, realistic ramp-up that is more likely to occur on site.

- ~~14 m and 7 m monopile foundations at the North OSP location, installed with a maximum blow energy of 4,000 kJ; and~~
- ~~14 m and 7 m monopile foundations at the South OSP location, installed with a maximum blow energy of 6,600 kJ.~~

For ~~SEL_{cum}~~ $L_{E,p,t}$ criteria, the soft start and ~~ramp-up~~ ramp-up of blow energies, along with the total duration of piling and strike rate, must ~~also~~ be considered. The ~~soft start and ramp-up have been designed to reduce the cumulative effect on marine fauna, especially by minimising the blow energy and strike rate at the start of the piling event.~~ The scenarios used for modelling assumptions used for this study are summarised in Table 3-2 to Table 3-5: for the WTG and OSP foundation scenarios. Each piling scenario considers a single monopile being installed in a 24-hour period, and the only difference between the WTG and OSP scenarios is the pile diameter.

Table 3-2: Summary of the soft start and ~~ramp-up scenario~~ ramp-up used for ~~both the 11 m and 7 m~~ ~~Precautionary WTG foundation monopile~~ ~~foundation modelling at the NW and C WTG locations~~ scenario.

WTG foundation (11 m diameter)	825 kJ		1,550 kJ	2,275 kJ	3,000 kJ	3,300 kJ	3,600 kJ	4,000 kJ
Number of strikes	612	600	400	400	400	450	450	3,300
Duration (s)	10 min	20 min	13 min 20 sec	13 min 20 sec	13 min 20 sec	15 min	15 min	110 min
Strike rate (bl/min)	0.6 bl/min	30 blows/minute	30		30			30
6,000 strikes over 3 hours 30 minutes duration per pile								

Table 3-3: Summary of the soft start and ~~ramp-up scenario~~ ramp-up used for the ~~11 m and 7 m~~ ~~Alternative WTG foundation monopile~~ ~~foundation modelling at the SW WTG location~~ scenario.

Alternative WTG foundation (11 m diameter)	825 kJ	800 kJ	1,550 kJ	2,275 kJ	3,000 kJ	4,000 kJ	4,450 kJ	6,600 kJ	
Number of strikes	18	6450	600	400600	400600	400	2,750	450	4,0003,600
Duration (s)	1,800	10 min 1,800	20 min 1,200	13 min 20 sec 1,200	13 min 20 sec 1,200	13 min 20 sec	94 min 40 sec	15 min	133 min 20 sec 7,200
Strike rate (bl/min)	0.6	0.6 bl/min 15	30 blows/minute	30		30			30
9,006,360 strikes, 5 over 4 hours 10 minutes duration per pile									

Table 3-4: Summary of the soft start and ~~ramp-up scenario~~ *ramp-up* used for ~~both the 14 m and 7 m~~ *both the 14 m and 7 m* ~~Precautionary OSP foundation monopile foundation modelling at the North OSP location~~ *Precautionary OSP foundation monopile foundation modelling at the North OSP location scenario.*

OSP foundation (14 m diameter)	825 kJ	1,550 kJ	2,275 kJ	3,000 kJ	3,300 kJ	3,600 kJ	4,000 kJ	
Number of strikes	612	600	400	400	450	450	3,3004,600	
Duration (s)	10 min 1,200	20 min 1,200	13 min 20 sec 800	13 min 20 sec 800	13 min 20 sec	15 min	15 min	110 min 9,200
Strike rate (bl/min)	0.6 bl/min	30 blows/minute	30		30		30	
6,006,012 strikes, over 3 hours 30 minutes duration per pile								

Table 3-5: Summary of the soft start and ~~ramp-up scenario~~ *ramp-up* used for ~~both the 14 m and 7 m~~ *both the 14 m and 7 m* ~~Alternative OSP foundation monopile foundation modelling at the South OSP location~~ *Alternative OSP foundation monopile foundation modelling at the South OSP location scenario.*

Alternative OSP foundation (14 m diameter)	825 kJ	800 kJ	1,550 kJ	2,275 kJ	3,000 kJ	4,000 kJ	4,450 kJ	6,600 kJ	
Number of strikes	18	6450	600	400600	400600	400	2,750	450	4,0003,600
Duration (s)	1,800	10 min 1,800	20 min 1,200	13 min 20 sec 1,200	13 min 20 sec 1,200	13 min 20 sec	94 min 40 sec	15 min	133 min 20 sec 7,200
Strike rate (bl/min)	0.6	0.6 bl/min 15	30 blows/minute	30		30			30
9,006,360 strikes, 5 over 4 hours 10 minutes duration per pile									

3.2.4 Mitigation

The effect of noise abatement systems (NAS) as mitigation for impact piling noise has also been investigated at ABWP2. For both the Precautionary WTG and OSP monopile foundation scenarios detailed in Table 3-2 and Table 3-4, an at-source reduction of 4 dB has been included alongside the unabated scenarios, representing the noise reduction from using a low-noise hammer.

No NAS has been considered for the Alternative ramp-up scenarios, as the ramp-up was designed to reduce impact ranges caused by the noise exposure.

3.2.5 3.2.3 Apparent source levels

Noise modelling requires knowledge of ~~the~~ source level, which is the theoretical noise level at one metre from the noise source. ~~The INSPIRE model assumes that the noise source — that is, the hammer striking the pile — acts as an effective single point, as it will appear at long distance. The source level is estimated based on the pile diameter and the blow energy imparted on the pile by the hammer. This is adjusted depending on the water depth at the modelling location to allow for the length of pile (and effective surface area) in contact with the water, which can affect the amount of noise that is transmitted from the pile into its surroundings.~~

It is ~~it is~~ worth noting that the ‘source level’ technically does not exist in the context of many shallow water (~~around~~ 100 m or less) noise sources (Heaney *et al.*, 2020). ~~The actual noise level one metre from the pile will be highly complex and vary up and down the water column by the pile, which is a long, extended noise source, rather than being one simple noise level.~~ In practice, for underwater noise modelling such as this, it is effectively an ‘apparent source level’ that is used, essentially a value that can be used to produce correct noise levels at range (for a specific ~~sound propagation~~ model), as required in impact assessments.

The INSPIRE model requires an apparent source level, which is estimated based on the pile diameter and the blow energy imparted on the pile by the hammer. This is adjusted depending on the water depth at the modelling location to allow for the length of the pile (and effective surface area) in contact with the water, which can affect the amount of noise that is transmitted from the pile into its surroundings. The unweighted, and unmitigated single strike $SPL_{peakL_p,pk}$ and $SEL_{ssL_{E,p,ss}}$ apparent source levels estimated for this study are provided in Table 3-6, ~~based on~~ and Table 3-7 for the maximum ~~blow~~ hammer energy and first pile strike hammer energy respectively, with no mitigation considered. Both WTG foundation scenarios have the same maximum hammer energy.

These figures are ~~in line with what has been previously requested~~ presented in accordance with requests commonly made by regulatory authorities ~~in England, Scotland and Wales and have been presented as it is not yet known what requirements are expected by Irish regulatory authorities.~~, although as indicated above, they are not necessarily compatible ~~or comparable~~ with any other model or predicted apparent source level. ~~In~~ Due to the similar water depths at each ~~case~~ modelling location, the differences in apparent source level ~~for each location~~ are minimal.

Table 3-6: Summary of the unweighted *and unmitigated apparent* source levels used for modelling, at maximum *blowhammer* energy.

Apparent source levels	Modelling location	$L_{p,pk}$ @ 1 m	$L_{E,p,ss}$ @ 1 m
WTG foundation Source levels (11 m diameter pile / 3,500 kJ maximum energy)	North location (WT03)	Large monopile foundation 11 m / 4,000 kJ (NW and C-WTG)	Smaller monopile foundation 7 m / 4,000 kJ (NW, C-WTG and North-OSP)
		11 m / 6,600 kJ (SW-WTG)	7 m / 6,600 kJ (SW-WTG and South-OSP)
OSP foundation (14 m diameter pile / 3,500 kJ maximum energy)	NW-WTG Central location (WT28)	242.4 246.8 dB re 1 μ Pa @ 1 m	242.2 219.0 dB re 1 μ Pa @ 1 m ² s
	South west location (WT53)	246.8 dB re 1 μ Pa	218.9 dB re 1 μ Pa ² s
	North OSP location	246.9 dB re 1 μ Pa	219.1 dB re 1 μ Pa ² s
	South OSP location	246.9 dB re 1 μ Pa	219.1 dB re 1 μ Pa ² s

Table 3-7: Summary of the unweighted and unmitigated apparent source levels used for modelling of the first pile strike.

Apparent source levels	C-WTG Modelling location	242.4 dB re 1 $\mu\text{Pa}_{L_{p,pk}}$ @ 1 m	242.4 dB re 1 $\mu\text{Pa}_{L_{E,p,ss}}$ @ 1 m
WTG foundation (11 m diameter pile / 825 kJ energy)	North location (WT03)	244.1 dB re 1 μPa	214.7 dB re 1 $\mu\text{Pa}^2\text{s}$
	Central location (WT28)	244.2 dB re 1 μPa	214.8 dB re 1 $\mu\text{Pa}^2\text{s}$
	SW-WTG South west location (WT53)	243.1 244.1 dB re 1 μPa @ 1 m	243.1 214.7 dB re 1 $\mu\text{Pa}^2\text{s}$ @ 1 m
Alternative WTG ramp-up (11 m diameter pile / 400 kJ energy)	North location (WT03)	241.4 dB re 1 μPa	212.3 dB re 1 $\mu\text{Pa}^2\text{s}$
	North-OSP Central location (WT28)	242.4 241.4 dB re 1 μPa @ 1 m	242.4 212.3 dB re 1 $\mu\text{Pa}^2\text{s}$ @ 1 m
	South west location (WT53)	241.4 dB re 1 μPa	212.3 dB re 1 $\mu\text{Pa}^2\text{s}$
OSP foundation (14 m diameter pile / 825 kJ energy)	South North OSP location	243.1 244.3 dB re 1 μPa @ 1 m	243.0 214.9 dB re 1 $\mu\text{Pa}^2\text{s}$ @ 1 m
	South OSP location	244.3 dB re 1 μPa	214.9 dB re 1 $\mu\text{Pa}^2\text{s}$
Alternative OSS ramp-up (14 m diameter pile / 400 kJ energy)	North OSP location	241.5 dB re 1 μPa	212.5 dB re 1 $\mu\text{Pa}^2\text{s}$
Unweighted SEL_{ss}	NW-WTG South OSP location	223.5 241.5 dB re 1 $\mu\text{Pa}^2\text{s}$ @ 1 m	223.1 212.5 dB re 1 $\mu\text{Pa}^2\text{s}$ @ 1 m

3.2.6 Predicted noise levels at 750 m from the noise source

In addition to the apparent source levels, it is useful to look at the potential noise levels at a range of 750 m from the noise source, which is a common feature of underwater noise studies where the primary consideration is impact piling.

These levels have the added advantage of being comparable with other modelling or measurements (as a valid measurement can be taken at this range; for example, von Pein *et al.*, 2022), where the source level (or apparent source level) may not. A summary of the modelled unweighted levels at a range of 750 m are given in Table 3-8 and Table 3-9, considering the transect with the greatest noise transmission at each location while piling at the maximum hammer energy and the hammer energy at first pile strike with no mitigation considered.

Table 3-8: Summary of the unweighted and unmitigated $L_{p,pk}$ and $L_{E,p,ss}$ (single strike) noise levels at a range of 750 m from the noise source when considering the maximum hammer energy.

Predicted levels	Modelling location	$L_{p,pk}$ @ 750 m	$L_{E,p,ss}$ @ 750 m
WTG foundation (11 m diameter pile / 3,500 kJ maximum energy)	C-WTG North location (WT03)	223.5 200.2 dB re 1 $\mu\text{Pa}^2\text{s}$ @ 1 m	223.4 178.8 dB re 1 $\mu\text{Pa}^2\text{s}$ @ 1 m
	Central location (WT28)	201.3 dB re 1 μPa	179.4 dB re 1 $\mu\text{Pa}^2\text{s}$
	South west location (WT53)	200.2 dB re 1 μPa	178.8 dB re 1 $\mu\text{Pa}^2\text{s}$
OSP foundation (14 m diameter pile / 3,500 kJ maximum energy)	North OSP location	199.2 dB re 1 μPa	178.4 dB re 1 $\mu\text{Pa}^2\text{s}$
	SW-WTG South OSP location	224.3 197.2 dB re 1 $\mu\text{Pa}^2\text{s}$ @ 1 m	224.3 177.0 dB re 1 $\mu\text{Pa}^2\text{s}$ @ 1 m

Table 3-9: Summary of the unweighted and unmitigated $L_{p,pk}$ and $L_{E,p,ss}$ (single strike) noise levels at a range of 750 m from the noise source when considering the hammer blow energy at the first pile strike.

Predicted levels	Modelling location	$L_{p,pk}$ @ 750 m	$L_{E,p,ss}$ @ 750 m
WTG foundation (11 m diameter pile / 825 kJ energy)	North OSP location (WT03)	223.5 197.5 dB re 1 $\mu\text{Pa}^2\text{s}$ @ 1 m	223.4 174.6 dB re 1 $\mu\text{Pa}^2\text{s}$ @ 1 m
	Central location (WT28)	198.6 dB re 1 μPa	175.2 dB re 1 $\mu\text{Pa}^2\text{s}$
	South west location (WT53)	197.6 dB re 1 μPa	174.6 dB re 1 $\mu\text{Pa}^2\text{s}$
Alternative WTG ramp-up (11 m diameter pile / 400 kJ energy)	South-OSP North location (WT03)	224.3 194.8 dB re 1 $\mu\text{Pa}^2\text{s}$ @ 1 m	224.3 172.2 dB re 1 $\mu\text{Pa}^2\text{s}$ @ 1 m
	Central location (WT28)	195.9 dB re 1 μPa	172.7 dB re 1 $\mu\text{Pa}^2\text{s}$

	South west location (WT53)	194.8 dB re 1 μ Pa	172.2 dB re 1 μ Pa ² s
OSP foundation (14 m diameter pile / 825 kJ energy)	North OSP location	196.6 dB re 1 μ Pa	174.2 dB re 1 μ Pa ² s
	South OSP location	194.5 dB re 1 μ Pa	172.8 dB re 1 μ Pa ² s
Alternative WTG ramp-up (11 m diameter pile / 400 kJ energy)	North OSP location	193.8 dB re 1 μ Pa	171.8 dB re 1 μ Pa ² s
	South OSP location	191.8 dB re 1 μ Pa	170.4 dB re 1 μ Pa ² s

3.2.7 Predicted noise levels against range

Figure 3-8 has been provided in order to show the modelled noise transmission, which can be used as a basis to compare and validate the levels against future noise monitoring. This plot presents the predicted unweighted $L_{p,pk}$ and $L_{E,p,ss}$ noise levels against range over the longest calculated transect (130°; SE) from the Central (WT28) modelling location during installation of a WTG foundation monopile using the maximum hammer energy (3,500 kJ). It should not be assumed necessarily comparable to any other transect or blow energy, although it is expected to present a precautionary scenario.

3.2.4 Environmental conditions

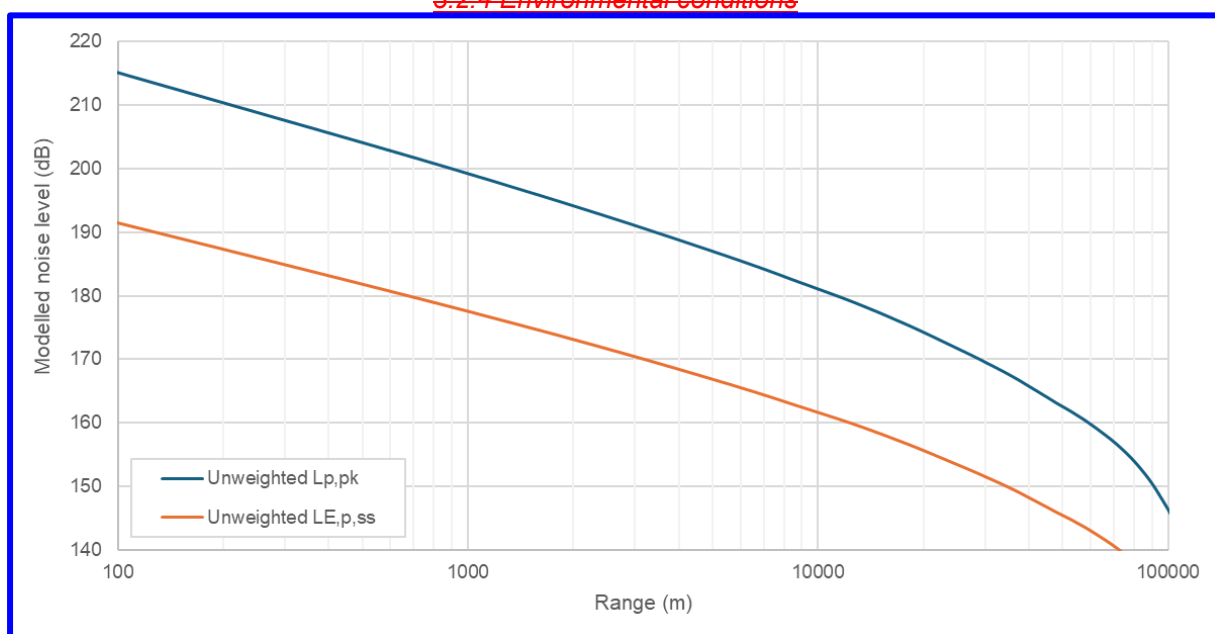


Figure 3-8: Modelled unweighted $L_{p,pk}$ and $L_{E,p,ss}$ noise levels with range for the monopile scenario assuming maximum blow energy along a SE transect (130°) from the Central (WT28) modelling location.

With the inclusion of measured noise propagation data for similar offshore piling operations in UK waters, the INSPIRE model intrinsically accounts for various environmental conditions. This includes the differences that can occur with the temperature and salinity of the water, as well as the sediment type surrounding the site. Data from the British Geological Survey (BGS) show that the seabed in and around ABWP2 is generally made up of variations of gravel, sand, and muddy sand, which are typical of the seabed conditions at which many of the measurements of piling included in the INSPIRE model have been sampled (McBreen *et al.*, 2011).

Digital bathymetry from the European Marine Observation and Data Network (EMODnet) has been used for this modelling. Mean tidal depth has been used throughout.

3.3 Cumulative SELs $L_{E,p,t}$ and fleeing receptors

Expanding on the information in section 2.3.2.3 regarding SEL_{cum} criteria $L_{E,p,t}$ and the fleeing animal model assumptions used for modelling, it is important to understand the meaning of the results presented in the following sections. This section lays out the methodology behind calculating these results to aid with interpretation.

When an SEL_{cum} $L_{E,p,t}$ impact range is presented for a fleeing animal, this range can essentially be considered a starting position (at the commencement of piling) for the fleeing animal receptor. For example, if a receptor began to flee in a straight line away from the noise source, starting at the position (distance from a pile) denoted by a modelled PTS contour, the receptor would receive exactly the noise exposure as per the PTS onset criterion under consideration (see Figure 3-11).

To help explain this, it is helpful to examine how the multiple pulse SEL_{cum} ranges are calculated. As explained in section 2.1.2, the SEL_{cum} is a measure of the total received noise over a whole operation: in the cases of the Southall et al. (2019) and Popper et al. (2014) criteria, this covers noise in a 24-hour period unless otherwise specified.

When considering a stationary receptor (i.e., one that stays at the same position throughout the piling operation, with no flee response), calculating the SEL_{cum} $L_{E,p,t}$ is straightforward: all the noise levels produced and received at a single point along a transect are aggregated to calculate the SEL_{cum} $L_{E,p,t}$. If this calculated level is greater than the threshold being modelled, the model steps considered, a location slightly further from the noise source is selected and the noise levels from that new location are aggregated to calculate a new SEL_{cum}. This continues outward until the threshold is met.

For a fleeing animal model, the receptor's varying distance from the noise source while moving away also needs to be relative to the receptor is considered. To model this, a nominal starting point close to the source is chosen and the received noise level for each noise event (e.g., pile strike) while the receptor is fleeing is noted; the receptor moves away from the source at a defined speed through the piling operation. For example, if a noise event (i.e., a pulse from a pile strike) occurs every six seconds, and an animal is fleeing at a rate of 1.5 m/s, it is 9 m further from the noise source after each noise pulse, resulting in a slightly reduced noise level each time. These values are then aggregated into an SEL_{cum} $L_{E,p,t}$ value over the entire operation. The faster an animal is fleeing, the greater the distance travelled between noise events. The impact range outputted by the model for this situation is the distance represents the location the receptor must be at the start of the operation to exactly meet the exposure threshold.

As an example, the graphs in Figure 3-3-9 and Figure 3-4-10 show the difference in the received SEL_{ss} $L_{E,p,ss}$ and $L_{E,p,t}$ from a stationary receptor and a fleeing receptor travelling at a constant speed of 1.5 m/s using the 11 m monopile transect with the largest transmission (130°) for the worst-case WTG foundation scenario at the SE-WTG installation at the Central (WT28) modelling location.

The received SEL_{ss} single strike $L_{E,p,ss}$ from the stationary receptor, as illustrated in Figure 3-3-9, shows the noise level gradually increasing levels rising as the blow energy increases throughout the piling operation monopile installation. These step changes are also visible for the fleeing receptor, but as the fleeing receptor is further from the noise source by the time the levels increase, the total received exposure reduces, resulting in progressively lower received noise cumulative levels. As an example, for the first 4020 minutes of the piling scenario, during the slow start, where the blow energy is 660825 kJ, a receptor fleeing at a rate of 1.5 m/s, the fleeing receptor will have moved 0.9 ms⁻¹ has the potential to move 1.8 km away from the noise source. After the full piling installation of more than five 3 hours, the and 40 minutes, a receptor will have the potential to be almost 2820 km from the pile noise source.

Figure 3-4-10 shows the effect that these different differing single strike received levels have when calculating the $SEL_{cum} = L_{E,p,t}$, clearly showing the difference in the cumulative effect of the levels between a receptor remaining still, as opposed to fleeing. To use an extreme example, starting at a range of 1 m, the first strike results in a received level of 218.3214.8 dB re 1 $\mu Pa^2 s$. If the receptor were to remain stationary throughout for the entire piling operation it would receive a cumulative level of 262.7 noise exposure of 256.3 dB re 1 $\mu Pa^2 s$, whereas when fleeing at a receptor flees at a constant speed of 1.5 m/s over the same scenario would result in a, the cumulative received level of exposure is calculated to be just 218.5214.9 dB re 1 $\mu Pa^2 s$ for the receptor, only slightly higher than the first pile strike alone.

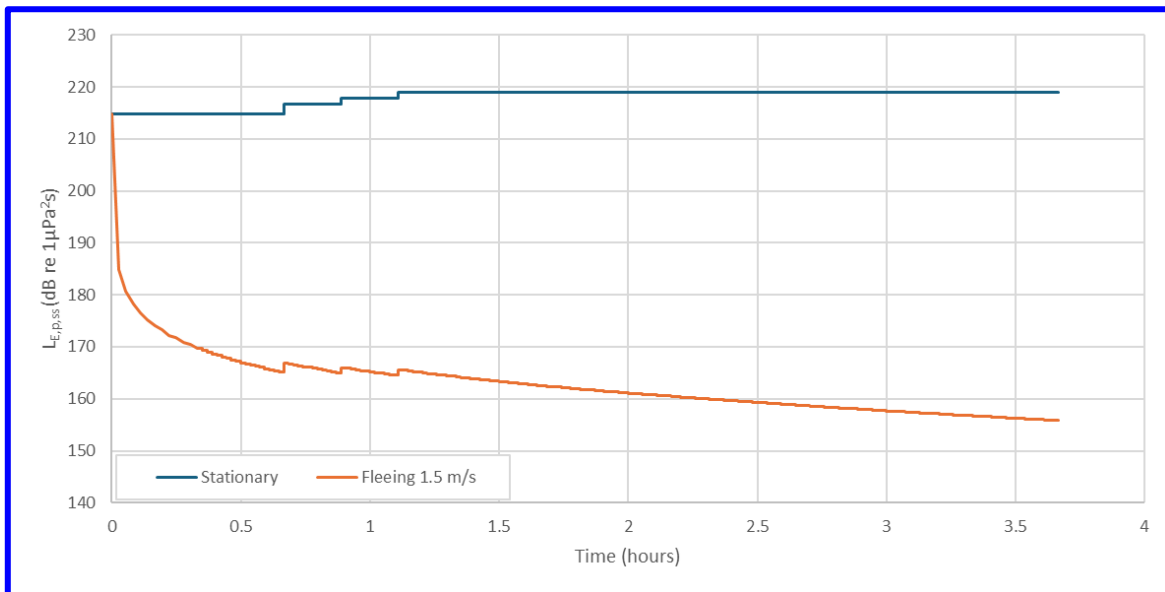
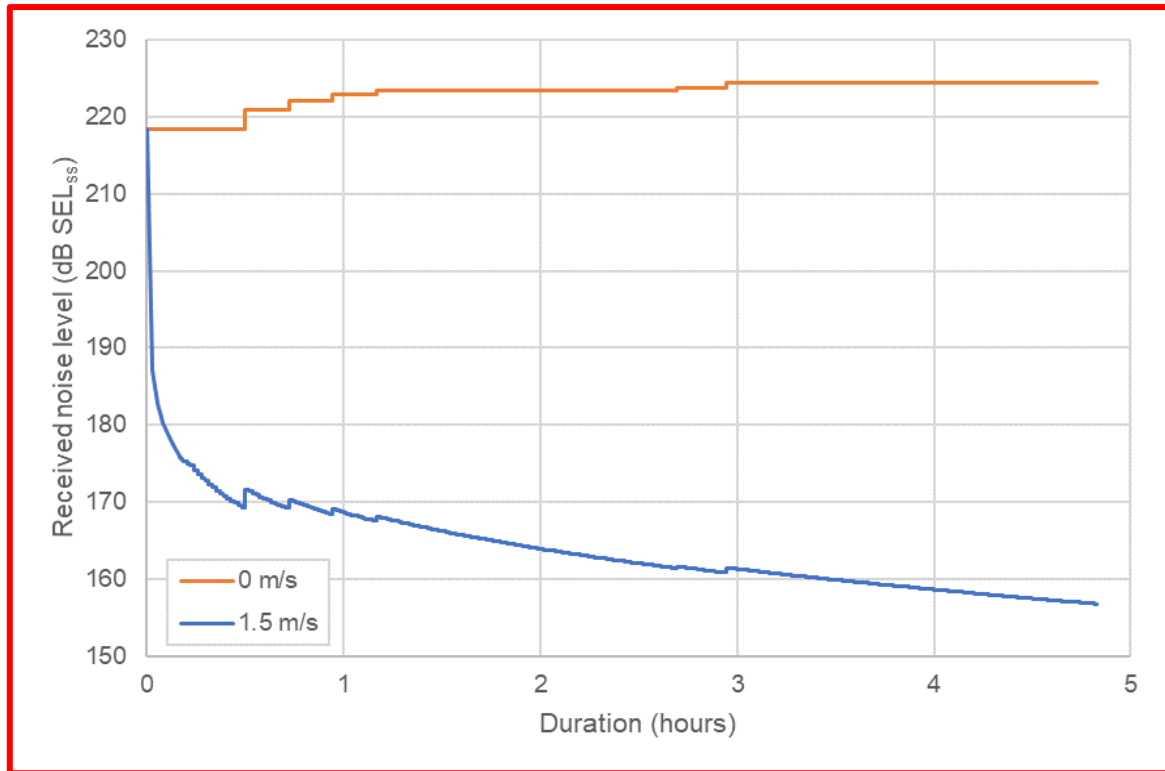


Figure 3-3-9: Received single strike noise levels ($SEL_{ss} = L_{E,p,ss}$) for receptors using the 11 m monopile during the worst-case WTG foundation installation parameters at the SE Central (WT28) modelling

location, assuming both a stationary and fleeing receptor/receptors starting at a location 1 m from the noise source.

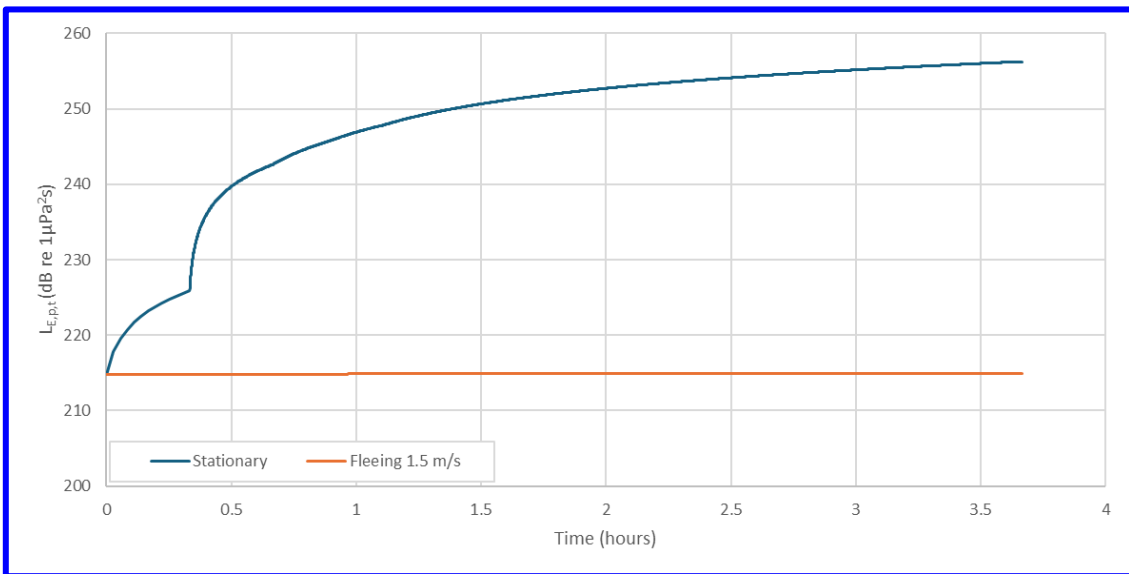
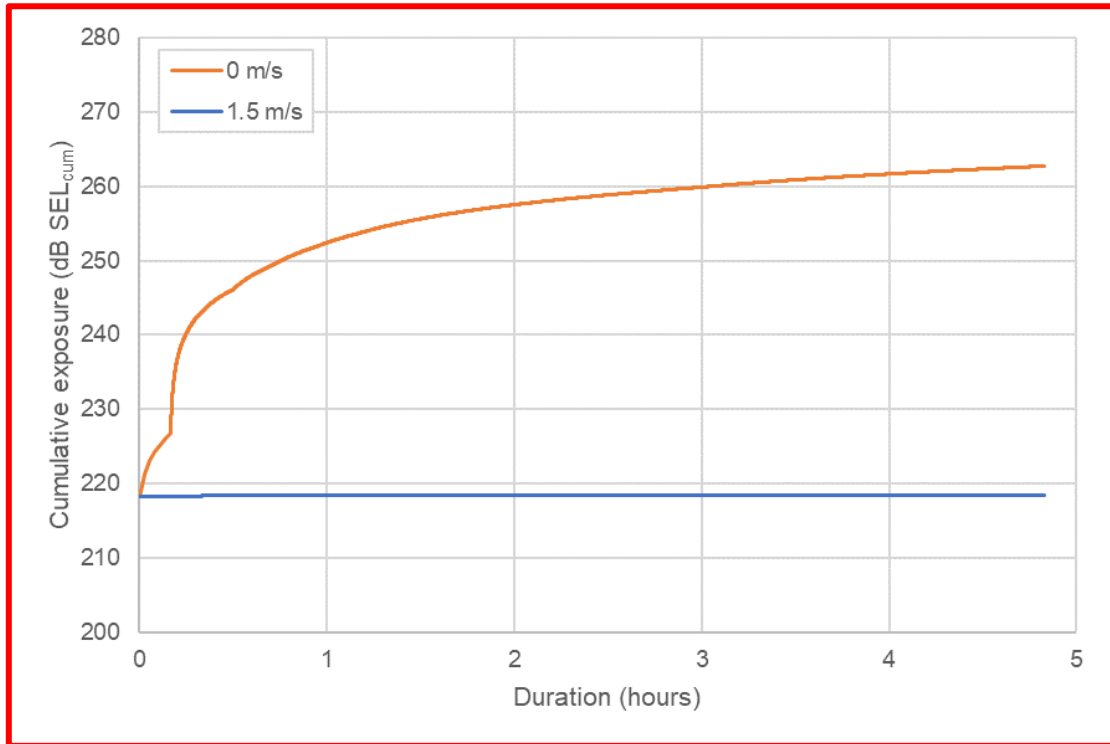


Figure 3-4-10: Cumulative received noise levels (SEL_{cum} level ($L_{E,p,t}$)) for receptors using the 11 m monopile during the worst-case WTG foundation installation parameters at the SE Central (WT28) modelling location, assuming both a stationary and fleeing receptor/receptors starting at a location 1 m from the noise source.

To summarise, if the receptor were to start fleeing in a straight line away from the noise source starting at a range closer than the modelled value impact range, it would receive a noise exposure in excess of the criteria criterion, and if the receptor were to start fleeing from a range further than the modelled value, it would receive a noise exposure below the criteria criterion. This is illustrated in Figure 3-11.

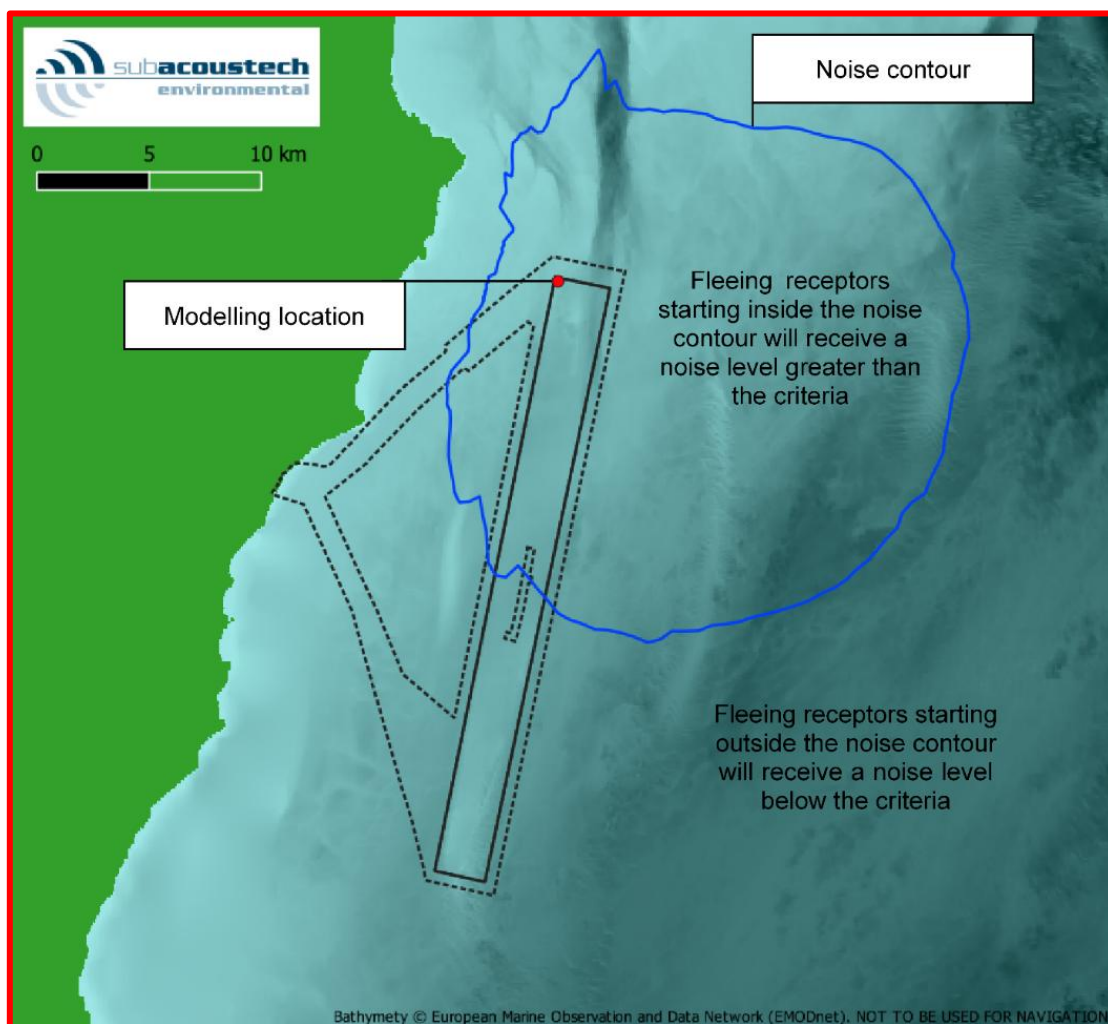


Figure 3-5 Plot showing a fleeing animal SEL_{cum} criteria contour and the areas where the cumulative noise exposure will exceed the impact criteria

Some modelling approaches include the effects of Acoustic Deterrent Devices (ADDs) that cause receptors to flee from the immediate area around the pile before activity commences. ~~Subacoustech Environmental's modelling approach does not include this, however the effects of using~~ This is not recommended for inclusion in modelling, as there are many ADDs with different performances and species-specific reactions and effectiveness, and instead the efficacy requirement from an ADD can ~~still~~ be inferred from the results. For example, if a receptor were to flee for 20 minutes from an ADD at a rate of 1.5 m/sms^{-1} , it would travel 1.8 km before piling begins. If a ~~calculated~~ cumulative $SEL_{LE,p,t}$ impact range ~~from INSPIRE~~ was ~~calculated to be~~ below 1.8 km, it can ~~safely~~ be assumed that the ADD will be effective in eliminating the risk of ~~injury~~ ~~on exceedance of the receptor threshold~~. The noise from an ADD is of a much lower level than impact piling, and as such ~~the~~ its overall effect on the SEL_{cum} total $L_{E,p,t}$ exposure ~~on a receptor~~ would be minimal.

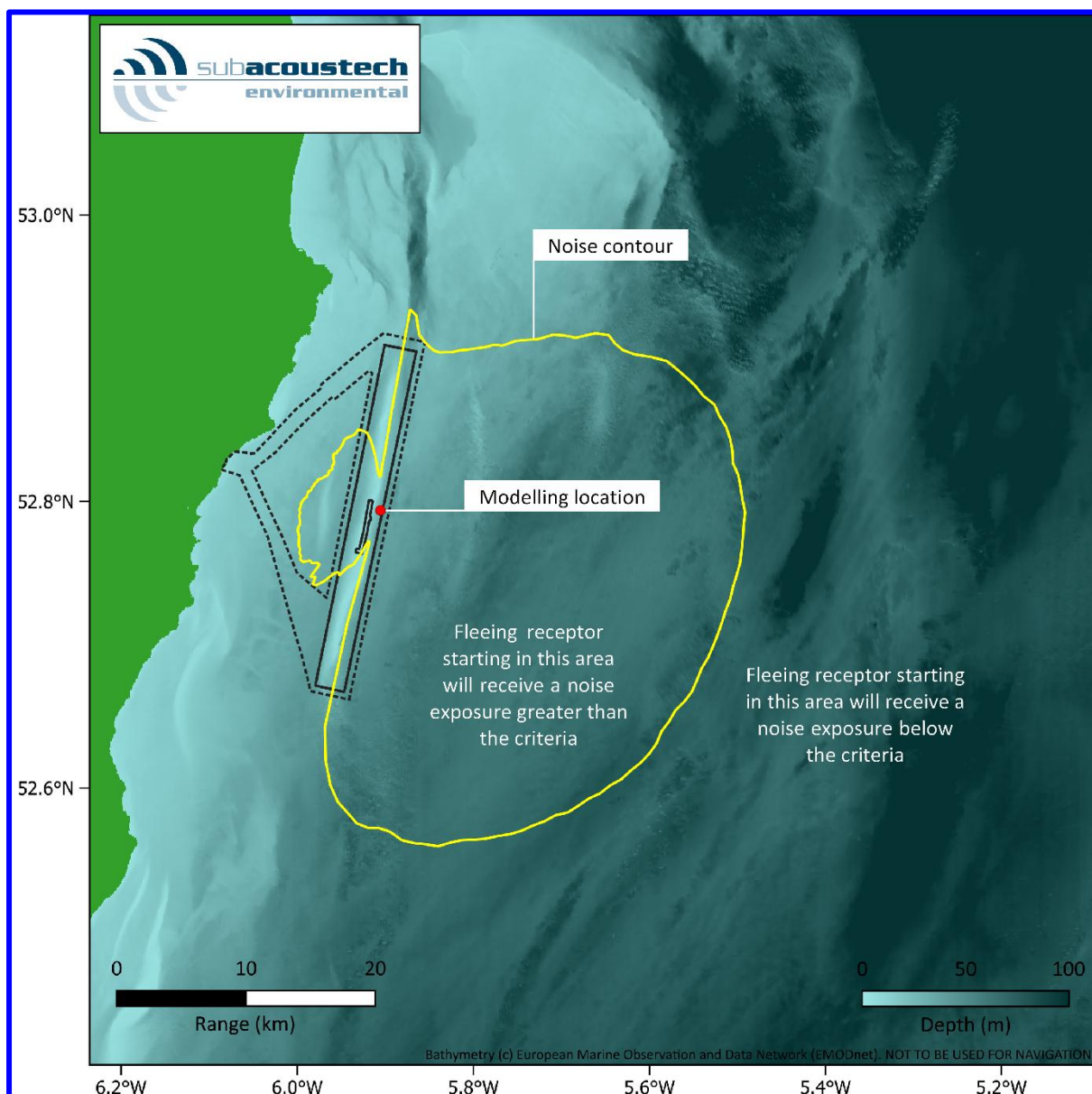


Figure 3-11: Example plot showing a fleeing animal $LE_{p,t}$ criteria contour and the areas where the cumulative noise exposure will exceed a given impact.

3.3.1 The effects of input effect of parameters on SEL_{cum} cumulative levels and fleeing receptors

As discussed in section 3.2.3, Parameters such as bathymetry, hammer blow energies, piling ramp up, strike rate and duration all have an effect on predicted noise levels. When considering $SEL_{cum,LE,p,t}$ and a fleeing animal model, some of these parameters can have a greater influence on the predicted noise levels than others.

~~Parameters like~~ Hammer blow energy can have a clear effect on the impact ranges, with higher energies resulting in ~~higher~~ high apparent source noise levels and therefore larger impact ranges. When considering cumulative noise levels, these higher levels are compounded, sometimes thousands of times, due to the number of pile strikes. With this in mind, the ramp up from ~~lower~~ to higher blow energies ~~to higher ones~~ requires careful consideration for fleeing ~~animals~~ receptors, as ~~the~~ levels while the ~~receptors are relatively~~ closer receptor is closer to the noise source will have a greater effect on the overall cumulative exposure level.

Linked to the effect of the ramp up is the strike rate, as the more pile strikes that occur while the receptor is close to the noise source, the greater the exposure and the greater effect it will have on the SEL_{cum} overall received $L_{E,p,t}$. The faster the strike rate, the shorter the distance the receptor can flee between each pile strike, which in turn leads to a greater exposure overall.

In general, the greatest contribution to the received exposure is found when a receptor is close to the noise source. If high blow energies or a fast strike rate are implemented at or close to the start of piling activities, it will tend to make impact ranges worse. This was one of the main considerations when creating the alternative ramp-up scenario in Table 3-3.

Another factor that can cause big differences in calculated impact ranges is the bathymetry, as deeper water results in less attenuation of noise (i.e., levels remain higher over greater distances). However, it is not feasible to limit piling activity in or near to deep water at ABWP2. This effect can be seen clearly in the example in Figure 3-11.

3.4 Precaution in underwater noise modelling

It is worth reiterating the precaution that is included in the modelling when assessing environmental impacts. In an effort to minimise the risk of under-prediction for potential impact ranges that occur in respect of sensitive marine mammal and fish receptors, conservative parameters are included for every element, which can be broken down into three basic steps for acoustic modelling (i.e. source, transmission and receiver). The possibility that the most conservative parameters could all occur together is highly unlikely, but necessary for the purposes of the assessment.

3.4.1 Source

The modelling locations were chosen to provide the greatest extents of the site, and specifically the locations likely to lead to underwater noise transmission over the greatest distances. The largest diameter for all types of pile has been used for the modelling scenarios. The maximum blow energies were used for a duration unlikely to occur in practice. The total piling duration is at the top of expectations and not expected to be exceeded on site.

3.4.2 Transmission path

Sound attenuates over distance from the source. The model considers fundamental noise spreading predictions adjusted to empirical data, accounting for frequency content, water depth, and other environmental factors, but fits to this data can still overestimate predicted levels (see section 3.1).

3.4.3 Receiver

The thresholds used for the sensitivity of marine mammals and fish are based on respective guidance for species groups (e.g., Southall *et al.*, 2019; Popper *et al.*, 2014). However, these tend to be precautionary in themselves. Frequency-specific hearing thresholds are not used for fish as they are with marine mammals, effectively assuming that fish are sensitive to sound at all frequencies, which is not the case. Also, the thresholds calculated for PTS and TTS are the 'onset' to these effects for both fish and marine mammals, which means that this is the threshold at which the effect starts to be detected in test species, rather than where this effect is widespread.

Concerning the flee speeds used in the modelling, studies have shown that these are typical swimming speeds (Williams, 2009), however flee speeds would be expected to be much faster during high noise conditions (McGarry *et al.*, 2017; Kastelein *et al.*, 2018). Using a faster flee speed would lead to much smaller cumulative impact ranges and consequently fewer impacted individuals.

The risk of PTS will not remain constant throughout an entire modelled PTS area either, although the model cannot account for this. The further away from the noise source the lower the risk will be, and thus at the PTS contour, this is effectively the position of onset of the risk of PTS.

Modelling does not include any assessment of impulsiveness, as criteria do not yet exist to specify a transition (Matei *et al.*, 2024). This means that any injury (or PTS) or TTS onset thresholds beyond a distance of the order of 3.5 km to 5 km (Hastie *et al.*, 2019, Matei *et al.*, 2024) are likely to over-estimate the risk to marine mammals, especially at greater distances.

All of these elements are not acting in isolation but will combine (i.e., at the greatest ranges there is a low risk of PTS, and reduced impulsivity, and are unlikely due to the slower-than-expected swim speeds), and contribute to the significant degree of precaution in the assessment.

4 Modelling results

This section presents the modelled ~~impact ranges for results from~~ impact piling noise to install monopile foundations at ABWP2 following the parameters detailed in section 3.2, ~~covering~~. The calculated impact ranges and areas cover the Southall *et al.* (2019) marine mammal criteria (section ~~2.3.1~~2.3.1) and ~~the~~ Popper *et al.* (2014) fish criteria (section ~~2.3.2~~2.3.2). ~~To aid navigation~~2.3.2). For completeness, Table 4-1 ~~contains~~ gives a list of the ~~impact range tables included in this section~~. The biggest results tables presented in sections 4.1 and 4.2. The largest modelled ranges from impact piling at ABWP2 are predicted ~~for the larger monopile scenarios at the SW and South OSP locations due to the combination of larger blow energies used and the at the~~ Central (WT28) location due to the deeper water at that location as well as proximity to ~~deep~~the deeper water ~~out~~in the Irish Sea to the ~~south and~~ east of ~~the site~~.

~~that location. For marine mammals, the maximum PTS injury ranges are predicted for LF cetaceans using the SEL_{cum} criteria, with ranges of up to 19 km. For VHF cetaceans, PTS ranges are predicted up to 10 km for the same scenario.~~

~~For fish, the largest recoverable injury ranges (203 dB SEL_{cum} threshold) are predicted to be 7.9 km assuming a stationary receptor; if a fleeing animal is assumed, these ranges reduce to less than 100 m. Maximum TTS ranges (186 dB SEL_{cum} threshold) are predicted up to 50 km for a stationary animal, reducing to 36 km for a fleeing animal.~~

The modelling results for the ~~first pile strike and the~~ Southall *et al.* (2019) non-impulsive criteria are presented in ~~Appendix A~~Appendix A.

~~For the results presented~~ Throughout this report, any predicted ranges smaller than 50 m ~~and areas less than 0.01 km² for single strike criteria, and ranges smaller than 100 m and areas less than 0.1 km² for cumulative criteria,~~ have not been presented in detail. At ranges this close to the noise source, the modelling processes are unable to model to a sufficient level of accuracy due to complex acoustic effects present near the ~~pile source~~. These ranges are given as “less than” this limit (e.g., “~~<100~~ 50 m”). Similarly, areas smaller than 0.01 km² have not been presented.

Table 4-1 ~~Summary:~~ List of the impact piling modelling results ~~tables~~ presented in this section.

Table (page)	Parameters (section) Scenario	Location	Criteria
Table 4-3 Summary of the 1 (p48)	NW-WTG location (4.1) foundations (Precautionary) (4.1)	11 m monopile, 4,000 kJ (North location (WT03) (4.1.1))	Southall <i>et al.</i> (2019) Unweighted SPL _{peak} L _{p,pk} (Impulsive)
Table 4-4 Summary of the 1 (p49)			Weighted SEL _{cum} L _{E,p,24h,wtd} (Impulsive)
Table 4-5 Summary of the 1 (p49)			NOAA (2005) Unweighted SPL _{RMS} L _{p,RMS} (Level B)
Table 4-6 Summary of the 5 (p50)		7 m monopile, 4,000 kJ	Popper <i>et al.</i> (2014) Unweighted SPL _{peak} L _{p,pk} (Pile driving)
Table 4-7 Summary of the 5 (p51)			Unweighted SEL _{cum} L _{E,p,24h} (Pile driving)
Table 4-8 Summary of the 7 (p54)		Southall <i>et al.</i> (2019) Unweighted SPL _{peak} L _{p,pk} (Impulsive)	

Table (page)	Parameters (section) scenario	Location	Criteria	
Table 4-9 Summary of the 3 (p 513)		(0) Central location (WT28) (4.1.2)	Weighted SEL_{cum}L_{E,p,24h,wtd} (Impulsive)	
Table 4-10 Summary of the 3 (p 523)			NOAA (2005) Unweighted SPL_{RMS}L_{p,RMS} (Level B)	
Table 4-11 Summary of the 10 (p 523)			Popper <i>et al.</i> (2014) Unweighted SPL_{peak}L_{p,pk} (Pile driving)	
Table 4-12 Summary of the 11 (p 523)			Popper <i>et al.</i> (2014) Unweighted SEL_{cum}L_{E,p,24h} (Pile driving)	
Table 4-13 Summary of the 12 (p 533)	C-WTG location (0)		Southall <i>et al.</i> (2019) Unweighted SPL_{peak}L_{p,pk} (Impulsive)	
Table 4-14 Summary of the 13 (p 533)	11-m monopile, 4,000 kJ (0) South west location (WT53) (4.1.3)		Southall <i>et al.</i> (2019) Weighted SEL_{cum}L_{E,p,24h,wtd} (Impulsive)	
Table 4-15 Summary of the 14 (p 543)		NOAA (2005) Unweighted SPL_{RMS}L_{p,RMS} (Level B)		
Table 4-16 Summary of the 15 (p 543)		Popper <i>et al.</i> (2014) Unweighted SPL_{peak}L_{p,pk} (Pile driving)		
Table 4-17 Summary of the 16 (p 543)		Popper <i>et al.</i> (2014) Unweighted SEL_{cum}L_{E,p,24h} (Pile driving)		
Table 4-17 (p 573)		7-m monopile, 4,000 kJ	Southall <i>et al.</i> (2019) Unweighted SPL_{peak} Weighted L_{E,p,24h,wtd} (Impulsive)	
Table 4-18 (p 593)		(0) North location (WT03) (4.1.4.1)	Popper <i>et al.</i> (2014) Weighted SEL_{cum} (Impulsive) Unweighted L_{E,p,24h} (Pile driving)	
Table 4-19 (p 593)		WTG foundations (Alternative ramp-up) (4.1.4)	Central location (WT28) (4.1.4.2)	NOAA (2005) Southall <i>et al.</i> (2019) Unweighted SPL_{RMS} (Level B) Weighted L_{E,p,24h,wtd} (Impulsive)
Table 4-20 (p 593)			South west location (WT53) (4.1.4.3)	Popper <i>et al.</i> (2014) Unweighted SPL_{peak}L_{E,p,24h} (Pile driving)
Table 4-21 (p 57)			South west location (WT53) (4.1.4.3)	Southall <i>et al.</i> (2019) Weighted L_{E,p,24h,wtd} (Impulsive)
Table 4-22 (p 593)			South west location (WT53) (4.1.4.3)	Popper <i>et al.</i> (2014) Unweighted SEL_{cum}L_{E,p,24h} (Pile driving)
Table 4-18 Summary of the 13 (p 613)	SW-WTG location (4.1.5.2) foundations (Mitigated) (4.1.5)	(0) North location (WT03) (4.1.5.1)	Southall <i>et al.</i> (2019) Unweighted SPL_{peak}L_{p,pk} (Impulsive)	
Table 4-19 Summary of the 14 (p 613)			Southall <i>et al.</i> (2019) Weighted SEL_{cum}L_{E,p,24h,wtd} (Impulsive)	
Table 4-20 Summary of the 15 (p 613)			NOAA (2005) Unweighted SPL_{RMS}L_{p,RMS} (Level B)	
Table 4-21 Summary of the 16 (p 623)			Popper <i>et al.</i> (2014) Unweighted SPL_{peak}L_{p,pk} (Pile driving)	

Table (page)	Parameters (section) scenario	Location	Criteria		
Table 4-22 Summary of the 27 (p 63.5)		7 m monopile, 6,600 kJ (0 Central location (WT28) (4.1.5.2))	Unweighted $SEL_{cumL,E,p,24h}$ (Pile driving)		
Table 4-23 Summary of the 28 (p 63.6)			Southall <i>et al.</i> (2019)	Unweighted $SPL_{peakL,p,pk}$ (Impulsive)	
Table 4-24 Summary of the 29 (p 63.7)				Weighted $SEL_{cumL,E,p,24h,wtd}$ (Impulsive)	
Table 4-25 Summary of the 30 (p 64.1)				NOAA (2005)	Unweighted $SPL_{RMSL,p,RMS}$ (Level B)
Table 4-26 Summary of the 31 (p 64.2)				Popper <i>et al.</i> (2014)	Unweighted $SPL_{peakL,p,pk}$ (Pile driving)
Table 4-27 Summary of the 32 (p 64.3)					Unweighted $SEL_{cumL,E,p,24h}$ (Pile driving)

Table (page)	Scenario	Location	Criteria	
Table 4-28 Summary of the 33 (p 65.1)	North OSP location (4.2.1 WTG foundations (Mitigated) (4.1.5))	14 m monopile, 4,000 kJ (0 South west location (WT53) (4.1.5.3))	Southall <i>et al.</i> (2019)	Unweighted $SPL_{peakL,p,pk}$ (Impulsive)
Table 4-29 Summary of the 34 (p 66.1)				Weighted $SEL_{cumL,E,p,24h,wtd}$ (Impulsive)
Table 4-30 Summary of the 35 (p 66.2)			NOAA (2005)	Unweighted $SPL_{RMSL,p,RMS}$ (Level B)
Table 4-31 Summary of the 36 (p 66.3)			Popper <i>et al.</i> (2014)	Unweighted $SPL_{peakL,p,pk}$ (Pile driving)
Table 4-32 Summary of the 37 (p 66.4)				Unweighted $SEL_{cumL,E,p,24h}$ (Pile driving)
Table 4-33 Summary of the 38 (p 68.1)	OSP foundations (Precautionary) (4.2)	7 m monopile, 4,000 kJ (0 North OSP location (4.2.1))	Southall <i>et al.</i> (2019)	Unweighted $SPL_{peakL,p,pk}$ (Impulsive)
Table 4-34 Summary of the 39 (p 68.2)				Weighted $SEL_{cumL,E,p,24h,wtd}$ (Impulsive)
Table 4-35 Summary of the 40 (p 68.3)			NOAA (2005)	Unweighted $SPL_{RMSL,p,RMS}$ (Level B)
Table 4-36 Summary of the 41 (p 68.4)			Popper <i>et al.</i> (2014)	Unweighted $SPL_{peakL,p,pk}$ (Pile driving)
Table 4-37 Summary of the 42 (p 70.1)				Unweighted $SEL_{cumL,E,p,24h}$ (Pile driving)
Table 4-38 Summary of the 43 (p 71.1)	South OSP location (4.2.3.2)	14 m monopile, 6,600 kJ (0 South OSP location (4.2.2))	Southall <i>et al.</i> (2019)	Unweighted $SPL_{peakL,p,pk}$ (Impulsive)

Table (page)	Scenario	Location	Criteria
Table 4-39 Summary of the 44 (p735)			Weighted $SEL_{cum}L_{E,p,24h,wtd}$ (Impulsive)
Table 4-40 Summary of the 45 (p735)			NOAA (2005) Unweighted $SPL_{RMS}L_{p,RMS}$ (Level B)
Table 4-41 Summary of the 46 (p735)			Popper <i>et al.</i> (2014) Unweighted $SPL_{peak}L_{p,pk}$ (Pile driving)
Table 4-47 (p737)			Unweighted $SEL_{cum}L_{E,p,24h}$ (Pile driving)
Table 4-48 (p70)	OSP foundations (Alternative ramp-up) (4.2.3)	North OSP location (4.2.3.1)	Southall <i>et al.</i> (2019) Weighted $L_{E,p,24h,wtd}$ (Impulsive)
Table 4-42 Summary of the 49 (p70)		Popper <i>et al.</i> (2014) Unweighted $L_{E,p,24h}$ (Pile driving)	
Table 4-50 (p71)		South OSP location (4.2.3.2)	Southall <i>et al.</i> (2019) Weighted $L_{E,p,24h,wtd}$ (Impulsive)
Table 4-51 (p71)		Popper <i>et al.</i> (2014) Unweighted $L_{E,p,24h}$ (Pile driving)	
Table 4-43 Summary of the 52 (p757)	OSP foundations (Mitigated) (4.2.4)	7 m monopile, 6,600 kJ (0) North OSP location (4.2.4.1)	Southall <i>et al.</i> (2019) Unweighted $SPL_{peak}L_{p,pk}$ (Impulsive)
Table 4-44 Summary of the 53 (p757)			Weighted $SEL_{cum}L_{E,p,24h,wtd}$ (Impulsive)
Table 4-45 Summary of the 54 (p757)			NOAA (2005) Unweighted $SPL_{RMS}L_{p,RMS}$ (Level B)
Table 4-46 Summary of the 55 (p767)			Popper <i>et al.</i> (2014) Unweighted $SPL_{peak}L_{p,pk}$ (Pile driving)
Table 4-47 Summary of the 56 (p73)		Unweighted $L_{E,p,24h}$ (Pile driving)	
Table 4-48 Summary of the 57 (p75)		Southall <i>et al.</i> (2019) Unweighted $L_{p,pk}$ (Impulsive)	
Table 4-49 Summary of the 58 (p75)		Weighted $L_{E,p,24h,wtd}$ (Impulsive)	
Table 4-50 Summary of the 59 (p75)		NOAA (2005) Unweighted $L_{p,RMS}$ (Level B)	
Table 4-51 Summary of the 60 (p76)		Popper <i>et al.</i> (2014) Unweighted $L_{p,pk}$ (Pile driving)	
Table 4-52 Summary of the 61 (p777)		Unweighted $SEL_{cum}L_{E,p,24h}$ (Pile driving)	

Table 4-3 ~~Summary of the 2~~ to Table 4-52 ~~Summary of the 61~~ present the impact piling modelling results for ABWP2, covering WTG and OSP foundations. For the WTG foundation scenarios, the largest predicted impact ranges were calculated for the LF cetaceans at the Central (WT28) modelling location, with maximum PTS ranges

of 7.1 km. For fish, maximum recoverable injury ($203 \text{ dB } L_{E,p,24h}$) ranges of up to 6.0 km were predicted for stationary receptors. When a fleeing receptor was considered, this range reduced to less than 50 m.

When considering the Alternative ramp-up scenario (Table 3-3) the maximum PTS ranges reduced to 870 m for LF cetacean and 5.5 km for recoverable injury in stationary fish. The use of mitigation (Mitigated scenarios) also reduced the maximum impact ranges down to 720 m for LF cetaceans PTS and 3.4 km for stationary fish. Note that the $L_{p,pk}$ impact ranges for the alternative ramp up have not been run as they use the same maximum blow energy to the worst case and thus would have identical impact ranges.

In accordance with the Danish requirements, an “ r_{safe} ” distance of 1 km was targeted for the marine mammal PTS impact ranges, which is mitigated using an ADD for key marine mammal species. As above, this is achieved using the WTG alternative ramp up, and also using the Precautionary scenario with a -4 dB noise reduction (the Mitigated scenario).

4.1 Predicted noise levels at 750 m from the noise source

In addition to the source levels given in section 3.2.5, it is useful to look at the potential noise levels at a range of 750 m from the noise source, which although not a requirement in the Irish sector, is a common consideration for underwater noise studies at offshore windfarms. It has the added advantage of being comparable with other modelling or on-site measurements. A summary of the modelled unweighted levels at a range of 750 m are given in Table 4-2 Summary of the maximum predicted unweighted SPL_{peak} and SEL_{ss} noise levels at a range of 750 m from the noise source when considering maximum hammer blow energy considering the transect with the greatest noise transmission at each location while piling at the maximum hammer energy. These show the variation between the different locations, but that there is only a minimal difference between the different monopile diameters.

Table 4-2 Summary of the maximum predicted unweighted SPL_{peak} and SEL_{ss} noise levels at a range of 750 m from the noise source when considering maximum hammer blow energy

Predicted level at 750 m range	Location	Large monopile foundation 11 m / 4,000 kJ (NW and C-WTG) 11 m / 6,600 kJ (SW-WTG) 14 m / 4,000 kJ (North-OSP) 14 m / 6,600 kJ (South-OSP)	Smaller monopile foundation 7 m / 4,000 kJ (NW, C-WTG and North-OSP) 7 m / 6,600 kJ (SW-WTG and South-OSP)
Unweighted SPL_{peak}	NW-WTG	200.3 dB re 1 μ Pa	200.3 dB re 1 μ Pa
	C-WTG	202.4 dB re 1 μ Pa	202.4 dB re 1 μ Pa
	SW-WTG	203.1 dB re 1 μ Pa	203.0 dB re 1 μ Pa
	North-OSP	201.1 dB re 1 μ Pa	201.1 dB re 1 μ Pa
	South-OSP	202.3 dB re 1 μ Pa	202.3 dB re 1 μ Pa
Unweighted SEL_{ss}	NW-WTG	181.5 dB re 1 μ Pa	181.5 dB re 1 μ Pa
	C-WTG	183.5 dB re 1 μ Pa	183.5 dB re 1 μ Pa
	SW-WTG	184.3 dB re 1 μ Pa	184.2 dB re 1 μ Pa
	North-OSP	182.3 dB re 1 μ Pa	182.2 dB re 1 μ Pa
	South-OSP	183.7 dB re 1 μ Pa	183.6 dB re 1 μ Pa

For the OSP foundation scenarios, maximum impact ranges were predicted at the deeper North OSP location, with maximum ranges for marine mammals calculated for LF cetaceans with PTS ranges of 3.0 km. For fish the maximum recoverable injury ranges for stationary fish were 4.9 km, reducing to less than 50 m when a fleeing receptor was considered.

4.1 4.2 NW-WTG location foundations

4.2.1 11 m monopile foundation

4.1.1 North location (WT03)

Table 4-3 Summary of the 2: Unweighted $SPL_{peak}L_{p,pk}$ impact areas and ranges for marine mammals using the Southall et al. (2019) impulsive criteria for the 11 m monopile maximum expected hammer blow energy during the WTG foundation modelling at the NW-WTG installation scenario at the North (WT03) modelling location.

Southall et al. (2019) Unweighted $SPL_{peak}L_{p,pk}$		Area WTG foundation (maximum range Minimum range Mean range energy)			
		Area	Maximum range	Minimum range	Mean range
PTS (impulsive)	LF (219 dB)	< 0.01 km ²	< 50 m	< 50 m	< 50 m
	HF (230 dB)	< 0.01 km ²	< 50 m	< 50 m	< 50 m
	VHF (202 dB)	0.731.0 km ²	500580 m	480550 m	480560 m
	PCW (218 dB)	0.01 km ²	< 5060 m	< 5060 m	< 5060 m
TTS (impulsive)	LF (213 dB)	0.030.05 km ²	90120 m	90120 m	90120 m
	HF (224 dB)	< 0.01 km ²	< 50 m	< 50 m	< 50 m
	VHF (196 dB)	4.34.9 km ²	1.21.4 km	1.11.2 km	1.21.3 km
	PCW (212 dB)	0.040.06 km ²	110140 m	110140 m	110140 m

Table 4-4 Summary of the 3: Weighted $SEL_{cum}L_{E,p,24h,wtd}$ impact areas and ranges for marine mammals using the Southall et al. (2019) impulsive criteria for the 11 m monopile WTG foundation modelling at the NW-WTG installation scenario at the North (WT03) modelling location assuming a fleeing animal receptors.

Southall et al. (2019) Weighted SEL_{cum} Weighted $L_{E,p,24h,wtd}$		Area Maximum range Minimum range Mean range WTG foundation			
		Area	Maximum range	Minimum range	Mean range
PTS (impulsive)	LF (183 dB)	5816 km ²	7.64.1 km	200100 m	3.41.7 km
	HF (185 dB)	< 0.40.01 km ²	< 10050 m	< 10050 m	< 10050 m
	VHF (155 dB)	310.6 km ²	4.6 km820 m	1.1 km< 50 m	2.9 km330 m
	PCW (185 dB)	< 0.40.01 km ²	< 10050 m	< 10050 m	< 10050 m
TTS (impulsive)	LF (168 dB)	4,5003,100 km ²	6353 km	3.52.4 km	3025 km
	HF (170 dB)	< 0.40.01 km ²	< 10050 m	< 10050 m	< 10050 m
	VHF (140 dB)	2,0001,000 km ²	4230 km	4.42.6 km	2216 km
	PCW (170 dB)	380200 km ²	1813 km	2.8 km920 m	9.87.1 km

Table 4-5 Summary of the 4: Unweighted $SPL_{RMS}L_{p,RMS}$ impact areas and ranges for marine mammals using the NOAA (2005) Level B impulsive behavioural disturbance criteria for the 11 m monopile maximum expected hammer blow energy during the WTG foundation modelling at the NW-WTG installation scenario at the North (WT03) modelling location.

NOAA (2005) Unweighted $SPL_{RMS}L_{p,RMS}$		Area WTG foundation (maximum range Minimum range range			
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		Mean range energy)			
		Area	Maximum range	Minimum range	Mean range
Level B	160 dB	980760 km ²	2622 km	5-96.3 km	1615 km

Table 4-6 Summary of the 5: Unweighted $SPL_{peak,Lp,pk}$ impact areas and ranges for fish using the Popper et al. (2014) pile driving criteria for the 11 m monopile maximum expected hammer blow energy during the WTG foundation modelling at the NW WTG installation scenario at the North (WT03) modelling location.

Popper et al. (2014) Unweighted $SPL_{peak,Lp,pk}$		Area WTG foundation (maximum range Minimum range Mean range energy)			
213 dB		0.03 km ² Area	90 m Maximum range	90 m Minimum range	90 m Mean range
Pile driving	213 dB	0.05 km ²	120 m	120 m	120 m
207 dB	207 dB	0.160.25 km ²	230290 m	230280 m	230280 m

Table 4-7 Summary of the 6: Unweighted $SEL_{cumLE,p,24h}$ impact areas and ranges for marine mammals/fish using the Popper et al. (2014) pile driving criteria for the 11-m monopile WTG foundation modelling at the NW-WTG installation scenario at the North (WT03) modelling location assuming both fleeing and stationary animals/receptors.

Popper et al. (2014) Unweighted $SEL_{cumLE,p,24h}$		Area Maximum range Minimum range Mean range range WTG foundation			
		Area	Maximum range	Minimum range	Mean range
Fleeing Fleeing (1.5 m/s ^{ms⁻¹})	219 dB	< 0.40.01 km ²	< 40050 m	< 40050 m	< 40050 m
	216 dB	< 0.40.01 km ²	< 40050 m	< 40050 m	< 40050 m
	210 dB	< 0.40.01 km ²	< 40050 m	< 40050 m	< 40050 m
	207 dB	< 0.40.01 km ²	< 40050 m	< 40050 m	< 40050 m
	203 dB	< 0.40.01 km ²	< 40050 m	< 40050 m	< 40050 m
	186 dB	440580 km ²	4922 km	2.81.9 km	4012 km
Stationary Stationary (0 m/s ^{ms⁻¹})	219 dB	0.60.68 km ²	500470 m	400460 m	440470 m
	216 dB	1.31.8 km ²	700770 m	600730 m	700750 m
	210 dB	6.811 km ²	1.62.0 km	1.41.7 km	1.51.9 km
	207 dB	1525 km ²	2.43.1 km	2.02.4 km	2.22.9 km
	203 dB	4268 km ²	4.15.5 km	3.23.3 km	3.64.8 km
	186 dB	1,2001,500 km ²	2932 km	5.87.5 km	4824 km

4.2.2 7 m monopile foundation

4.1.2 Central location (WT28)

Table 4-8 Summary of the 7: Unweighted $SPL_{peakLp,pk}$ impact areas and ranges for marine mammals using the Southall et al. (2019) impulsive criteria for the 7-m monopile maximum expected hammer blow energy during the WTG foundation modelling at the NW-WTG installation scenario at the Central (WT28) modelling location.

Southall et al. (2019) Unweighted $SPL_{peakLp,pk}$		Area WTG foundation (maximum range Minimum range Mean range range energy)			
		Area	Maximum range	Minimum range	Mean range
PTS (impulsive)	LF (219 dB)	< 0.01 km ²	< 5060 m	< 5060 m	< 5060 m
	HF (230 dB)	< 0.01 km ²	< 50 m	< 50 m	< 50 m
	VHF (202 dB)	0.71.3 km ²	490680 m	470570 m	480630 m
	PCW (218 dB)	< 0.01 km ²	< 5070 m	< 5070 m	< 5070 m
TTS (impulsive)	LF (213 dB)	0.030.06 km ²	90140 m	90130 m	90130 m
	HF (224 dB)	< 0.01 km ²	< 50 m	< 50 m	< 50 m
	VHF (196 dB)	4.25.8 km ²	1.21.6 km	1.1 km960 m	1.21.3 km
	PCW (212 dB)	0.030.07 km ²	110160 m	110150 m	110150 m

Table 4-9 Summary of the 8: Weighted $SEL_{cumLE,p,24h,wtd}$ impact areas and ranges for marine mammals using the Southall et al. (2019) impulsive criteria for the 7-m monopile WTG foundation modelling at the NW-WTG installation scenario at the Central (WT28) modelling location assuming a fleeing animal/receptors.

Southall et al. (2019) Weighted SEL_{cum} Weighted $L_{E,p,24h}$		Area Maximum range Minimum range Mean range WTG foundation			
		Area	Maximum range	Minimum range	Mean range
PTS (Impulsive)	LF (183 dB)	5657 km ²	7.57.1 km	200110 m	3.43.0 km
	HF (185 dB)	< 0.40.01 km ²	< 40050 m	< 40050 m	< 40050 m
	VHF (155 dB)	343.3 km ²	4.61.7 km	1.1 km < 50 m	2.9 km 720 m
	PCW (185 dB)	< 0.40.01 km ²	< 40050 m	< 40050 m	< 40050 m
TTS (Impulsive)	LF (168 dB)	4,500 km ²	6360 km	3.52.7 km	30 km
	HF (170 dB)	< 0.40.01 km ²	< 40050 m	< 40050 m	< 40050 m
	VHF (140 dB)	2,0001,600 km ²	4235 km	4.42.9 km	2219 km
	PCW (170 dB)	380340 km ²	1816 km	2.8 km 820 m	9.78.7 km

Table 4-10 Summary of the 9: Unweighted SPL_{RMS} $L_{p,RMS}$ impact areas and ranges for marine mammals using the NOAA (2005) Level B impulsive behavioural disturbance criteria for the 7-m monopile maximum expected hammer blow energy during the WTG foundation modelling at the NW WTG installation scenario at the Central (WT28) modelling location.

NOAA (2005) Unweighted SPL_{RMS} $L_{p,RMS}$		Area WTG foundation (maximum range Minimum range Mean range energy)			
		Area	Maximum range	Minimum range	Mean range
Level B	160 dB	9701,100 km ²	2526 km	5.96.6 km	1618 km

Table 4-11 Summary of the 10: Unweighted SPL_{peak} $L_{p,pk}$ impact areas and ranges for fish using the Popper et al. (2014) pile driving criteria for the 7-m monopile maximum expected hammer blow energy during the WTG foundation modelling installation scenario at the NW Central (WT28) modelling location.

Popper et al. (2014) Unweighted SPL_{peak} $L_{p,pk}$		Area WTG foundation (maximum range Minimum range Mean range energy)			
		Area	Maximum range	Minimum range	Mean range
213-dB		0.03 km ² Area	90-m Maximum range	90-m Minimum range	90-m Mean range
Pile driving	213 dB	0.06 km ²	140 m	130 m	130 m
207-dB		0.460.31 km ²	230330 m	220300 m	230320 m

Table 4-12 Summary of the 11: Unweighted SEL_{cum} $L_{E,p,24h}$ impact areas and ranges for marine mammals fish using the Popper et al. (2014) pile driving criteria for the 7-m monopile WTG foundation modelling at the NW WTG installation scenario at the Central (WT28) modelling location assuming both fleeing and stationary animals receptors.

Popper et al. (2014) Unweighted SEL_{cum} Unweighted $L_{E,p,24h}$		Area Maximum range Minimum range Mean range WTG foundation			
		Area	Maximum range	Minimum range	Mean range
Fleeing Fleeing (1.5 m/s ms ⁻¹)	219 dB	< 0.4 0.01 km ²	< 400 50 m	< 400 50 m	< 400 50 m
	216 dB	< 0.4 0.01 km ²	< 400 50 m	< 400 50 m	< 400 50 m
	210 dB	< 0.4 0.01 km ²	< 400 50 m	< 400 50 m	< 400 50 m
	207 dB	< 0.4 0.01 km ²	< 400 50 m	< 400 50 m	< 400 50 m
	203 dB	< 0.4 0.01 km ²	< 400 50 m	< 400 50 m	< 400 50 m
	186 dB	420 1,000 km ²	19 28 km	2.7 2.5 km	40 15 km
Stationary Stationary (0 m/s ms ⁻¹)	219 dB	0.6 0.79 km ²	500 520 m	400 480 m	400 500 m
	216 dB	1.3 2.0 km ²	700 840 m	600 710 m	600 810 m
	210 dB	6.5 11 km ²	1.5 2.2 km	1.4 1.3 km	1.4 2.0 km
	207 dB	15 26 km ²	2.4 3.4 km	2.0 1.7 km	2.2 3.0 km
	203 dB	44 73 km ²	4.0 6.0 km	3.2 2.4 km	3.6 4.9 km
	186 dB	1,200 2,300 km ²	28 39 km	5.8 7.3 km	18 27 km

4.1.3 [South west location \(WT53\)](#)

4.3 C-WTG location

4.3.1 11-m monopile foundation

Table 4-13 ~~Summary of the 12:~~ Unweighted SPL_{peak} $L_{p,pk}$ impact areas and ranges for marine mammals using the Southall et al. (2019) impulsive criteria for the 11-m monopile maximum expected hammer blow energy during the WTG foundation modelling at the C-WTG installation scenario at the South west (WT53) modelling location.

Southall et al. (2019) Unweighted SPL_{peak} $L_{p,pk}$		Area WTG foundation (maximum range Minimum range Mean range energy)			
		Area	Maximum range	Minimum range	Mean range
PTS (Impulsive)	LF (219 dB)	< 0.01 km ²	< 50 m	< 50 m	< 50 m
	HF (230 dB)	< 0.01 km ²	< 50 m	< 50 m	< 50 m
	VHF (202 dB)	1.5 0.9 km ²	700 580 m	600 480 m	600 530 m
	PCW (218 dB)	0.01 km ²	50 60 m	50 60 m	50 60 m
TTS (Impulsive)	LF (213 dB)	0.04 km ²	120 m	120 110 m	120 m
	HF (224 dB)	< 0.01 km ²	< 50 m	< 50 m	< 50 m
	VHF (196 dB)	9.5 4.6 km ²	1.8 1.4 km	1.7 1.0 km	1.7 1.2 km
	PCW (212 dB)	0.06 km ²	140 m	140 130 m	140 130 m

Table 4-14 ~~Summary of the 13:~~ Weighted SEL_{cum} $L_{E,p,24h,wtg}$ impact areas and ranges for marine mammals using the Southall et al. (2019) impulsive criteria for the 11-m monopile WTG foundation modelling at the C-WTG installation scenario at the South west (WT53) modelling location assuming a fleeing animal receptors.

Southall et al. (2019) Weighted SEL_{cum} $L_{E,p,24h}$		Area Maximum range Minimum range Mean range WTG foundation			
		Area	Maximum range	Minimum range	Mean range
PTS (Impulsive)	LF (183 dB)	40037 km ²	175.8 km	3.2 km	90 m
	HF (185 dB)	< 0.1 km ²	< 100 m	< 100 m	< 100 m
	VHF (155 dB)	1502.8 km ²	9.1 km	3.7 km	< 50 m
	PCW (185 dB)	0.3 km ²	< 50 m	< 100 m	< 50 m
TTS (Impulsive)	LF (168 dB)	8,5004,100 km ²	8659 km	8.3 km	15 km
	HF (170 dB)	< 0.1 km ²	< 100 m	< 100 m	< 100 m
	VHF (140 dB)	4,5001,600 km ²	5634 km	8.9 km	1.7 km
	PCW (170 dB)	1,500370 km ²	3216 km	7.2 km	570 m

Table 4-15 Summary of the 14: Unweighted SPL_{RMS} $L_{p,RMS}$ impact areas and ranges for marine mammals using the NOAA (2005) Level B impulsive behavioural disturbance criteria for the 11-m monopile maximum expected hammer blow energy during the WTG foundation modelling at the C-WTG installation scenario at the South west (WT53) modelling location.

NOAA (2005) Unweighted SPL_{RMS} $L_{p,RMS}$		Area WTG foundation (maximum range Minimum range Mean range energy)			
		Area	Maximum range	Minimum range	Mean range
Level B	160 dB	2,3001,100 km ²	3724 km	144.9 km	2518 km

Table 4-16 Summary of the 15: Unweighted SPL_{peak} $L_{p,pk}$ impact areas and ranges for fish using the Popper et al. (2014) pile driving criteria for the 11-m monopile maximum expected hammer blow energy during the WTG foundation modelling at the C-WTG installation scenario at the South west (WT53) modelling location.

Popper et al. (2014) Unweighted SPL_{peak} $L_{p,pk}$		Area WTG foundation (maximum range Minimum range Mean range energy)			
		Area	Maximum range	Minimum range	Mean range
213-dB Pile driving	213 dB	0.04 km ²	120 m	120 m	120 m
207-dB	207 dB	0.30.22 km ²	340280 m	340250 m	340270 m

Table 4-17 Summary of the 16: Unweighted SEL_{cum} $L_{E,p,24h}$ impact areas and ranges for marine mammals fish using the Popper et al. (2014) pile driving criteria for the 11-m monopile WTG foundation modelling at the C-WTG installation scenario at the South west (WT53) modelling location assuming both fleeing and stationary animals receptors.

Popper et al. (2014) Unweighted $L_{E,p,24h}$		WTG foundation			
		Area	Maximum range	Minimum range	Mean range
Fleeing (1.5 ms ⁻¹)	219 dB	< 0.01 km ²	< 50 m	< 50 m	< 50 m
	216 dB	< 0.01 km ²	< 50 m	< 50 m	< 50 m
	210 dB	< 0.01 km ²	< 50 m	< 50 m	< 50 m

	207 dB	< 0.01 km ²	< 50 m	< 50 m	< 50 m
	203 dB	< 0.01 km ²	< 50 m	< 50 m	< 50 m
	186 dB	890 km ²	25 km	1.3 km	15 km
Stationary (0 ms ⁻¹)	219 dB	0.63 km ²	470 m	420 m	450 m
	216 dB	1.6 km ²	770 m	660 m	720 m
	210 dB	11 km ²	2.0 km	1.6 km	1.8 km
	207 dB	26 km ²	3.1 km	2.1 km	2.8 km
	203 dB	78 km ²	5.4 km	2.7 km	4.9 km
	186 dB	2,100 km ²	36 km	5.4 km	26 km

4.1.4 Alternative ramp-up

4.1.4.1 North location (WT03)

Table 4-17: Weighted $L_{E,p,24h,wtd}$ impact areas and ranges for marine mammals using the Southall et al. (2019) impulsive criteria for the WTG foundation installation (alternative ramp-up) scenario at the North (WT03) modelling location assuming fleeing receptors.

Southall et al. (2019) Weighted $L_{E,p,24h,wtd}$		WTG foundation (alternative ramp-up)			
		Area	Maximum range	Minimum range	Mean range
PTS (Impulsive)	LF (183 dB)	0.04 km ²	170 m	60 m	100 m
	HF (185 dB)	< 0.01 km ²	< 50 m	< 50 m	< 50 m
	VHF (155 dB)	< 0.01 km ²	< 50 m	< 50 m	< 50 m
	PCW (185 dB)	< 0.01 km ²	< 50 m	< 50 m	< 50 m
TTS (Impulsive)	LF (168 dB)	2,100 km ²	44 km	1.3 km	19 km
	HF (170 dB)	< 0.01 km ²	< 50 m	< 50 m	< 50 m
	VHF (140 dB)	690 km ²	25 km	620 m	12 km
	PCW (170 dB)	78 km ²	8.8 km	70 m	3.9 km

Table 4-18: Unweighted $L_{E,p,24h}$ impact areas and ranges for fish using the Popper et al. (2014) pile driving criteria for the WTG foundation installation (alternative ramp-up) scenario at the North (WT03) modelling location assuming both fleeing and stationary receptors.

Popper et al. (2014) Unweighted-SEL _{cum} $L_{E,p,24h}$	Area Maximum range Minimum range Mean range range WTG foundation (alternative ramp-up)				
	Area	Maximum range	Minimum range	Mean range	
Fleeing Fleeing (1.5 m/s ms ⁻¹)	219 dB	< 0.4 0.01 km ²	< 400 50 m	< 400 50 m	< 400 50 m
	216 dB	< 0.4 0.01 km ²	< 400 50 m	< 400 50 m	< 400 50 m
	210 dB	< 0.4 0.01 km ²	< 400 50 m	< 400 50 m	< 400 50 m
	207 dB	< 0.4 0.01 km ²	< 400 50 m	< 400 50 m	< 400 50 m
	203 dB	< 0.4 0.01 km ²	< 400 50 m	< 400 50 m	< 400 50 m
	186 dB	340 km ²	18 km	320 m	8.5 km
Stationary (0 ms ⁻¹)	219 dB	0.57 km ²	430 m	420 m	430 m
	216 dB	1.5 km ²	710 m	670 m	690 m
	210 dB	9.6 km ²	1.9 km	1.6 km	1.8 km
	207 dB	22 km ²	2.9 km	2.3 km	2.7 km
	203 dB	59 km ²	5.1 km	3.1 km	4.5 km
	186 dB	1,400 km ²	31 km	6.9 7.5 km	19 23 km

4.1.4.2 Central location (WT28)

Table 4-19: Weighted $L_{E,p,24h,wtd}$ impact areas and ranges for marine mammals using the Southall et al. (2019) impulsive criteria for the WTG foundation installation (alternative ramp-up) scenario at the Central (WT28) modelling location assuming fleeing receptors.

Southall et al. (2019) Weighted $L_{E,p,24h,wtd}$	WTG foundation (alternative ramp-up)				
	Area	Maximum range	Minimum range	Mean range	
PTS (Impulsive)	LF (183 dB)	0.58 km ²	870 m	70 m	300 m
	HF (185 dB)	< 0.01 km ²	< 50 m	< 50 m	< 50 m
	VHF (155 dB)	< 0.01 km ²	< 50 m	< 50 m	< 50 m
	PCW (185 dB)	< 0.01 km ²	< 50 m	< 50 m	< 50 m
TTS (Impulsive)	LF (168 dB)	3,200 km ²	51 km	1.7 km	24 km
	HF (170 dB)	< 0.01 km ²	< 50 m	< 50 m	< 50 m
	VHF (140 dB)	1,100 km ²	30 km	870 m	15 km
	PCW (170 dB)	170 km ²	12 km	80 m	5.5 km

Table 4-20: Unweighted $L_{E,p,24h}$ impact areas and ranges for fish using the Popper et al. (2014) pile driving criteria for the WTG foundation installation (alternative ramp-up) scenario at the Central (WT28) modelling location assuming both fleeing and stationary receptors.

Popper et al. (2014) Unweighted $L_{E,p,24h}$	WTG foundation (alternative ramp-up)				
	Area	Maximum range	Minimum range	Mean range	
Fleeing (1.5 ms ⁻¹)	219 dB	< 0.01 km ²	< 50 m	< 50 m	< 50 m
	216 dB	< 0.01 km ²	< 50 m	< 50 m	< 50 m
	210 dB	< 0.01 km ²	< 50 m	< 50 m	< 50 m
	207 dB	< 0.01 km ²	< 50 m	< 50 m	< 50 m
	203 dB	< 0.01 km ²	< 50 m	< 50 m	< 50 m
	186 dB	690 km ²	24 km	620 m	12 km
Stationary Stationary (0 m/s ms ⁻¹)	219 dB	1.1 0.66 km ²	600 470 m	600 440 m	600 460 m
	216 dB	2.5 1.7 km ²	900 770 m	900 670 m	900 750 m
	210 dB	14 9.7 km ²	2.2 2.0 km	2.1 1.2 km	2.1 1.8 km
	207 dB	33 22 km ²	3.4 3.2 km	3.2 1.6 km	3.3 2.7 km
	203 dB	96 64 km ²	5.9 5.5 km	5.2 2.3 km	5.6 4.6 km

186 dB ~~2,800~~2,100 km² ~~44~~37 km ~~147.3~~ km ~~27~~26 km
4.3.2 7 m monopile foundation

4.1.4.3 South west location (WT53)

Table 4-21: Weighted $L_{E,p,24h,wt}$ impact areas and ranges for marine mammals using the Southall et al. (2019) impulsive criteria for the WTG foundation installation (alternative ramp-up) scenario at the South west (WT53) modelling location assuming fleeing receptors.

Southall et al. (2019) Weighted $L_{E,p,24h,wt}$		WTG foundation (alternative ramp-up)			
		Area	Maximum range	Minimum range	Mean range
PTS (Impulsive)	LF (183 dB)	0.08 km ²	330 m	60 m	140 m
	HF (185 dB)	< 0.01 km ²	< 50 m	< 50 m	< 50 m
	VHF (155 dB)	< 0.01 km ²	< 50 m	< 50 m	< 50 m
	PCW (185 dB)	< 0.01 km ²	< 50 m	< 50 m	< 50 m
TTS (Impulsive)	LF (168 dB)	2,900 km ²	50 km	1.0 km	24 km
	HF (170 dB)	< 0.01 km ²	< 50 m	< 50 m	< 50 m
	VHF (140 dB)	1,100 km ²	29 km	340 m	16 km
	PCW (170 dB)	180 km ²	12 km	60 m	6.5 km

Table 4-22: Unweighted $L_{E,p,24h}$ impact areas and ranges for fish using the Popper et al. (2014) pile driving criteria for the WTG foundation installation (alternative ramp-up) scenario at the South west (WT53) modelling location assuming both fleeing and stationary receptors.

Popper et al. (2014) Unweighted $L_{E,p,24h}$		WTG foundation (alternative ramp-up)			
		Area	Maximum range	Minimum range	Mean range
Fleeing (1.5 ms ⁻¹)	219 dB	< 0.01 km ²	< 50 m	< 50 m	< 50 m
	216 dB	< 0.01 km ²	< 50 m	< 50 m	< 50 m
	210 dB	< 0.01 km ²	< 50 m	< 50 m	< 50 m
	207 dB	< 0.01 km ²	< 50 m	< 50 m	< 50 m
	203 dB	< 0.01 km ²	< 50 m	< 50 m	< 50 m
Stationary (0 ms ⁻¹)	186 dB	560 km ²	21 km	180 m	12 km
	219 dB	0.53 km ²	430 m	390 m	410 m
	216 dB	1.4 km ²	710 m	610 m	650 m
	210 dB	9.5 km ²	1.9 km	1.5 km	1.7 km
	207 dB	23 km ²	2.9 km	2.0 km	2.6 km
	203 dB	68 km ²	5.1 km	2.7 km	4.6 km
	186 dB	1,900 km ²	34 km	5.5 km	25 km

4.1.5 Mitigation

4.1.5.1 North location (WT03)

Table 4-18 Summary of the 23: Unweighted $SPL_{L_{p,pk}}$ impact areas and ranges for marine mammals using the Southall et al. (2019) impulsive criteria for the **7 m monopile** maximum expected hammer blow energy during the mitigated WTG foundation **modelling** installation scenario at the **C-WTG** North (WT03) modelling location.

Southall et al. (2019) Unweighted $SPL_{L_{p,pk}}$		Area WTG foundation (maximum range Minimum range Mean range energy), mitigated			
		Area	Maximum range	Minimum range	Mean range
PTS (Impulsive)	LF (219 dB)	< 0.01 km ²	< 50 m	< 50 m	< 50 m
	HF (230 dB)	< 0.01 km ²	< 50 m	< 50 m	< 50 m
	VHF (202 dB)	4.5 0.3 km ²	690 330 m	680 320 m	690 320 m
	PCW (218 dB)	< 0.01 km ²	< 50 m	< 50 m	< 50 m

TTS (Impulsive)	LF (213 dB)	0.04 0.02 km ²	120 70 m	120 70 m	120 70 m
	HF (224 dB)	< 0.01 km ²	< 50 m	< 50 m	< 50 m
	VHF (196 dB)	0.4 1.7 km ²	1.8 km 770 m	1.7 km 710 m	1.7 km 740 m
	PCW (212 dB)	0.06 0.02 km ²	140 80 m	140 80 m	140 80 m

Table 4-19 Summary of the 24: Weighted $SEL_{cumLE,p,24h,wtd}$ impact areas and ranges for marine mammals using the Southall et al. (2019) impulsive criteria for the 7-m monopile mitigated WTG foundation modelling installation scenario at the C-WTG North (WT03) modelling location assuming a fleeing animal receptors.

Southall et al. (2019) Weighted $SEL_{cumLE,p,24h,wtd}$		Area WTG foundation, mitigated			
		Area	Maximum range	Minimum range	Mean range
PTS (Impulsive)	LF (183 dB)	400.03 km ²	17 km	3.2 km	9.9 km
	HF (185 dB)	< 0.4 km ²	< 100 m	< 100 m	< 100 m
	VHF (155 dB)	150 < 0.01 km ²	9.1 km < 50 m	3.7 km < 50 m	6.6 km < 50 m
	PCW (185 dB)	0.3 < 0.01 km ²	400 < 50 m	< 100 m	300 < 50 m
TTS (Impulsive)	LF (168 dB)	8,500 1,400 km ²	86 36 km	8.2 1.4 km	43 17 km
	HF (170 dB)	< 0.4 km ²	< 100 m	< 100 m	< 100 m
	VHF (140 dB)	4,500 410 km ²	56 18 km	8.9 1.6 km	33 10 km
	PCW (170 dB)	1,500 43 km ²	32 5.7 km	7.2 km	20 3.3 km

Table 4-20 Summary of the 25: Unweighted $SPL_{RMSLp,RMS}$ impact areas and ranges for marine mammals using the NOAA (2005) Level B impulsive behavioural disturbance criteria for the 7-m monopile maximum expected hammer blow energy during the mitigated WTG foundation modelling installation scenario at the C-WTG North (WT03) modelling location.

NOAA (2005) Unweighted $SPL_{RMSLp,RMS}$		Area WTG foundation (maximum range energy), mitigated			
		Area	Maximum range	Minimum range	Mean range
Level B	160 dB	2,300 390 km ²	37 14 km	14 5.2 km	25 11 km

Table 4-21 Summary of the 26: Unweighted $SPL_{peakLp,pk}$ impact areas and ranges for fish using the Popper et al. (2014) pile driving criteria for the 7-m monopile maximum expected hammer blow energy during the mitigated WTG foundation modelling installation scenario at the C-WTG North (WT03) modelling location.

Popper et al. (2014) Unweighted $SPL_{peakLp,pk}$		Area WTG foundation (maximum range energy), mitigated			
		Area	Maximum range	Minimum range	Mean range
213 dB		0.04 km ²	120 m	120 m	120 m
Pile driving	213 dB	0.02 km ²	70 m	70 m	70 m
207 dB	207 dB	0.30 0.8 km ²	340 160 m	340 160 m	340 160 m

Table 4-22 Summary of the 27: Unweighted $SEL_{cumLE,p,24h}$ impact areas and ranges for marine mammals fish using the Popper et al. (2014) pile driving criteria for the 7-m monopile mitigated WTG foundation modelling installation scenario at the C-WTG North (WT03) modelling location assuming both fleeing and stationary animals receptors.

Popper <i>et al.</i> (2014) Unweighted-SEL _{cum} Unweighted $L_{50,24h}$		Area Maximum range Minimum range Mean range WTC foundation, mitigated			
		Area	Maximum range	Minimum range	Mean range
Fleeing Fleeing (1.5 m/s ms ⁻¹)	219 dB	< 0.4 0.01 km ²	< 100 50 m	< 100 50 m	< 100 50 m
	216 dB	< 0.4 0.01 km ²	< 100 50 m	< 100 50 m	< 100 50 m
	210 dB	< 0.4 0.01 km ²	< 100 50 m	< 100 50 m	< 100 50 m
	207 dB	< 0.4 0.01 km ²	< 100 50 m	< 100 50 m	< 100 50 m
	203 dB	< 0.4 0.01 km ²	< 100 50 m	< 100 50 m	< 100 50 m
Stationary Stationary (0 m/s ms ⁻¹)	186 dB	1,400 180 km ²	3 112 km	6.9 km 930 m	19 6.8 km
	219 dB	1.1 0.2 km ²	600 250 m	600 250 m	600 250 m
	216 dB	2.5 0.5 km ²	900 400 m	900 390 m	900 400 m
	210 dB	14 3.5 km ²	2.2 1.1 km	2.1 km 980 m	2.1 1.1 km
	207 dB	33 8.5 km ²	3.3 1.8 km	3.2 1.5 km	3.2 1.7 km
	203 dB	95 25 km ²	5.8 3.1 km	5.1 2.4 km	5.5 2.9 km
	186 dB	2,700 800 km ²	4 122 km	14 6.4 km	27 17 km

4.1.5.2 **4.4 SW-WTG Central location (WT28)**

4.4.1 11 m monopile foundation

Table 4-23 Summary of the 28: Unweighted $SPL_{peak,L,p,k}$ impact areas and ranges for marine mammals using the Southall et al. (2019) impulsive criteria for the 11-m monopile maximum expected hammer blow energy during the mitigated WTG foundation modelling at the SW-WTG installation scenario at the Central (WT28) modelling location.

Southall et al. (2019) Unweighted $SPL_{peak,L,p,k}$		Area WTG foundation (maximum range Minimum range Mean range energy), mitigated			
		Area	Maximum range	Minimum range	Mean range
PTS (Impulsive)	LF (219 dB)	< 0.01 km ²	< 50 m	< 50 m	< 50 m
	HF (230 dB)	< 0.01 km ²	< 50 m	< 50 m	< 50 m
	VHF (202 dB)	1.6-0.4 km ²	750-380 m	690-350 m	720-360 m
	PCW (218 dB)	< 0.01 km ²	60-50 m	60-50 m	60-50 m
TTS (Impulsive)	LF (213 dB)	0.05-0.02 km ²	130-80 m	130-80 m	130-80 m
	HF (224 dB)	< 0.01 km ²	< 50 m	< 50 m	< 50 m
	VHF (196 dB)	102.1 km ²	1.9-1.9 km	1.7-1.7 km	1.8-1.8 km
	PCW (212 dB)	0.07-0.02 km ²	150-90 m	150-90 m	150-90 m

Table 4-24 Summary of the 29: Weighted $SEL_{cum,L,E,p,24h,wt,d}$ impact areas and ranges for marine mammals using the Southall et al. (2019) impulsive criteria for the 11-m monopile mitigated WTG foundation modelling at the SW-WTG installation scenario at the Central (WT28) modelling location assuming a fleeing animal receptors.

Southall et al. (2019) Weighted $SEL_{cum,L,E,p,24h,wt,d}$		Area Maximum range Minimum range Mean range WTG foundation, mitigated			
		Area	Maximum range	Minimum range	Mean range
PTS (Impulsive)	LF (183 dB)	4700.42 km ²	19-19 km	3.7-3.7 km	11-11 km
	HF (185 dB)	< 0.1-0.01 km ²	< 100-50 m	< 100-50 m	< 100-50 m
	VHF (155 dB)	180-0.01 km ²	10-10 km	4.1-4.1 km	7.3-7.3 km
	PCW (185 dB)	0.2-0.01 km ²	400-50 m	< 100-50 m	200-50 m
TTS (Impulsive)	LF (168 dB)	9,200-2,300 km ²	98-43 km	9.0-1.8 km	45-21 km
	HF (170 dB)	< 0.1-0.01 km ²	< 100-50 m	< 100-50 m	< 100-50 m
	VHF (140 dB)	5,500-680 km ²	63-23 km	9.6-1.9 km	3.6-12 km
	PCW (170 dB)	1,800-84 km ²	358.2 km	7.9-100 m	224.1 km

Table 4-25 Summary of the 30: Unweighted $SPL_{RMS,L,p,RMS}$ impact areas and ranges for marine mammals using the NOAA (2005) Level B impulsive behavioural disturbance criteria for the 11-m monopile maximum expected hammer blow energy during the mitigated WTG foundation modelling at the SW-WTG installation scenario at the Central (WT28) modelling location.

NOAA (2005) Unweighted $SPL_{RMS,L,p,RMS}$		Area WTG foundation (maximum range Minimum range Mean range)			
		Area	Maximum range	Minimum range	Mean range

		range energy), mitigated			
		Area	Maximum range	Minimum range	Mean range
Level B	160 dB	2,600,540 km ²	3917 km	145.1 km	2712 km

Table 4-26 Summary of the 31: Unweighted $SPL_{peak,L,p,k}$ impact areas and ranges for fish using the Popper et al. (2014) pile driving criteria for the 11-m monopile maximum expected hammer blow energy during the mitigated WTG foundation modelling at the SW-WTG installation scenario at the Central (WT28) modelling location.

Popper et al. (2014) Unweighted $SPL_{peak,L,p,k}$		Area WTG foundation (maximum range Minimum range Mean range energy), mitigated			
	213 dB	0.05 km ² Area	130 m Maximum range	130 m Minimum range	130 m Mean range
Pile driving	213 dB	0.02 km ²	80 m	80 m	80 m
	207 dB	0.340.1 km ²	340180 m	320170 m	330180 m

Table 4-27 Summary of the 32: Unweighted $SEL_{cumLE,p,24h}$ impact areas and ranges for marine mammals/fish using the Popper et al. (2014) pile driving criteria for the 11-m monopile mitigated WTG foundation modelling at the SW WTG installation scenario at the Central (WT28) modelling location assuming both fleeing and stationary animals/receptors.

Popper et al. (2014) Unweighted SEL_{cum} Unweighted $L_{E,p,24h}$		Area Maximum range Minimum range Mean range WTG foundation, mitigated			
		Area	Maximum range	Minimum range	Mean range
Fleeing Fleeing (1.5 m/s ms ⁻¹)	219 dB	< 0.4 km ²	< 400 m	< 400 m	< 400 m
	216 dB	< 0.4 km ²	< 400 m	< 400 m	< 400 m
	210 dB	< 0.4 km ²	< 400 m	< 400 m	< 400 m
	207 dB	< 0.4 km ²	< 400 m	< 400 m	< 400 m
	203 dB	< 0.4 km ²	< 400 m	< 400 m	< 400 m
	186 dB	1,900 km ²	361 km	7.6 km	22.9 km
Stationary (0m/s ms ⁻¹)	219 dB	2.0 km ²	800 m	700 m	800 m
	216 dB	4.7 km ²	1.3 km	1.1 km	1.2 km
	210 dB	263.8 km ²	3.1 km	2.6 km	2.9 km
	207 dB	618.7 km ²	4.7 km	3.8 km	4.4 km
	203 dB	17026 km ²	7.9 km	6.3 km	7.4 km
	186 dB	4,000 km ²	5027 km	126.7 km	3319 km

4.4.2 7 m monopile foundation

4.1.5.3 South west location (WT53)

Table 4-28 Summary of the 33: Unweighted $SPL_{peakLp,pk}$ impact areas and ranges for marine mammals using the Southall et al. (2019) impulsive criteria for the 7-m monopile maximum expected hammer blow energy during the mitigated WTG foundation modelling at the SW WTG installation scenario at the South west (WT53) modelling location.

Southall et al. (2019) Unweighted SPL_{peak} $L_{p,pk}$		Area WTG foundation (maximum range Minimum range Mean range energy), mitigated			
		Area	Maximum range	Minimum range	Mean range
PTS (Impulsive)	LF (219 dB)	< 0.01 km ²	< 50 m	< 50 m	< 50 m
	HF (230 dB)	< 0.01 km ²	< 50 m	< 50 m	< 50 m
	VHF (202 dB)	1.6 km ²	740 m	680 m	720 m
	PCW (218 dB)	< 0.01 km ²	< 50 m	< 50 m	< 50 m
TTS (Impulsive)	LF (213 dB)	0.05 km ²	130 m	120 m	130 m
	HF (224 dB)	< 0.01 km ²	< 50 m	< 50 m	< 50 m
	VHF (196 dB)	101.5 km ²	1.9 km	1.7 km	1.8 km
	PCW (212 dB)	0.07 km ²	150 m	150 m	150 m

Table 4-29 Summary of the 34: Weighted $SEL_{cumLE,p,24h,wt}$ impact areas and ranges for marine mammals using the Southall et al. (2019) impulsive criteria for the 7-m monopile mitigated WTG foundation modelling at the SW WTG installation scenario at the South west (WT53) modelling location assuming a fleeing animal/receptors.

Southall et al. (2019) Weighted SEL_{cum} Weighted $L_{E,p,24h,wt}$		Area Maximum range
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		Minimum range Mean			
		Area	Maximum range	Minimum range	Mean range
PTS (Impulsive)	LF (183 dB)	460.06 km ²	18 km	3.6 km	11 km
	HF (185 dB)	< 0.1 km ²	< 100 m	< 100 m	< 100 m
	VHF (155 dB)	180 < 0.01 km ²	10 km < 50 m	4.1 km < 50 m	7.3 km < 50 m
	PCW (185 dB)	0.1 < 0.01 km ²	380 < 50 m	< 100 m	150 < 50 m
TTS (Impulsive)	LF (168 dB)	9,200,2,000 km ²	9840 km	9.0 km	4521 km
	HF (170 dB)	< 0.1 km ²	< 100 m	< 100 m	< 100 m
	VHF (140 dB)	5,500,710 km ²	6322 km	9.6 km	3613 km
	PCW (170 dB)	1,800,95 km ²	357.9 km	7.9 km	245.1 km

Table 4-30 Summary of the 35: Unweighted $SPL_{RMS,Lp,RMS}$ impact areas and ranges for marine mammals using the NOAA (2005) Level B impulsive behavioural disturbance criteria for the 7-m monopile maximum expected hammer blow energy during the mitigated WTG foundation modelling at the SW WTG installation scenario at the South west (WT53) modelling location.

NOAA (2005) Unweighted $SPL_{RMS,Lp,RMS}$		Area WTG foundation (maximum range Minimum range Mean range energy), mitigated			
		Area	Maximum range	Minimum range	Mean range
Level B	160 dB	2,600,530 km ²	3916 km	114.2 km	2713 km

Table 4-31 Summary of the 36: Unweighted $SPL_{peak,Lp,pk}$ impact areas and ranges for fish using the Popper et al. (2014) pile driving criteria for the 7-m monopile maximum expected hammer blow energy during the mitigated WTG foundation modelling at the SW WTG installation scenario at the South west (WT53) modelling location.

Popper et al. (2014) Unweighted $SPL_{peak,Lp,pk}$		Area WTG foundation (maximum range Minimum range Mean range energy), mitigated			
		Area	Maximum range	Minimum range	Mean range
213 dB		0.05 km ²	130 m	120 m	130 m
Pile driving	213 dB	0.01 km ²	70 m	60 m	70 m
207 dB	207 dB	0.33, 0.07 km ²	330, 160 m	320, 150 m	330, 150 m

Table 4-32 Summary of the 37: Unweighted $SEL_{cum,Lp,p,24h}$ impact areas and ranges for marine mammals using the Popper et al. (2014) pile driving criteria for the 7-m monopile mitigated WTG foundation modelling at the SW WTG installation scenario at the South west (WT53) modelling location assuming both fleeing and stationary animals receptors.

Popper et al. (2014) Unweighted $SEL_{cum,Lp,p,24h}$		Area Maximum range Minimum range Mean			
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		range WTG foundation, mitigated			
		Area	Maximum range	Minimum range	Mean range
Fleeing Fleeing (1.5 m/s ms ⁻¹)	219 dB	< 0.1 km ²	< 100 m	< 100 m	< 100 m
	216 dB	< 0.1 km ²	< 100 m	< 100 m	< 100 m
	210 dB	< 0.1 km ²	< 100 m	< 100 m	< 100 m
	207 dB	< 0.1 km ²	< 100 m	< 100 m	< 100 m
	203 dB	< 0.1 km ²	< 100 m	< 100 m	< 100 m
Stationary Stationary (0 m/s ms ⁻¹)	186 dB	1,800 km ²	35 km	7.5 km	229.1 km
	219 dB	1.7 km ²	780 m	700 m	740 m
	216 dB	4.3 km ²	1.2 km	1.1 km	1.2 km
	210 dB	253.2 km ²	3.0 km	2.5 km	2.8 km
	207 dB	598.0 km ²	4.6 km	3.8 km	4.3 km
	203 dB	17026 km ²	7.8 km	6.2 km	7.4 km
	186 dB	3,900 km ²	50 km	44.5 km	32 km

4.2 OSP foundations

4.2.1 4.5 North OSP location

4.5.1 14 m diameter pile foundation

Table 4-33 Summary of the 38: Unweighted $SPL_{peak,Lp,pk}$ impact areas and ranges for marine mammals using the Southall et al. (2019) impulsive criteria for the 14 m monopile maximum expected hammer blow energy during the OSP foundation modelling installation scenario at the North OSP modelling location.

Southall et al. (2019) Unweighted $SPL_{peak,Lp,pk}$		Area OSP foundation maximum range Minimum range Mean range (energy)			
		Area	Maximum range	Minimum range	Mean range
PTS (Impulsive)	LF (219 dB)	< 0.01 km ²	< 50 m	< 50 m	< 50 m
	HF (230 dB)	< 0.01 km ²	< 50 m	< 50 m	< 50 m
	VHF (202 dB)	1.0 km ²	580 m	570 m	570 m
	PCW (218 dB)	< 0.01 km ²	< 50 m	< 50 m	< 50 m
TTS (Impulsive)	LF (213 dB)	0.03 km ²	110 m	100 m	110 m
	HF (224 dB)	< 0.01 km ²	< 50 m	< 50 m	< 50 m
	VHF (196 dB)	5.9 km ²	1.4 km	1.3 km	1.4 km
	PCW (212 dB)	0.05 km ²	120 m	120 m	120 m

Table 4-34 Summary of the 39: Weighted $SEL_{cum}L_{E,p,24h,wt}$ impact areas and ranges for marine mammals using the Southall et al. (2019) impulsive criteria for the 14-m monopile OSP foundation modelling installation scenario at the North OSP modelling location assuming a fleeing animal receptors.

Southall et al. (2019) Weighted SEL_{cum} Weighted $L_{E,p,24h,wt}$		Area Maximum range Minimum range Mean range range OSP foundation			
		Area	Maximum range	Minimum range	Mean range
PTS (Impulsive)	LF (183 dB)	1007.1 km ²	9.53.0 km	700100 m	4.81.1 km
	HF (185 dB)	< 0.40.01 km ²	< 40050 m	< 40050 m	< 40050 m
	VHF (155 dB)	490.23 km ²	5.4 km520 m	1.5 km< 50 m	3.7 km200 m
	PCW (185 dB)	< 0.40.01 km ²	< 40050 m	< 40050 m	< 40050 m
TTS (Impulsive)	LF (168 dB)	5,4002,500 km ²	6749 km	4.02.2 km	3322 km
	HF (170 dB)	< 0.40.01 km ²	< 40050 m	< 40050 m	< 40050 m
	VHF (140 dB)	2,500900 km ²	4529 km	4.82.3 km	2414 km
	PCW (170 dB)	570160 km ²	2112 km	3.4 km850 m	126.4 km

Table 4-35 Summary of the 40: Unweighted $SPL_{RMS}L_{p,RMS}$ impact areas and ranges for marine mammals using the NOAA (2005) Level B impulsive behavioural disturbance criteria for the 14-m monopile maximum expected hammer blow energy during the OSP foundation modelling installation scenario at the North OSP modelling location.

NOAA (2005) Unweighted $SPL_{RMS}L_{p,RMS}$		Area OSP foundation maximum range Minimum range Mean range range energy)			
		Area	Maximum range	Minimum range	Mean range
Level B	160 dB	1,200650 km ²	2820 km	6.35.8 km	1814 km

Table 4-36 Summary of the 41: Unweighted $SPL_{peak}L_{p,pk}$ impact areas and ranges for fish using the Popper et al. (2014) pile driving criteria for the 14-m monopile maximum expected hammer blow energy during the OSP foundation modelling installation scenario at the North OSP modelling location.

Popper et al. (2014) Unweighted $SPL_{peak}L_{p,pk}$		Area OSP foundation maximum range Minimum range Mean range range energy)			
		Area	Maximum range	Minimum range	Mean range
213 dB		0.03 km ² Area	140 m Maximum range	100 m Minimum range	110 m Mean range
Pile driving	213 dB	0.04 km ²	110 m	110 m	110 m
207 dB	207 dB	0.220.21 km ²	270260 m	270250 m	270260 m

Table 4-37 Summary of the 42: Unweighted $SEL_{cum}L_{E,p,24h}$ impact areas and ranges for marine mammals fish using the Popper et al. (2014) pile driving criteria for the 14-m monopile OSP foundation modelling installation scenario at the North OSP modelling location assuming both fleeing and stationary animals receptors.

Popper et al. (2014)		Area Maximum			
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Unweighted-SEL _{cum} Unweighted <i>L_{0,5m}</i>	range Minimum range Mean range				
	Area	Maximum range	Minimum range	Mean range	
Fleeing Fleeing (1.5 m/s ms ⁻¹)	219 dB	< 0.40.01 km ²	< 40050 m	< 40050 m	< 40050 m
	216 dB	< 0.40.01 km ²	< 40050 m	< 40050 m	< 40050 m
	210 dB	< 0.40.01 km ²	< 40050 m	< 40050 m	< 40050 m
	207 dB	< 0.40.01 km ²	< 40050 m	< 40050 m	< 40050 m
	203 dB	< 0.40.01 km ²	< 40050 m	< 40050 m	< 40050 m
	186 dB	600450 km ²	2220 km	3.31.7 km	4210 km
Stationary Stationary (0 m/s ms ⁻¹)	219 dB	0.80.62 km ²	500450 m	500430 m	500440 m
	216 dB	1.81.6 km ²	800720 m	700680 m	800700 m
	210 dB	0.39.9 km ²	1.81.9 km	1.6 km	1.71.8 km
	207 dB	2022 km ²	2.82.9 km	2.42.1 km	2.62.7 km
	203 dB	5659 km ²	4.74.9 km	3.72.8 km	4.24.4 km
	186 dB	1,5001,300 km ²	3130 km	6.36.9 km	2022 km

4.5.2 7 m diameter pile foundation

4.2.2 South OSP location

Table 4-38 ~~Summary of the 43~~: Unweighted $SPL_{peak,L,p,k}$ impact areas and ranges for marine mammals using the Southall et al. (2019) impulsive criteria for the 7-m monopile maximum expected hammer blow energy during the OSP foundation modelling installation scenario at the NorthSouth OSP modelling location.

Southall et al. (2019) Unweighted $SPL_{peak,L,p,k}$		Area OSP foundation maximum range Minimum range Mean range (energy)			
		Area	Maximum range	Minimum range	Mean range
PTS (Impulsive)	LF (219 dB)	< 0.01 km ²	< 50 m	< 50 m	< 50 m
	HF (230 dB)	< 0.01 km ²	< 50 m	< 50 m	< 50 m
	VHF (202 dB)	1.0-0.4 km ²	570-380 m	560-310 m	570-340 m
	PCW (218 dB)	< 0.01 km ²	< 50 m	< 50 m	< 50 m
TTS (Impulsive)	LF (213 dB)	0.03-0.02 km ²	110-80 m	100-80 m	100-80 m
	HF (224 dB)	< 0.01 km ²	< 50 m	< 50 m	< 50 m
	VHF (196 dB)	5.8-1.7 km ²	1.4 km-880 m	1.3 km-620 m	1.4 km-730 m
	PCW (212 dB)	0.05-0.03 km ²	120-100 m	120-90 m	120-90 m

Table 4-39 ~~Summary of the 44~~: Weighted $SEL_{cum,L,E,p,24h,wtd}$ impact areas and ranges for marine mammals using the Southall et al. (2019) impulsive criteria for the 7-m monopile OSP foundation modelling installation scenario at the NorthSouth OSP modelling location assuming a fleeing animal receptors.

Southall et al. (2019) Weighted $SEL_{cum,L,E,p,24h,wtd}$		Area Maximum range Minimum range Mean range OSP foundation			
		Area	Maximum range	Minimum range	Mean range
PTS (Impulsive)	LF (183 dB)	100-0.28 km ²	9.4 km-790 m	700-70 m	4.7 km-240 m
	HF (185 dB)	< 0.4-0.01 km ²	< 100-50 m	< 100-50 m	< 100-50 m
	VHF (155 dB)	49-0.09 km ²	5.4 km-370 m	1.5 km-< 50 m	3.7 km-130 m
	PCW (185 dB)	< 0.4-0.01 km ²	< 100-50 m	< 100-50 m	< 100-50 m
TTS (Impulsive)	LF (168 dB)	5,300-1,900 km ²	67-41 km	4.0 km-590 m	33-19 km
	HF (170 dB)	< 0.4-0.01 km ²	< 100-50 m	< 100-50 m	< 100-50 m
	VHF (140 dB)	2,500-1,000 km ²	45-28 km	4.8 km-250 m	24-15 km
	PCW (170 dB)	570-180 km ²	24-11 km	3.4 km-60 m	12-6.7 km

Table 4-40 ~~Summary of the 45~~: Unweighted $SPL_{RMS,L,p,RMS}$ impact areas and ranges for marine mammals using the NOAA (2005) Level B impulsive behavioural disturbance criteria for the 7-m monopile maximum expected hammer blow energy during the OSP foundation modelling installation scenario at the NorthSouth OSP modelling location.

NOAA (2005) Unweighted $SPL_{RMS,L,p,RMS}$		Area OSP foundation maximum range Minimum range Mean range (energy)			
		Area	Maximum range	Minimum range	Mean range
Level B	160 dB	1,200-560 km ²	28-17 km	6.3-1.4 km	18-13 km

Table 4-41 ~~Summary of the 46~~: Unweighted $SPL_{peak,Lp,pk}$ impact areas and ranges for fish using the Popper et al. (2014) pile driving criteria for the ~~7-m monopile~~ maximum expected hammer blow energy during the OSP foundation ~~modelling~~ installation scenario at the ~~North~~South OSP modelling location.

Popper et al. (2014) Unweighted $SPL_{peak,Lp,pk}$		Area OSP foundation (maximum range Minimum range Mean range energy)			
213-dB		0.03 km² Area	110 mMaximum range	100 m Minimum range	100 m Mean range
Pile driving	213 dB	0.02 km ²	80 m	80 m	80 m
207-dB	207 dB	0.22 0.1 km ²	270 190 m	260 170 m	260 180 m

Table 4-47: Unweighted $L_{E,p,24h}$ impact areas and ranges for fish using the Popper et al. (2014) pile driving criteria for the OSP foundation installation scenario at the South OSP modelling location assuming both fleeing and stationary receptors.

	Popper et al. (2014) Unweighted $L_{E,p,24h}$	OSP foundation			
		Area	Maximum range	Minimum range	Mean range
Fleeing (1.5 ms ⁻¹)	219 dB	< 0.01 km ²	< 50 m	< 50 m	< 50 m
	216 dB	< 0.01 km ²	< 50 m	< 50 m	< 50 m
	210 dB	< 0.01 km ²	< 50 m	< 50 m	< 50 m
	207 dB	< 0.01 km ²	< 50 m	< 50 m	< 50 m
	203 dB	< 0.01 km ²	< 50 m	< 50 m	< 50 m
Stationary (0 ms ⁻¹)	186 dB	350 km ²	16 km	120 m	9.3 km
	219 dB	0.33 km ²	350 m	300 m	320 m
	216 dB	0.81 km ²	570 m	450 m	500 m
	210 dB	5.3 km ²	1.5 km	980 m	1.3 km
	207 dB	13 km ²	2.4 km	1.3 km	2.0 km
	203 dB	37 km ²	4.1 km	1.6 km	3.4 km
	186 dB	1,100 km ²	26 km	5.0 km	20 km

4.2.3 Alternative ramp-up

4.2.3.1 North OSP location

Table 4-48: Weighted $L_{E,p,24h,wt}$ impact areas and ranges for marine mammals using the Southall et al (2019) impulsive criteria for the OSP foundation installation (alternative ramp-up) scenario at the North OSP modelling location assuming fleeing receptors.

	Southall et al. (2019) Weighted $L_{E,p,24h,wt}$	OSP foundation (alternative ramp-up)			
		Area	Maximum range	Minimum range	Mean range
PTS (Impulsive)	LF (183 dB)	0.02 km ²	120 m	60 m	80 m
	HF (185 dB)	< 0.01 km ²	< 50 m	< 50 m	< 50 m
	VHF (155 dB)	< 0.01 km ²	< 50 m	< 50 m	< 50 m
	PCW (185 dB)	< 0.01 km ²	< 50 m	< 50 m	< 50 m
TTS (Impulsive)	LF (168 dB)	1,600 km ²	41 km	1.2 km	17 km
	HF (170 dB)	< 0.01 km ²	< 50 m	< 50 m	< 50 m
	VHF (140 dB)	570 km ²	24 km	520 m	11 km
	PCW (170 dB)	58 km ²	7.9 km	60 m	3.3 km

Table 4-42 Summary of the 49: Unweighted $SEL_{cum}L_{E,p,24h}$ impact areas and ranges for marine mammals using the Popper et al. (2014) pile driving criteria for the 7-m monopile OSP foundation modelling installation (alternative ramp-up) scenario at the North OSP modelling location assuming both fleeing and stationary animals receptors.

	Popper et al. (2014) Unweighted SEL_{cum} Unweighted $L_{E,p,24h}$	OSP foundation (alternative ramp-up)			
		Area	Maximum range	Minimum range	Mean range
Fleeing Fleeing (1.5 m/s ms ⁻¹)	219 dB	< 0.4 0.01 km ²	< 400 50 m	< 400 50 m	< 400 50 m
	216 dB	< 0.4 0.01 km ²	< 400 50 m	< 400 50 m	< 400 50 m
	210 dB	< 0.4 0.01 km ²	< 400 50 m	< 400 50 m	< 400 50 m
	207 dB	< 0.4 0.01 km ²	< 400 50 m	< 400 50 m	< 400 50 m
	203 dB	< 0.4 0.01 km ²	< 400 50 m	< 400 50 m	< 400 50 m
	186 dB	590 250 km ²	21 16 km	3.3 km 270 m	12 7.2 km
Stationary	219 dB	0.80 52 km ²	500 420 m	500 400 m	500 410 m
	216 dB	1.7 1.3 km ²	800 660 m	700 630 m	700 650 m

Stationary (0 m/s ms ⁻¹)	210 dB	8.9 8.5 km ²	1.8 1.7 km	1.6 1.5 km	1.7 1.6 km
	207 dB	20 19 km ²	2.7 km ² km	2.4 2.0 km	2.5 km
	203 dB	54 52 km ²	4.6 4.5 km	3.7 2.7 km	4.2 4.1 km
	186 dB	1,500 1,200 km ²	31 29 km	6.3 6.8 km	20 21 km

4.2.3.2 ~~4.6~~ South OSP location

Table 4-50: Weighted $L_{E,p,24h,wtd}$ impact areas and ranges for marine mammals using the Southall et al (2019) impulsive criteria for the OSP foundation installation (alternative ramp-up) scenario at the South OSP modelling location assuming fleeing receptors.

Southall et al. (2019) Weighted $L_{E,p,24h,wtd}$		OSP foundation (alternative ramp-up)			
		Area	Maximum range	Minimum range	Mean range
PTS (Impulsive)	LF (183 dB)	< 0.01 km ²	70 m	50 m	60 m
	HF (185 dB)	< 0.01 km ²	< 50 m	< 50 m	< 50 m
	VHF (155 dB)	< 0.01 km ²	< 50 m	< 50 m	< 50 m
	PCW (185 dB)	< 0.01 km ²	< 50 m	< 50 m	< 50 m
TTS (Impulsive)	LF (168 dB)	1,100 km ²	33 km	430 m	14 km
	HF (170 dB)	< 0.01 km ²	< 50 m	< 50 m	< 50 m
	VHF (140 dB)	670 km ²	24 km	130 m	12 km
	PCW (170 dB)	61 km ²	7.4 km	< 50 m	3.5 km

4.6.1 14 m diameter pile foundation

Table 4-51: Unweighted $L_{E,p,24h}$ impact areas and ranges for marine mammals using the Popper et al. (2014) pile driving criteria for the OSP foundation installation (alternative ramp-up) scenario at the South OSP modelling location assuming both fleeing and stationary receptors.

Popper et al. (2014) Unweighted $L_{E,p,24h}$		OSP foundation (alternative ramp-up)			
		Area	Maximum range	Minimum range	Mean range
Fleeing (1.5 ms ⁻¹)	219 dB	< 0.01 km ²	< 50 m	< 50 m	< 50 m
	216 dB	< 0.01 km ²	< 50 m	< 50 m	< 50 m
	210 dB	< 0.01 km ²	< 50 m	< 50 m	< 50 m
	207 dB	< 0.01 km ²	< 50 m	< 50 m	< 50 m
	203 dB	< 0.01 km ²	< 50 m	< 50 m	< 50 m
	186 dB	170 km ²	12 km	80 m	6.1 km
Stationary (0 ms ⁻¹)	219 dB	0.27 km ²	320 m	280 m	290 m
	216 dB	0.68 km ²	520 m	420 m	460 m
	210 dB	4.5 km ²	1.4 km	920 m	1.2 km
	207 dB	11 km ²	2.2 km	1.3 km	1.8 km
	203 dB	32 km ²	3.8 km	1.6 km	3.2 km
	186 dB	1,000 km ²	25 km	4.9 km	19 km

4.2.4 Mitigation

4.2.4.1 North OSP location

Table 4-~~43~~ Summary of the 52: Unweighted $SPL_{L_{p,pk}}$ impact areas and ranges for marine mammals using the Southall et al. (2019) impulsive criteria for the 14 m monopile maximum expected hammer blow energy during the mitigated OSP foundation modelling installation scenario at the SouthNorth OSP modelling location.

Southall et al. (2019) Unweighted $SPL_{L_{p,pk}}$		Area OSP foundation (maximum range Minimum range Mean range energy), mitigated			
		Area	Maximum range	Minimum range	Mean range
PTS	LF (219 dB)	< 0.01 km ²	< 50 m	< 50 m	< 50 m

[Impulsive]	HF (230 dB)	< 0.01 km ²	< 50 m	< 50 m	< 50 m
	VHF (202 dB)	1.4 0.27 km ²	680 300 m	660 290 m	670 300 m
	PCW (218 dB)	< 0.01 km ²	60 < 50 m	60 < 50 m	60 < 50 m
TTS [Impulsive]	LF (213 dB)	0.05 0.01 km ²	120 60 m	120 60 m	120 60 m
	HF (224 dB)	< 0.01 km ²	< 50 m	< 50 m	< 50 m
	VHF (196 dB)	8.9 1.3 km ²	1.7 km 680 m	1.6 km 620 m	1.7 km 650 m
	PCW (212 dB)	0.06 0.01 km ²	140 70 m	140 70 m	140 70 m

Table 4-44 Summary of the 53: Weighted $SEL_{cumLE,p,24h,wtd}$ impact areas and ranges for marine mammals using the Southall et al. (2019) impulsive criteria for the 14-m monopile mitigated OSP foundation modelling installation scenario at the SouthNorth OSP modelling location assuming a fleeing animal receptors.

Southall et al. (2019) Weighted SEL_{cum} $L_{E,p,24h,wtd}$		Area Maximum range Minimum range Mean range range OSP foundation, mitigated			
		Area	Maximum range	Minimum range	Mean range
PTS (Impulsive)	LF (183 dB)	4500.02 km ²	18 km	3.4 km	10 km
	HF (185 dB)	< 0.4 km ²	< 100 m	< 100 m	< 100 m
	VHF (155 dB)	170 < 0.01 km ²	10 km < 50 m	4.0 km < 50 m	7.1 km < 50 m
	PCW (185 dB)	< 0.4 km ²	200 < 50 m	< 100 m	100 < 50 m
TTS (Impulsive)	LF (168 dB)	9,100 1,100 km ²	95 33 km	9.3 1.2 km	45 15 km
	HF (170 dB)	< 0.4 km ²	< 100 m	< 100 m	< 100 m
	VHF (140 dB)	5,500 340 km ²	62 17 km	9.9 1.4 km	36 9.1 km
	PCW (170 dB)	1,700 33 km ²	34 5.2 km	7.9 km	120 m

Table 4-45 Summary of the 54: Unweighted $SPL_{RMSLp,RMS}$ impact areas and ranges for marine mammals using the NOAA (2005) Level B impulsive behavioural disturbance criteria for the 14-m monopile maximum expected hammer blow energy during the mitigated OSP foundation modelling installation scenario at the SouthNorth OSP modelling location.

NOAA (2005) Unweighted $SPL_{RMSLp,RMS}$		Area OSP foundation maximum range Minimum range Mean range range energy), mitigated			
		Area	Maximum range	Minimum range	Mean range
Level B	160 dB	2,600 330 km ²	39 13 km	12 4.8 km	27 10 km

Table 4-46 Summary of the 55: Unweighted $SPL_{peakLp,pk}$ impact areas and ranges for fish using the Popper et al. (2014) pile driving criteria for the 14-m monopile maximum expected hammer blow energy during the mitigated OSP foundation modelling installation scenario at the SouthNorth OSP modelling location.

Popper et al. (2014) Unweighted $SPL_{peakLp,pk}$		Area OSP foundation maximum range Minimum range Mean range range energy), mitigated			
		Area	Maximum range	Minimum range	Mean range
213 dB		0.05 km ²	120 m	120 m	120 m
Pile driving	213 dB	0.01 km ²	60 m	60 m	60 m
207 dB	207 dB	0.30 0.7 km ²	310 150 m	310 150 m	310 150 m

Table 4-47 Summary of the 56: Unweighted $SEL_{cumLE,p,24h}$ impact areas and ranges for marine mammals/fish using the Popper et al. (2014) pile driving criteria for the 14-m monopile mitigated OSP foundation modelling installation scenario at the SouthNorth OSP modelling location assuming both fleeing and stationary animals receptors.

Popper et al. (2014)	Area
----------------------	------

Unweighted SEL _{cum} Unweighted <i>L_{0.5}</i>		Maximum range Minimum range Mean <i>range OSP foundation, mitigated</i>			
		Area	Maximum range	Minimum range	Mean range
Fleeing Fleeing (1.5 m/s ⁻¹)	219 dB	< 0.4 0.01 km ²	< 100 50 m	< 100 50 m	< 100 50 m
	216 dB	< 0.4 0.01 km ²	< 100 50 m	< 100 50 m	< 100 50 m
	210 dB	< 0.4 0.01 km ²	< 100 50 m	< 100 50 m	< 100 50 m
	207 dB	< 0.4 0.01 km ²	< 100 50 m	< 100 50 m	< 100 50 m
	203 dB	< 0.4 0.01 km ²	< 100 50 m	< 100 50 m	< 100 50 m
	186 dB	1,800 130 km ²	35 11 km	7.6 km840 m	22 5.8 km
Stationary Stationary (0 m/s ⁻¹)	219 dB	1.7 0.18 km ²	800 240 m	700 230 m	700 240 m
	216 dB	4.0 0.45 km ²	1.2 km390 m	1.1 km370 m	1.1 km380 m
	210 dB	23 2.9 km ²	2.8 km990 m	2.6 km900 m	2.7 km950 m
	207 dB	54 7.4 km ²	4.4 1.6 km	3.8 1.4 km	4.2 1.5 km
	203 dB	160 22 km ²	7.7 2.9 km	6.3 2.1 km	7.4 2.7 km
	186 dB	4,000 680 km ²	50 21 km	12 5.9 km	33 16 km

4.6.2 7 m diameter pile foundation

4.2.4.2 South OSP location

Table 4-48 Summary of the 57: Unweighted $SPL_{peak,L,p,k}$ impact areas and ranges for marine mammals using the Southall et al. (2019) impulsive criteria for the 7-m monopile maximum expected hammer blow energy during the mitigated OSP foundation modelling installation scenario at the South OSP modelling location.

Southall et al. (2019) Unweighted $SPL_{peak,L,p,k}$		Area OSP foundation maximum range Minimum range Mean range energy), mitigated			
		Area	Maximum range	Minimum range	Mean range
PTS (Impulsive)	LF (219 dB)	< 0.01 km ²	< 50 m	< 50 m	< 50 m
	HF (230 dB)	< 0.01 km ²	< 50 m	< 50 m	< 50 m
	VHF (202 dB)	1.40-13 km ²	670-220 m	660-190 m	670-200 m
	PCW (218 dB)	< 0.01 km ²	60 < 50 m	60 < 50 m	60 < 50 m
TTS (Impulsive)	LF (213 dB)	0.04 < 0.01 km ²	120 < 50 m	120 < 50 m	120 < 50 m
	HF (224 dB)	< 0.01 km ²	< 50 m	< 50 m	< 50 m
	VHF (196 dB)	8.80-6 km ²	1.7 km-500 m	1.6 km-390 m	1.7 km-440 m
	PCW (212 dB)	0.06 < 0.01 km ²	140-60 m	140-50 m	140-60 m

Table 4-49 Summary of the 58: Weighted $SEL_{cum,L,E,p,24h,wt}$ impact areas and ranges for marine mammals using the Southall et al. (2019) impulsive criteria for the 7-m monopile mitigated OSP foundation modelling installation scenario at the South OSP modelling location assuming a-fleeing animal receptors.

Southall et al. (2019) Weighted $SEL_{cum,L,E,p,24h,wt}$		Area Maximum range Minimum range Mean range OSP foundation, mitigated			
		Area	Maximum range	Minimum range	Mean range
PTS (Impulsive)	LF (183 dB)	440 < 0.01 km ²	18 km-50 m	3.4 km < 50 m	10 km < 50 m
	HF (185 dB)	< 0.40-0.01 km ²	< 100-50 m	< 100-50 m	< 100-50 m
	VHF (155 dB)	170 < 0.01 km ²	10 km < 50 m	4.0 km < 50 m	7.1 km < 50 m
	PCW (185 dB)	< 0.40-0.01 km ²	200 < 50 m	< 100-50 m	100 < 50 m
TTS (Impulsive)	LF (168 dB)	9,000-740 km ²	94-26 km	9.3 km-340 m	45-12 km
	HF (170 dB)	< 0.40-0.01 km ²	< 100-50 m	< 100-50 m	< 100-50 m
	VHF (140 dB)	5,500-420 km ²	62-18 km	9.9 km-100 m	36-10 km
	PCW (170 dB)	1,700-32 km ²	34-4.9 km	7.9 km < 50 m	21-2.9 km

Table 4-50 Summary of the 59: Unweighted $SPL_{RMS,L,p,RMS}$ impact areas and ranges for marine mammals using the NOAA (2005) Level B impulsive behavioural disturbance criteria for the 7-m monopile maximum expected hammer blow energy during the mitigated OSP foundation modelling installation scenario at the South OSP modelling location.

NOAA (2005) Unweighted $SPL_{RMS,L,p,RMS}$		Area OSP foundation maximum range Minimum range Mean range energy), mitigated			
		Area	Maximum range	Minimum range	Mean range
Level B	160 dB	2,600-280 km ²	38-11 km	12-1.4 km	27-9.1 km

Table 4-51 Summary of the 60: Unweighted $SPL_{peak,Lp,pk}$ impact areas and ranges for fish using the Popper et al. (2014) pile driving criteria for the 7-m monopile maximum expected hammer blow energy during the mitigated OSP foundation modelling installation scenario at the South OSP modelling location.

Popper et al. (2014) Unweighted $SPL_{peak,Lp,pk}$		Area OSP foundation maximum range Minimum range Mean range (energy), mitigated			
213-dB		0.04 km² Area	120 m Maximum range	120 m Minimum range	120 m Mean range
Pile driving	213 dB	< 0.01 km ²	< 50 m	< 50 m	< 50 m
207-dB	207 dB	0.29 0.04 km ²	310 110 m	310 100 m	310 110 m

Table 4-52 Summary of the 61: Unweighted $SEL_{cum,L,E,p,24h}$ impact areas and ranges for marine mammals/fish using the Popper et al. (2014) pile driving criteria for the 7-m monopile mitigated OSP foundation modelling installation scenario at the South OSP modelling location assuming both fleeing and stationary animals/receptors.

Popper et al. (2014) Unweighted- $SEL_{cum,L,E,p,24h}$		Area Maximum range Minimum range Mean range OSP foundation, mitigated			
		Area	Maximum range	Minimum range	Mean range
Fleeing Fleeing (1.5 m/s ms ⁻¹)	219 dB	< 0.4 km ²	< 400 m	< 400 m	< 400 m
	216 dB	< 0.4 km ²	< 400 m	< 400 m	< 400 m
	210 dB	< 0.4 km ²	< 400 m	< 400 m	< 400 m
	207 dB	< 0.4 km ²	< 400 m	< 400 m	< 400 m
	203 dB	< 0.4 km ²	< 400 m	< 400 m	< 400 m
	186 dB	1,800 km ²	358.0 km	7.6 km	24.8 km
Stationary Stationary (0 m/s ms ⁻¹)	219 dB	1.6 km ²	800 m	700 m	700 m
	216 dB	3.9 km ²	1.2 km	1.1 km	1.1 km
	210 dB	23.5 km ²	2.8 km	2.5 km	2.7 km
	207 dB	53.9 km ²	4.3 km	3.8 km	4.1 km
	203 dB	160 km ²	7.6 km	6.2 km	7.0 km
	186 dB	3,900 km ²	49 km	12.4 km	33 km

5 Other noise sources

Although impact piling is expected to generate the greatest overall noise source levels during offshore construction and development (Bailey *et al.*, 2014), several other anthropogenic underwater noise sources may be present. Each of these has been have the potential to be associated with the Proposed Development and need to be considered, and. These noise sources have been presented alongside relevant biological noise criteria presented, (see section 2.3) in this section.

Table 5-1 Summary of the possible noise making activities at ABWP2 other than impact piling provides a summary of The list below shows the various noise producing sources, aside from impact piling, that are expected to be present occur during the construction and operation of ABWP2.

- Cable laying

Table 5-1 Summary of the possible noise making activities at ABWP2 other than impact piling

Activity	Description
Cable laying	Noise from the Cable Laying Vessel and any other associated noise during the offshore cable installation.
Dredging	Dredging may be required on site for seabed preparation work for certain foundation options (e.g., gravity bases), as well as for the Export Cable, Array Cables and Interconnector Cable Installation. Suction dredging has been assumed as a worst case due to the louder processes involved compared to other considered methods such as backhoe dredging.
Drilling	There is the potential for WTG Foundations to be installed using drilling depending on seabed type or if a pile refuses during impact piling operations.
Rock placement	Potentially required on site for installation of offshore cables (Cable Crossings and Cable Protection) and Scour Protection around foundation structures.
Trenching	Plough trenching may be required during Offshore Cable installation.
Trenchless techniques	Trenchless techniques will also be considered as part of construction, which do not involve breaking the surface of the seabed, rather using an under-seabed directional drilling technique. Little directly measured noise data is available for this and its prediction is extremely complex. As there is no significant noisy activity in the water column itself for trenchless techniques, except briefly at the exit point, drilling (above) will be assumed as the worst case scenario.
Vessel activities	Jack-up barges for piling substructure and WTG installation. Other large and medium sized vessels to carry out other construction tasks and anchor handling. Other small vessels for crew transport and maintenance on site.
Operational WTG	Noise transmitted through the water from operational WTGs.
UXO clearance	There is a possibility, albeit highly unlikely, that Unexploded Ordnance (UXO) may exist within the boundaries of APWP2, which would need to be cleared before construction can begin.

- Dredging
- Drilling (including trenchless techniques)
- Rock placement
- Trenching
- Vessel activities
- Operational WTGs
- UXO clearance

The NPL Good Practice Guide 133 for underwater noise measurements (Robinson *et al.*, 2014) indicates that under certain circumstances, a simple modelling approach ~~may be~~ considered ~~acceptable~~ appropriate. Such an approach has been used for these noise sources, which are variously either quiet compared to impact piling (e.g., cable laying and dredging), or where detailed modelling would imply unjustified accuracy (e.g., where data is limited such as with UXO detonation). As such, the ~~high-level overview of~~ modelling that has been presented here is considered sufficient ~~for the purpose of understanding and describing the likely significant effects~~. The limitations of this approach are noted, including the lack of frequency or bathymetric dependence.

~~Most~~The majority of these activities are considered in section 5.1, with operational WTG noise and UXO clearance assessed in sections 5.2 and 5.3 respectively.

5.1 Noise making activities (construction)

For the purposes of identifying the greatest noise levels, approximate subsea noise levels have been predicted using a simple modelling approach based on measurement data from Subacoustech Environmental’s own underwater noise measurement database, scaled to relevant parameters for the site and to the specific noise sources to be used. The calculation of underwater noise transmission loss for the non-impulsive sources is based on an empirical analysis of the noise measurements taken along transects around these sources by Subacoustech Environmental. The predictions use the following principle fitted to the measured data, where R is the range from the source, N is the transmission loss, and α is the absorption loss:

$$\text{Received level} = \text{Source level (SL)} - N \log_{10} R - \alpha R$$

Predicted source levels and propagation calculations for the construction activities are presented in Table 5-2-1 along with a summary of the number of datasets used in each case. As previously, all $SEL_{cumLE,p,t}$ criteria use the same assumptions as presented in section 2-3.2.3, and ranges smaller than 50 m (~~single strike~~) and 100 m (~~cumulative~~) have not been presented. It should be reiterated that this modelling approach does not take bathymetry or any other environmental conditions into account, and as such can be applied to any location at, or surrounding, ABWP2.

Table 5-2-1: Summary of the estimated unweighted source levels and transmission losses for the ~~different~~ considered ~~noise sources related to~~ construction activities, based on directly measured data

Source activity	Estimated unweighted source level @ 1 m	Transmission loss parameters coefficients	Comments
Cable laying	171 dB re 1 μ Pa @ 1 m (RMS)	N 13, α 0.0 (no absorption)	Based on 11 datasets from a pipe laying vessel measuring 300 m in length; this is considered a worst case precautionary noise source for cable laying operations.
Dredging (backhoe)	165 dB re 1 μ Pa @ 1 m (RMS)	N 19, α 0.0009	Based on three datasets from backhoe dredgers.
Dredging (suction)	186 dB re 1 μ Pa @ 1 m (RMS)	N 19, α 0.0009	Based on five datasets from suction and cutter suction cutter-suction dredgers.
Drilling (including trenchless techniques)	169 dB re 1 μ Pa @ 1 m (RMS)	N 16, α 0.0006	Based on six datasets from various drilling operations covering ground investigations and pile installation. A 200kW 200 kW drill has been assumed for modelling.
Rock placement	172 166 dB re 1 μ Pa @ 1 m (RMS)	N 429, α 0.00050.0025	Based on four datasets from rock placement vessel ‘Rollingstone.’

Trenching	172 dB re 1 μ Pa @ 1 m (RMS)	N 13, α 0.0004	Based on three datasets of measurements from trenching vessels more than 100 m in length.
Vessel noise (large)	168 dB re 1 μ Pa @ 1 m (RMS)	N 12, α 0.0021	Based on five datasets of large vessels including container ships, Floating Production Storage and Offloading (FPSOs) and other vessels more than 100 m in length. Vessel speed assumed as 10 knots kn.
Vessel noise (medium)	161 dB re 1 μ Pa @ 1 m (RMS)	N 12, α 0.0021	Based on three datasets of moderate sized vessels less than below 100 m in length. Vessel speed assumed as 10 knots kn.

All values of N and α are empirically derived and will be linked to the size and shape of the machinery ~~and the noise source on it~~, the transect on which the measurements ~~are were~~ taken and the local environment at the time. It is noted that the depths at ABWP2 are deep relative to the locations where the original data here was derived, although the noise levels relative to the thresholds under consideration will mean that the relatively low impact ranges predicted are unlikely to be significantly affected.

For $SEL_{cumLE,p,t}$ calculations in this section, the duration the noise is present also needs to be considered, with all sources assumed to operate constantly for 24 hours to give a ~~worst case~~precautionary assessment of ~~the~~ noise. Due to the relatively low ~~noise~~ level of ~~the noise from these sources~~ ~~considered~~, both ~~fleeing~~moving and stationary ~~animals~~receptors have been included for all $SEL_{cumLE,p,t}$ criteria; ~~the same swim speeds as presented in section 2.3 have been assumed here~~.

To account for the weightings required for modelling using the Southall *et al.* (2019) $L_{E,p,t}$ criteria (see section 2.3.12.3.1), reductions ~~in source level~~ have been applied to the ~~source levels of the~~ various noise sources. Figure 5-1 shows the representative noise measurements used ~~for this to calculate these reductions~~, which have been adjusted ~~for based on~~ the source levels given in Table 5-2-1. Details of the reductions in ~~sources levels~~source level for each of the Southall *et al.* (2019) marine mammal weightings ~~used for modelling~~ are given in Table 5-3-2.

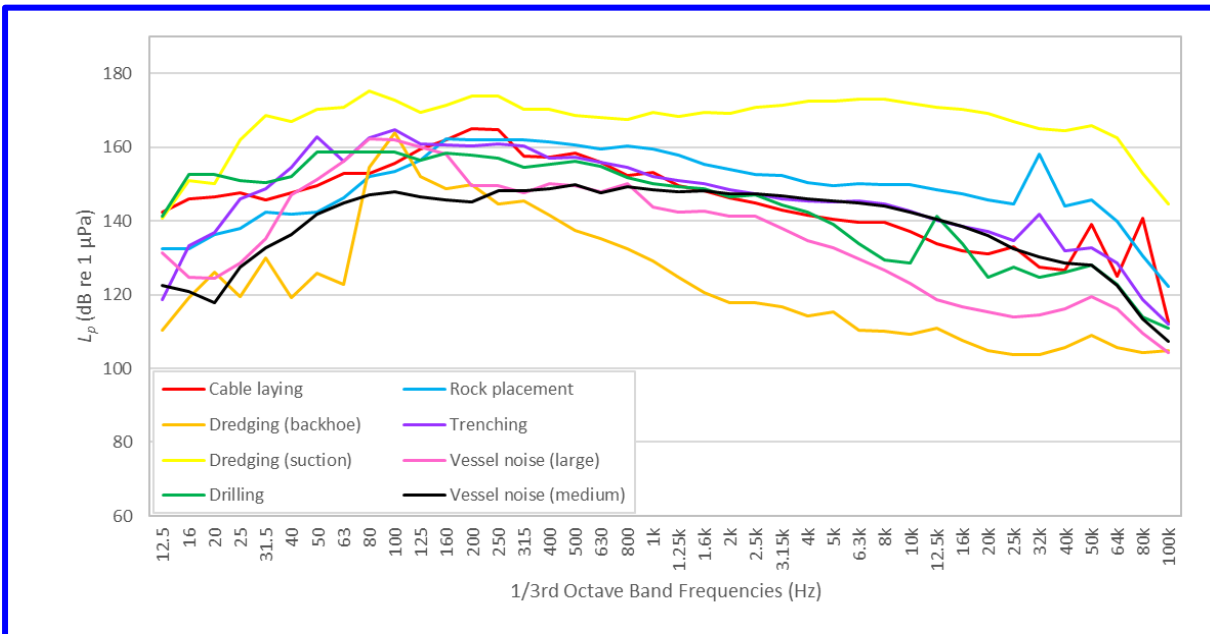
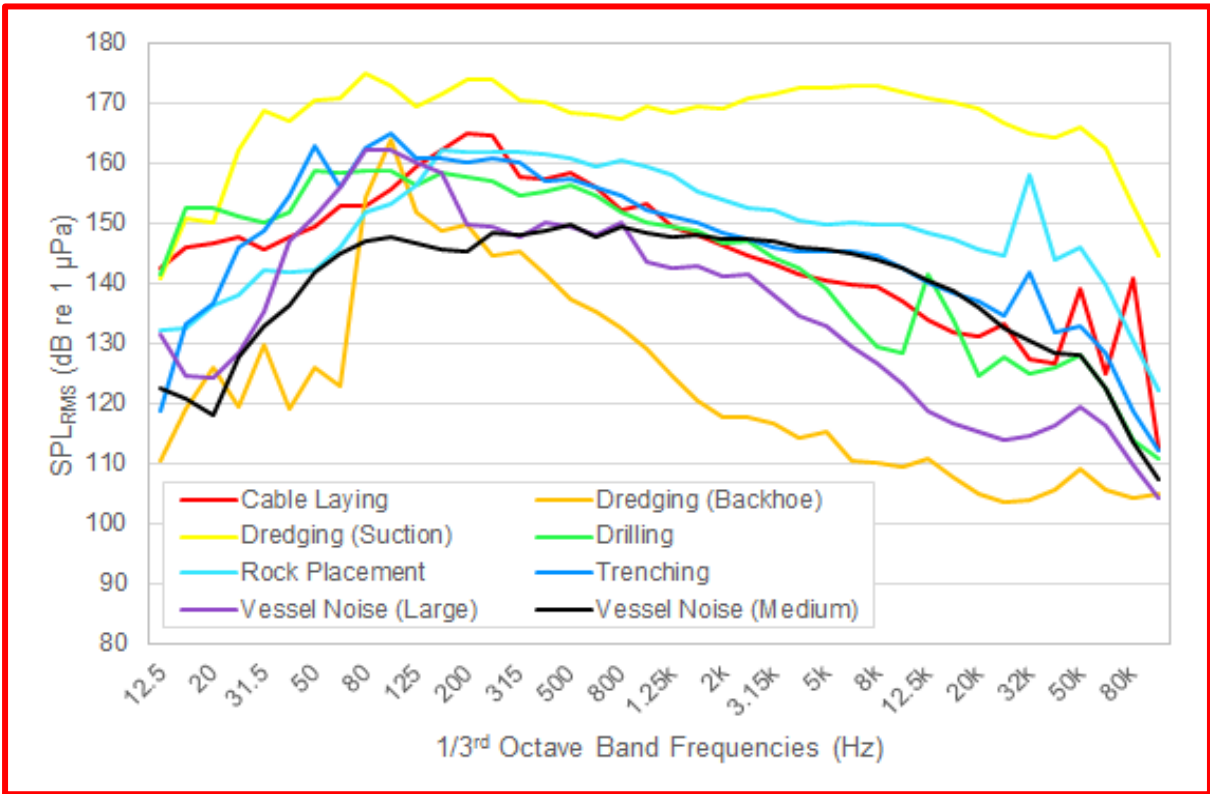


Figure 5-1: Summary of the *unweighted* 1/3rd octave frequency bands to which the *weightings* from Southall et al. (2019) *weightings were* have been applied *in the simple modelling*.

Table 5-3-2: Reductions in source level for the different construction noise sources activities considered when the weightings from Southall et al. (2019) weightings are applied.

Source Activity	Reduction in source level from the unweighted level			
	Reduction in L_p source level from the unweighted level (Southall et al., 2019)			
	LF	HF	VHF	PCW
Cable laying	3.6 2.5 dB re 1 μ Pa	22.9 25.6 dB re 1 μ Pa	23.9 26.6 dB re 1 μ Pa	13.2 13.8 dB re 1 μ Pa
Dredging (backhoe)	6.3 dB re 1 μ Pa	46.7 dB re 1 μ Pa	48.7 dB re 1 μ Pa	23.1 dB re 1 μ Pa
Dredging (backhoe and suction)	2.5 dB re 1 μ Pa	7.9 dB re 1 μ Pa	9.6 dB re 1 μ Pa	4.2 4.1 dB re 1 μ Pa
Drilling	4.0 dB re 1 μ Pa	25.8 dB re 1 μ Pa	48.7 28.4 dB re 1 μ Pa	13.2 dB re 1 μ Pa
Rock placement	1.6 dB re 1 μ Pa	11.9 dB re 1 μ Pa	12.5 dB re 1 μ Pa	8.2 dB re 1 μ Pa
Trenching	4.1 dB re 1 μ Pa	23.0 dB re 1 μ Pa	25.0 dB re 1 μ Pa	13.7 13.6 dB re 1 μ Pa
Vessel noise (large)	5.5 5.6 dB re 1 μ Pa	34.4 dB re 1 μ Pa	38.6 38.7 dB re 1 μ Pa	17.4 dB re 1 μ Pa
Vessel noise (medium)	1.3 dB re 1 μ Pa	13.2 dB re 1 μ Pa	16.1 dB re 1 μ Pa	5.1 dB re 1 μ Pa

Table 5-4 Summary of the 3 to Table 5-6 Summary of the 5 summarise the predicted impact ranges for these noise sources. All the sources in this section are considered non-impulsive or continuous. ~~As with the previous results, ranges smaller than 50 m (single strike) and 100 m (cumulative) have not been presented.~~

Given the modelled impact ranges, ~~any~~almost all marine ~~mammal~~mammals would have to be closer than ~~100~~50 m from the ~~continuous~~noise sources at the start of the activity to acquire the necessary exposure to induce for PTS onset as per Southall et al. (2019). ~~The exposure calculation assumes the same receptor swim speeds as the impact piling modelling in section 4. As explained in section 3.3, this would only mean that, with the possible exception of suction dredging and rock placement for stationary receptors. As previously iterated, these ranges only represent a range where the receptor reaches the 'onset' stage at these ranges, which is the minimum exposure that could potentially lead to the start of an effect and may only be marginal. In most hearing groups, the noise levels are low enough that there is this only represents a minimal risk.~~

, especially bearing in mind that many sources above are mobile. For fish, there is only a minimal risk of any injury or TTS ~~with reference to the SPL_{RMS}~~, using the L_p guidance for shipping and continuous noise sources in Popper et al. (2014), with all impact ranges predicted to be smaller than 50 m.

All the sources presented here produce much quieter levels than the results presented for impact piling in section 4.

Table 5-4 Summary of the 3: Weighted $L_{E,p,24h,wt}$ impact ranges for the different noise sources related to construction marine mammals using the ~~non-impulsive criteria from~~ Southall et al. (2019) for marine mammals ~~non-impulsive criteria for the various noise-making activities assuming a fleeing animal receptor.~~

Southall et al. (2019)	PTS (Non-impulsive)				TTS (Non-impulsive)			
	LF	HF	VHF	PCW	LF	HF	VHF	PCW
$L_{E,p,24h,wt}$ Weighted SEL_{cum} (Fleeing)	199 dB	198 dB	173 dB	201 dB	179 dB	178 dB	153 dB	181 dB
Cable laying	< 100 50 m	< 100 50 m	< 100 50 m	< 100 50 m	< 100 50 m	< 100 50 m	110 < 50 m	< 100 50 m
Dredging (backhoe)	< 100 50 m	< 100 50 m	< 100 50 m	< 100 50 m	< 100 50 m	< 100 50 m	< 100 50 m	< 100 50 m
Dredging (suction)	< 100 50 m	< 100 50 m	< 100 50 m	< 100 50 m	< 100 50 m	< 100 50 m	230 250 m	< 100 50 m

Southall <i>et al.</i> (2019) <i>L_{E,p,24h,wtd}</i> Weighted SEL_{cum} (Fleeting)	PTS (Non-impulsive)				TTS (Non-impulsive)			
	LF	HF	VHF	PCW	LF	HF	VHF	PCW
	199 dB	198 dB	173 dB	201 dB	179 dB	178 dB	153 dB	181 dB
Drilling	< 10050 m	< 10050 m	< 10050 m	< 10050 m	< 10050 m	< 10050 m	< 10050 m	< 10050 m
Rock placement	< 10050 m	< 10050 m	< 10050 m	< 10050 m	< 10050 m	< 10050 m	990 < 50 m	< 10050 m
Trenching	< 10050 m	< 10050 m	< 10050 m	< 10050 m	< 10050 m	< 10050 m	< 100820 m	< 10050 m
Vessel noise (large)	< 10050 m	< 10050 m	< 10050 m	< 10050 m	< 10050 m	< 10050 m	< 10050 m	< 10050 m
Vessel noise (medium)	< 10050 m	< 10050 m	< 10050 m	< 10050 m	< 10050 m	< 10050 m	< 10050 m	< 10050 m

Table 5-5 Summary of the 4: Weighted *L_{E,p,24h,wtd}* impact ranges for the different noise sources related to construction marine mammals using the non-impulsive criteria from Southall *et al.* (2019) for marine mammals non-impulsive criteria for the various noise-making activities assuming a stationary animal receptor.

Southall <i>et al.</i> (2019) <i>L_{E,p,24h,wtd}</i> Weighted SEL_{cum} (Stationary)	PTS (Non-impulsive)				TTS (Non-impulsive)				
	LF	HF	VHF	PCW	LF	HF	VHF	PCW	
	199 dB	198 dB	173 dB	173 dB 201 dB	179 dB	179 dB	178 dB	153 dB	181 dB
Cable laying	< 10050 m	< 10050 m	< 50 m	< 10050 m	< 10097 0 m	810 m	< 10050 m	2-31.3 km	11090 m
Dredging (backhoe)	< 10050 m	< 10050 m	< 50 m	< 10050 m	< 10050 m	< 100 m	< 10050 m	< 10050 m	< 10050 m
Dredging (suction)	< 10060 m	< 10050 m	560 m	570 < 50 m	< 10063 0 m	640 m	39038 0 m	4-34.2 km	420 m
Drilling	< 10050 m	< 10050 m	< 50 m	< 10050 m	< 10016 0 m	160 m	< 10050 m	200 m	< 10050 m
Rock placement	< 10050 m	< 10050 m	1.0 km	900 < 50 m	< 1002.0 km	2.1 km	41049 0 m	136.2 km	46056 0 m
Trenching	< 10050 m	< 10050 m	60 m	< 10050 m	< 10082 0 m	830 m	< 10050 m	1-91.8 km	12011 0 m
Vessel noise (large)	< 10050 m	< 10050 m	< 50 m	< 10050 m	< 10044 0 m	480 m	< 10050 m	14013 0 m	< 10050 m
Vessel noise (medium)	< 10050 m	< 10050 m	60 m	< 10050 m	< 10028 0 m	130 m	< 10050 m	< 1001.5 km	< 100 m

It should also be noted that that ranges for a stationary animal animals are theoretical only and are expected to be over-conservative as the assumption is for the animal receptor to remain stationary in respect to the noise source for the entire assessment period (24 hours), when in a number of these instances, the noise source itself is moving in most cases moves.

Table 5-6 Summary of the 5: Unweighted L_p impact ranges for ~~the different noise sources related to construction~~ fish using the Popper *et al* (2014) shipping and continuous noise criteria ~~from Popper *et al.* (2014) for fish (swim bladder involved in hearing)~~ for the various noise-making activities.

Popper <i>et al.</i> (2014) Unweighted SPL _{rms} [dBA]	Recoverable injury 170 dB re 1 μ Pa (48 hours)	TTS 158 dB re 1 μ Pa (12 hours)
Cable laying	< 50 m	< 50 m
Dredging (backhoe)	< 50 m	< 50 m
Dredging (suction)	< 50 m	< 50 m
Drilling	< 50 m	< 50 m
Rock placement	< 50 m	< 50 m
Trenching	< 50 m	< 50 m
Vessel noise (large)	< 50 m	< 50 m
Vessel noise (medium)	< 50 m	< 50 m

5.2 Operational WTG noise

When considering the ~~main source of underwater~~ noise from operational ~~WTGs will be~~ WTG, the primary noise source is a consequence of mechanically generated vibration from the rotating machinery in the ~~WTGs, which is~~ WTG transmitted into the ~~sea~~ water through the structure of the WTG tower and foundations (Nedwell *et al.*, 2003; Tougaard *et al.*, 2020). For a fixed-bottom foundation, this is the surface area of the cylindrical pile in the water column (or piles for multi-leg designs). The complexities of the acoustics in large structures such as these make it difficult to predict their effect on the resulting noise output (Tougaard *et al.*, 2020). Noise levels generated above the water surface are low enough that no significant airborne sound will pass from the air to the water.

Tougaard *et al.* (2020) published a study investigating ~~underwater~~ noise data from 17 operational WTGs in Europe and the United States, from 0.2 MW to 6.15 MW nominal power output. The paper identified the nominal power output and wind speed as the ~~two~~ primary driving factors for underwater noise generation. Although the datasets were acquired under different conditions, the authors devised a formula based on the published data for the operational ~~wind farms~~ wind farms, allowing a broadband noise level to be estimated based on the application of wind speed, turbine size (by nominal power output) and distance from the turbine:

$$L_{eq} = C + \alpha \log_{10} \left(\frac{\text{distance}}{100\text{m}} \right) + \beta \log_{10} \left(\frac{\text{wind speed}}{10\text{ms}^{-1}} \right) + \gamma \log_{10} \left(\frac{\text{turbine size}}{1\text{MW}} \right)$$

where C is a fixed constant and the coefficients α , β and γ are derived from the 17 empirical ~~data for the 17~~ datasets.

~~WTG~~ Indicative power outputs have been ~~used~~ assumed based on turbine size, ~~to calculate the impacts here.~~ For ABWP2, WTGs with rotor diameters of between 236 and 250 m ~~have been considered~~ ~~for this study.~~

~~As~~ The ~~maximum turbine~~ WTG sizes ~~considered~~ under consideration at ABWP2 are much larger than those used ~~for to develop~~ the estimation above, so caution must be ~~used~~ taken when considering the results presented in this section; no empirical data is available for large wind turbines close to the ~~specifications~~ specification proposed here. Research from Bellmann *et al.* (2024) using more up-to-date operational noise data from larger turbines currently installed (up to 8 MW) found that the predictions using the equation from Tougaard *et al.* (2020) are likely to overestimate the noise produced from the turbines, giving an extra level of conservatism for the estimations.

Figure 5-2 presents a level against range plot for the ~~two turbine~~ WTG sizes at ABWP2 using the Tougaard *et al.* (2020) ~~calculation~~ equation, assuming an average ~~6 m/s~~ wind speed ~~at hub height~~ of 6ms^{-1} . Although wind

speeds (and thus operational noise levels) may be greater than this, meaning this will not represent the typical condition. It is also worth noting that the background noise level will also naturally increase, due primarily to rougher seas, somewhat offsetting any additional impact this may have.

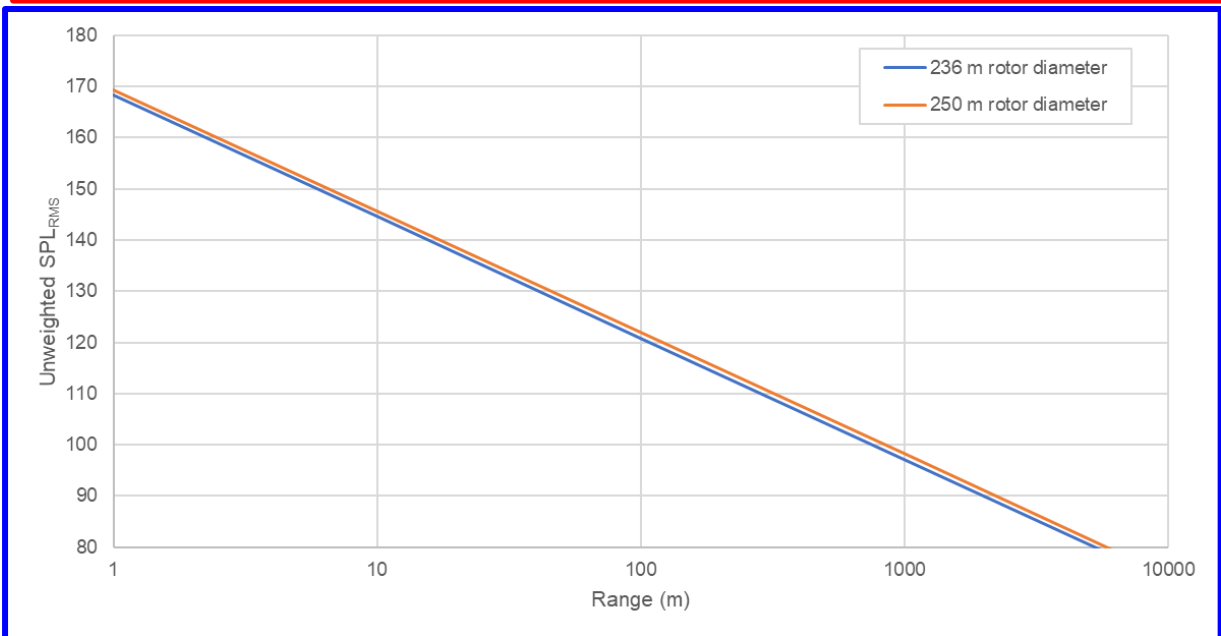
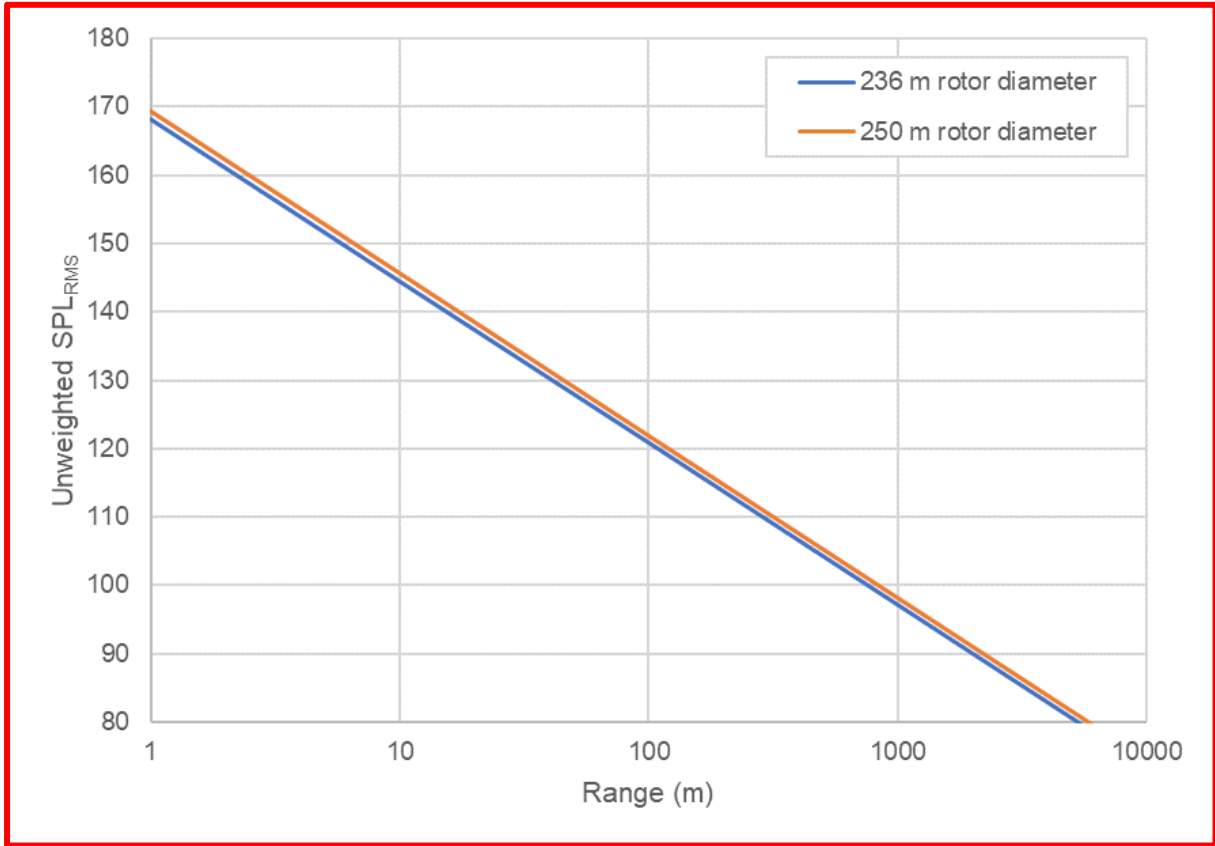


Figure 5-2: Predicted unweighted SPL_{RMS,L_p} from operational WTGs with rotor diameters of between 236 and 250 m using the calculation from Tougaard et al. (2020).

Using this data, a summary of the predicted impact ranges for operational WTG noise has been produced, shown presented in Table 5-8 Summary of the 6 and Table 5-9 Summary of the operational WTG noise-7. The operational WTG noisesource is considered a non-impulsive, or continuous, source.

For $SEL_{cum}L_{E,p,t}$ calculations, ita precautionary stationary animal has been used, and it is assumed that the operational WTG noise is present 24 hours a day, and similarly to the noise sources in section 5.1, both fleeing and stationary animals have been included for all SEL_{cum} criteria due to the low noise levels considered.

Table 5-7 Summary of the operational WTG noise impact ranges using the non-impulsive noise criteria from Southall et al. (2019) for marine mammals assuming a fleeing receptor

Southall et al. (2019) Weighted SEL_{cum}	PTS (non-impulsive)				TTS (non-impulsive)			
	LF	HF	VHF	PCW	LF	HF	VHF	PCW
	199 dB	198 dB	173 dB	201 dB	179 dB	178 dB	153 dB	181 dB
236 m rotor diameter	<100 m	<100 m	<100 m	<100 m	<100 m	<100 m	<100 m	<100 m
250 m rotor diameter	<100 m	<100 m	<100 m	<100 m	<100 m	<100 m	<100 m	<100 m

Table 5-8 Summary of the 6: Weighted $L_{E,p,24h,wtg}$ impact ranges for marine mammals using the Southall et al (2019) non-impulsive criteria for operational WTG noise impact ranges using the non-impulsive noise criteria from Southall et al. (2019) for marine mammals assuming a stationary receptor.

Southall et al. (2019) Weighted SEL_{cum} (Stationary)	PTS (Non-impulsive)				TTS (Non-impulsive)			
	LF	HF	VHF	PCW	LF	HF	VHF	PCW
	199 dB	198 dB	173 dB	201 dB	179 dB	178 dB	153 dB	181 dB
236 m rotor diameter	<10050 m	<10050 m	<10050 m	<10050 m	<10050 m	<10050 m	<10050 m	<10050 m
250 m rotor diameter	<10050 m	<10050 m	<10050 m	<10050 m	<10050 m	<10050 m	<10050 m	<10050 m

Based on the Southall et al. (2019) non-impulsive criteria, a marine mammal would need to remain within 50 m of the operational WTG for 12 hours to exceed threshold.

Table 5-9 Summary of the operational WTG noise-7: Unweighted L_p impact ranges for fish using the Popper et al (2014) shipping and continuous noise criteria from Popper et al. (2014) for fish (swim bladder involved in hearing) for the various noise-making activities.

Popper et al. (2014) Unweighted SPL_{RMS}	Recoverable injury 170 dB re 1 μ Pa (48 hours)	TTS 158 dB re 1 μ Pa (12 hours)
236 m rotor diameter	< 50 m	< 50 m
250 m rotor diameter	< 50 m	< 50 m

The results show that, for operational WTGs, injury risk is minimal. Increasing the wind speed does not lead to significant increases in the impact ranges.

Stöber and Thomsen (2021) produced a similar study of an operational wind turbine dataset to Tougaard et al. (2020) and it raises WTG noise datasets and raised the potential for behavioural disturbance caused by larger wind turbines WTGs. While prospective turbine WTG sizes are increasing, Stöber and Thomsen conclude (2021) concluded that these might only have limited impacts related to behavioural response on responses in marine mammals and fish, although there is considerable uncertainty in the criteria available to assess these. However, this. Based on the highly precautionary NOAA Level B behavioural threshold for continuous noise sources (120 dB SPL_{RMS} , (120 dB re 1 μ Pa (L_p) for non-impulsive noise; see NOAA, 2005)

that the study utilises. For ABWP2, it is estimated that ~~the WTGs~~ larger WTG may only ~~reach that~~ achieve the Level B behavioural threshold at ~~around~~ ranges of 120 m ~~away~~ using the Tougaard *et al.* (2020) equation (Figure 5-2). As the distance between turbines at ABWP2 is expected to be ~~considerably~~ greater than this ~~distance~~ ~~(the minimum separation distance is 944 m)~~, this would indicate that, any array effect from the turbines is not expected. Bellman *et al.* (2024) takes this further and shows that the predictions of underwater noise during the operational phase in Stöber and Thomsen (2021) represent significant over-estimations of the actual levels seen on site.

5.3 UXO clearance

It is possible, although highly unlikely, that UXO devices with a range of charge weights (or quantity of contained explosive), are present within ~~the boundaries of the proposed development~~ in and around the ABWP2 site. These would need to be cleared before any construction can begin. When modelling potential noise from UXO clearance, a variety of explosive types need to be considered, with the potential that many have been subject to degradation and burying over time. Two otherwise identical explosive devices are likely to produce different blasts in the case where one has spent an extended period on the seabed ~~or sits in a different topographical situation~~.

A selection of explosive sizes has been considered based on what might be present ~~and~~, in each case, it has been assumed that the maximum explosive charge in each device is present and ~~either~~ detonates with the clearance ~~(high-order) or a clearance method, such as deflagration (low-order), can be used~~. It is likely that a low-order technique will be the primary method of UXO clearance, with high-order clearance only to occur in exceptional circumstances.

5.3.1 Estimation of underwater noise levels

The noise produced by the detonation of explosives is affected by several different elements, only one of which can be easily ~~be~~ factored into a calculation: the charge weight ~~;~~ in this case, the charge weight is based on the equivalent weight of ~~T~~ ~~r~~ ~~i~~ ~~n~~ ~~i~~ ~~t~~ ~~r~~ ~~o~~ ~~t~~ ~~o~~ ~~l~~ ~~u~~ ~~e~~ ~~n~~ ~~e~~ ~~–~~ (TNT). Many other elements relating to its situation (e.g., its design, composition, age, position, orientation, whether it is covered by sediment) and exactly how they will affect the sound produced by detonation are usually unknown and cannot be directly considered in this type of assessment. This leads to a high ~~degree~~ level of uncertainty in the estimation of ~~the source~~ noise ~~level~~ levels. A ~~worst case~~ precautionary estimation has therefore been used for calculations in this study, assuming the UXO to be detonated is not buried, degraded or subject to any other significant attenuation from its ~~as new~~ 'as-new' condition. A 'high-order' clearance technique, using an external 'donor charge' initiator to detonate the explosive material in the UXO, theoretically produces a blast wave equivalent to the full detonation of the device.

The consequence of this is that the noise ~~levels~~ produced, particularly by the larger explosives under consideration, are likely to be over-estimated as some degree of ~~attenuation~~ (i.e., from topography, burying, degradation, ~~orientation~~) would be expected.

5.3.2 Low-order clearance

The primary choice for any UXO clearance at ABWP2 will be a low-order technique, in order to reduce the consequences of noise caused by detonation of the main charge of the UXO. Deflagration is one such alternative technique, intended to result in a low-order burn of the explosive material in a UXO, which destroys, but does not detonate, the internal explosive material.

Where the technique proceeds as intended, it is still not without noise impact. The process requires an initial shaped explosive donor charge, typically less than 0.5 kg, to breach the casing and ignite the internal high explosive (HE) material without full detonation. The shaped charge and burn will both produce noise, although it will be significantly less than the high-order detonation of the UXO. Deflagration may not destroy all of the HE,

which would necessitate further low-order clearance events or collection of the remnants. There is also the possibility (although rare) that the deflagration could produce an unintentional high-order event.

For calculation of the deflagration scenario, resulting in total destruction of the HE material, it is anticipated that the initial shaped charge is the greatest source of noise (Cheong *et al.*, 2020). The shaped charge is treated as a bulk charge with a net explosive quantity (NEQ) determined according to the size of UXO on which it is placed. The prediction of this impact is based on a charge weight of 0.5 kg. The most precautionary scenario would, of course, be a high-order detonation with maximum pressures from complete detonation of the UXO.

5.3.3 High-order clearance

High-order clearance technique is not a proposed UXO clearance methodology, although for completeness has been included in this assessment. The only reason for a high-order clearance at ABWP2 would be as a last resort if the use of a less intrusive or quieter technique is not possible, or if the low-order technique accidentally results in a high-order detonation. A high-order clearance would involve detonating the UXO including all the HE material contained within.

The maximum equivalent charge weight for the potential UXO devices that could be present ~~within the proposed development~~ at ABWP2 has been estimated as 800 kg. This has been modelled alongside a range of smaller devices, at charge weights of 25, 55, 120, 240, 525 and 700 kg. In each case, an additional donor weight of 0.5 kg has been included to initiate detonation. ~~Low-order deflagration has also been assessed, which assumes that the donor or shaped charge (charge weight of 0.5 kg) detonates fully to initiate a burnout of the explosive but without the follow-up detonation of the UXO.~~ No mitigation has been considered for this modelling.

Estimation of the source noise level for each charge weight has been carried out in accordance with the methodology of Soloway and Dahl (2014), which follows Arons (1954) and the Marine Technical Directorate Ltd (MTD) (1996). ~~This is covered in more detail in section 5.3.4.~~

5.3.4 ~~5.3.2~~ Estimation of underwater noise propagation

For this assessment, the attenuation of the noise from UXO detonation has been accounted for in calculations using geometric spreading and a sound absorption coefficient, primarily using the methodologies cited in Soloway and Dahl (2014), which establishes a trend based on measured data in open water. These are, for ~~SPL~~ $L_{p,pk}$:

$$L_{p,pk} = 52.4 \times 10^6 \left(\frac{R}{W^{1/3}} \right)^{-1.13}$$

and for ~~SEL~~ $L_{E,p}$:

$$L_{E,p} = 6.14 \times \log_{10} \left(W^{1/3} \left(\frac{R}{W^{1/3}} \right)^{-2.12} \right) + 219$$

where W is the equivalent charge weight for TNT in kilograms and R is the range from the source.

These equations give a relatively simple calculation which can be used to give an indication of the range of effect. The equation does not consider variable bathymetry or seabed type, and thus the calculation results will be the same regardless of where it is used. An attenuation correction can be added to the Soloway and Dahl (2014) equations for the absorption over long ranges (i.e., of the order of thousands of metres), based on ~~measurements of high intensity noise propagation taken in the North Sea and Irish Sea~~ typical sound frequencies associated with UXO clearance. This uses standard frequency-based absorption coefficients for the seawater conditions expected in the region.

Despite this attenuation correction, the resulting noise levels still need to be considered carefully, for example, $SPL_{peak,L,p,k}$ noise levels over larger distances are difficult to predict accurately (von Benda-Beckmann *et al.*, 2015). Soloway and Dahl (2014) only verify results from the equation above for small charges at ranges of less than 1 km, although the long range and high charge weight results are similar to the measurements presented by von Benda-Beckmann *et al.* (2015). At longer ranges, greater confidence is expected with the $SEL_{E,p}$ calculations. However, Ocean Winds (2024) indicates that, based on measurements of deflagration noise in the Moray Firth, these calculations are likely to produce a higher, and therefore precautionary, prediction of noise levels than are seen in practice.

A further limitation in the Soloway and Dahl (2014) equations ~~that must be considered~~ are that variations in noise levels at different depths are not considered. Where animals are swimming near the surface, the acoustics can cause the noise level, and hence the exposure, to be lower (MTD, 1996). The risk to animals near the surface may therefore be lower than indicated by the impact ranges, and therefore the results presented here can be considered conservative in respect of the ~~impact~~ impacts at different depths.

Additionally, an impulsive wave tends to be smoothed (i.e., the pulse becomes longer) over distance (Cudahy and Parvin, 2001), meaning that the injurious potential of a wave at greater range can be even lower than just a reduction in the absolute noise level. An assessment in respect of $SEL_{E,p}$ is considered preferential at long range as it considers the overall energy, and the degree of smoothing of the peak with increasing distance is ~~less~~ critical.

In light of this, the selection of assessment criteria ~~must also be considered in light of this~~ needs careful consideration. As discussed in section 2.2.1, the smoothing of the pulse at range means that a pulse may be considered non-impulsive with distance, suggesting that, at greater ranges, it may be more appropriate to use the non-impulsive criteria. Based on impulsive noise from piling, this consideration may begin at ranges of 3.5 km (Hastie *et al.*, 2019) to 5 km (Matei *et al.*, 2024), although, as blast noise is inherently more impulsive than piling, the transition from full impulsivity may occur at a greater distance from the UXO source location.

A summary of the unweighted UXO clearance source levels, calculated using the equations above, are given in Table 5-10 ~~Summary of the unweighted SPL_{peak} and SEL_{ss}~~ 8.

Table 5-10 ~~Summary of the unweighted SPL_{peak} and SEL_{ss}~~ 8: List of the $L_{p,pk}$ and $L_{E,p}$ source levels used for UXO clearance modelling.

Charge weight	SPL_{peak} source level (dB re 1 μPa @ 1 m)	SEL_{ss} source level (dB re 1 $\mu Pa^2 s$ @ 1 m)
Low order 0.5 kg	272.1 dB re 1 μPa	217.1 dB re 1 $\mu Pa^2 s$
25 kg + donor	284.9 dB re 1 μPa	228.0 dB re 1 $\mu Pa^2 s$
55 kg + donor	287.5 dB re 1 μPa	230.1 dB re 1 $\mu Pa^2 s$
120 kg + donor	290.0 dB re 1 μPa	232.3 dB re 1 $\mu Pa^2 s$
240 kg + donor	292.3 dB re 1 μPa	234.2 dB re 1 $\mu Pa^2 s$
525 kg + donor	294.8 dB re 1 μPa	236.4 dB re 1 $\mu Pa^2 s$
700 kg + donor	295.8 dB re 1 μPa	237.2 dB re 1 $\mu Pa^2 s$
800 kg + donor	296.2 dB re 1 μPa	237.5 dB re 1 $\mu Pa^2 s$

5.3.5 ~~5.3.3~~ Impact ranges

Table 5-11 ~~Summary of the PTS and TTS~~ 9 to Table 5-14 ~~Summary of the~~ 12 present the impact ranges for UXO ~~detonation~~ clearance, considering various charge weights and impact criteria. It should be noted that Popper *et al.* (2014) gives specific impact criteria for explosions (Table 2-8 ~~Summary of the qualitative effects on fish from continuous noise from~~ 5). A UXO detonation source is defined as a single pulse, and as such the ~~$SEL_{cum,L,E,p,wd}$~~ criteria from Southall *et al.* (2019) have been given as SEL_{ss} single pulse values in the following

tables below. Thus, and fleeing animal assumptions do not apply. As with the previous sections, ranges smaller than 50 m have not been presented.

Although the impact ranges in Table 5-11 Summary of the PTS and TTS 9 to Table 5-14 Summary of the 12 are large, the duration the noise is present must also be considered. For the detonation of a UXO, each explosion is a single noise event, compared to the multiple pulse nature and longer durations of impact piling.

Table 5-11 Summary of the PTS and TTS 9: Unweighted $L_{p,pk}$ impact ranges for UXO detonation marine mammals using the impulsive, unweighted SPL_{peak} noise criteria from Southall et al. (2019) for marine mammals impulsive criteria for UXO clearance noise.

Southall et al. (2019) $L_{p,pk}$ Unweighted SPL_{peak}	PTS (Impulsive)				TTS (Impulsive)			
	LF	HF	VHF	PCW	LF	HF	VHF	PCW
	219 dB	230 dB	202 dB	218 dB	213 dB	224 dB	196 dB	212 dB
Low order (0.5 kg)	220 m	70 m	1.2 km	240 m	410 m	130 m	2.3 km	450 m
25 kg + donor	820 m	260 m	4.6 km	910 m	1.5 km	490 m	8.5 km	1.6 km
55 kg + donor	1.0 km	340 m	6.0 km	1.1 km	1.9 km	640 m	11 km	2.1 km
120 kg + donor	1.3 km	450 m	7.8 km	1.5 km	2.5 km	830 m	14 km	2.8 km
240 kg + donor	1.7 km	560 m	9.8 km	1.9 km	3.2 km	1.0 km	18 km	3.5 km
525 kg + donor	2.2 km	730 m	12 km	2.5 km	4.1 km	1.3 km	23 km	4.6 km
700 kg + donor	2.4 km	810 m	14 km	2.7 km	4.5 km	1.4 km	25 km	5.0 km
800 kg + donor	2.6 km	840 m	14 km	2.8 km	4.7 km	1.5 km	26 km	5.3 km

Table 5-12 Summary of the PTS and TTS 10: Weighted $L_{p,wtd}$ (single pulse) impact ranges for UXO detonation marine mammals using the impulsive, weighted SEL_{ss} noise criteria from Southall et al. (2019) for marine mammals impulsive criteria for UXO clearance noise.

Southall et al. (2019) $L_{p,wtd}$ Weighted SEL_{ss} (Single pulse)	PTS (Impulsive)				TTS (Impulsive)			
	LF	HF	VHF	PCW	LF	HF	VHF	PCW
	183 dB	185 dB	155 dB	185 dB	168 dB	170 dB	140 dB	170 dB
Low order (0.5 kg)	320 m	< 50 m	110 m	60 m	4.5 km	< 50 m	930 930m	800 m
25 kg + donor	2.2 km	< 50 m	570 m	390 m	29 km	150 m	2.4 km	5.2 km
55 kg + donor	3.2 km	< 50 m	740 m	570 m	41 km	210 m	2.8 km	7.5 km
120 kg + donor	4.7 km	< 50 m	950 m	830 m	57 km	300 m	3.2 km	10 km
240 kg + donor	6.5 km	< 50 m	1.1 km	1.1 km	76 km	390 m	3.5 km	14 km
525 kg + donor	9.5 km	50 m	1.4 km	1.6 km	100 km	530 m	4.0 km	19 km
700 kg + donor	10 km	60 m	1.5 km	1.9 km	110 km	590 m	4.1 km	22 km
800 kg + donor	11 km	60 m	1.6 km	2.0 km	120 km	620 m	4.2 km	23 km

Table 5-13 Summary of the PTS and TTS-11: Weighted $L_{E,p,wt}$ (single pulse) impact ranges for UXO detonation marine mammals using the non-impulsive, weighted $SEL_{L_{ss}}$ noise criteria from Southall et al. (2019) for marine mammals non-impulsive criteria for UXO clearance noise.

Southall et al. (2019) $L_{E,p,wt}$ Weighted $SEL_{L_{ss}}$ (single pulse)	PTS (Non-impulsive)				TTS (Non-impulsive)			
	LF	HF	VHF	PCW	LF	HF	VHF	PCW
	199 dB	198 dB	173 dB	201 dB	179 dB	178 dB	153 dB	181 dB
Low order (0.5 kg)	< 50 m	< 50 m	< 50 m	< 50 m	650 m	< 50 m	150 m	110 m
25 kg + donor	130 m	< 50 m	< 50 m	< 50 m	4.4 km	< 50 m	730 m	790 m
55 kg + donor	190 m	< 50 m	< 50 m	< 50 m	6.4 km	60 m	940 m	1.1 km
120 kg + donor	280 m	< 50 m	70 m	< 50 m	9.4 km	80 m	1.1 km	1.6 km
240 kg + donor	390 m	< 50 m	100 m	70 m	13 km	110 m	1.4 km	2.3 km
525 kg + donor	570 m	< 50 m	130 m	100 m	18 km	160 m	1.7 km	3.3 km
700 kg + donor	660 m	< 50 m	150 m	110 m	21 km	180 m	1.8 km	3.8 km
800 kg + donor	700 m	< 50 m	160 m	120 m	22 km	190 m	1.8 km	4.1 km

Table 5-14 Summary of the 12: Unweighted $L_{p,pl}$ impact ranges for UXO detonation fish using the unweighted $SPL_{L_{peak}}$ explosion noise criteria from Popper et al. (2014) for species of fish explosions criteria for UXO clearance noise.

Popper et al. (2014) $L_{p,pl}$ Unweighted $SPL_{L_{RMS}}$	Mortality and potential mortal injury	
	234 dB	229 dB
Low order (0.5 kg)	< 50 m	80 m
25 kg + donor	170 m	290 m
55 kg + donor	230 m	380 m
120 kg + donor	300 m	490 m
240 kg + donor	370 m	620 m
525 kg + donor	490 m	810 m
700 kg + donor	530 m	890 m
800 kg + donor	560 m	930 m

5.3.6 5.3.4 Summary

The maximum PTS range onset ranges calculated for the largest high-order UXO clearance is 14 km for the VHF cetacean category, when considering the unweighted $SPL_{L_{peak}}$ criteria. For $SEL_{L_{ss}}$ criteria, the largest PTS onset range is calculated for LF cetaceans with a predicted impact range of 11 km using the impulsive noise criteria. As explained earlier previously mentioned, this assumes no degradation of the UXO and no smoothing of the pulse over that distance, which is a very precautionary approach. Although an assumption of non-pulse could underestimate using the non-impulsive criteria could underestimate the potential impact (Martin et al., 2020) (the equivalent non-impulsive criteria range based on for LF cetacean non-pulse criteria cetaceans is 700 m), it is likely that the long-range smoothing of the pulse peak would reduce its potential harm and, making the maximum 'impulsive' range for all species is very precautionary.

A low-order clearance would produce a maximum PTS onset impact range of 1.2 km for VHF cetaceans using the $L_{p,pk}$ criteria, with all other species groups lower than this. A low-order methodology is expected to be used for any potential UXO clearance at ABWP2, with high-order being a last resort.

6 Summary and conclusions

Subacoustech Environmental ~~have~~has undertaken a study on behalf of GoBe Consultants to ~~assess the potential~~model the impact ranges caused by underwater noise from impact piling and its effects on marine fauna during construction ~~and operation of the proposed~~of ABWP2, ~~located~~an offshore wind farm in the ~~southern~~Irish Sea.

The level of underwater noise from the installation of ~~turbine~~monopile foundations using impact piling during construction has been ~~estimated~~modelled using the INSPIRE, ~~a~~ semi-empirical underwater noise model. ~~The modelling~~This industry standard approach considers a wide variety of input parameters including bathymetry, and precautionary values for pile diameter, hammer blow energy, strike rate, ~~and receptor fleeing~~flee speed of the receptor.

Five ~~representative~~modelling locations (~~three for the WTGs~~WTG installations and two for ~~OSPs~~)OSP, were chosen to give spatial variation across ABWP2 as well as ~~account~~accounting for changes in water depth ~~around the Array Area and OSPs. At each location, two sizes of monopile have been considered covering the various foundation options.~~. The WTG foundation scenarios considered 11 m diameter piles installed with a maximum hammer energy of 3,500 kJ. The OSP foundation scenario considers the same maximum hammer energy, but with a larger pile diameter of 14 m.

~~The loudest levels of noise and greatest impact ranges have been predicted for the larger monopile foundation scenarios at the SW WTG location and the South OSP locations due to the proximity to deep water areas to the south and east of the site.~~

~~The modelling results were analysed in terms of relevant noise metrics and criteria to assess the effects of the impact piling on marine mammals (Southall et al., 2019) and fish (Popper et al., 2014), which have been used to aid biological assessments.~~

For marine mammals, the largest predicted impact ranges were calculated for the LF cetaceans at the Central (WT28) modelling location, with maximum PTS ranges ~~were predicted for LF cetaceans, with~~of 7.1 km. For fish, maximum recoverable injury (203 dB $L_{E,p,24h}$) ranges of up to ~~196.0 km predicted for the larger piles at the south of the site. For fish, the largest recoverable injury ranges (203 dB SEL_{cum})~~were predicted ~~to be less than 100 m for~~for stationary receptors. When a fleeing receptor, ~~increasing to a maximum of 7.9 km for a stationary receptor~~ was considered, this range reduced to less than 50 m.

An alternative ramp-up scenario was considered, to reduce the energy during the earlier stages of the piling (Table 3-3 and Table 3-5), and using this the maximum PTS impact ranges reduced to 870 m for LF cetaceans and 5.5 km for recoverable injury in stationary fish. The use of a low-noise hammer as mitigation also reduced the maximum impact ranges down to 720 m for LF cetaceans PTS and 3.4 km for stationary fish.

A review was undertaken of various underwater noise thresholds or limits for offshore wind farm installation in use in the EU as requested in the Further Information Request from ACP. The Danish requirements offer a combination of consideration of the varying sensitivities of different species groups expected to be present in Irish waters, site specific characteristics, and the total noise produced by an impact piling event. The Danish guidance uses the r_{safe} concept, effectively the PTS impact range, a distance which must be made clear of any marine mammals at the start of piling. Targeting an r_{safe} distance of 1 km, this target can be achieved using either the Mitigated (4 dB noise reduction) or the Alternative ramp up scenarios, without additional noise abatement.

Noise sources other than impact piling ~~were~~have been considered ~~using a high-level, simple modelling approach, including~~, which include noise from cable laying, dredging, drilling, rock placement, trenching, vessel movements noise, and operational ~~WTG noise. The predicted noise levels for the other construction noise~~

~~sources and during WTC operation are well below those predicted for impact piling noise~~ WTCs. The risk of any potentially injurious effects to fish or marine mammals from these sources are expected to be minimal as the noise emissions from these are close to, or below, the appropriate injury criteria, even when very close to the source of the noise.

Potential noise from UXO clearance ~~has~~ ~~was~~ also ~~been~~ considered ~~at the proposed development, and for the expected UXO clearance noise,~~ across the ABWP2 site. There is a risk of PTS onset up to 1.2 km for VHF cetaceans (unweighted $L_{p,pk}$ criteria) with the use of the expected technique of low-order clearance. In the event that a high-order detonation does occur, the maximum PTS onset range is up to 14 km ~~for~~ from the largest UXO device considered, ~~–~~ (800 kg + donor charge), using the unweighted ~~SPL~~ $L_{p,pk}$ criteria for VHF cetaceans. However, this is likely to be highly precautionary as the impact range is based on a ~~worst case~~ criterion and calculation methodology that does not account for any smoothing of the pulse over long ranges, which would reduce the pulse peak and other characteristics of the sound that cause injury.

By its nature, numerical modelling will produce results that indicate a precise range at which a criterion will be reached, but this does not reflect the inherent uncertainty in the physical processes, including many that change constantly under real world conditions. While the results present specific ranges at which each impact threshold is met based on the modelling results, the ranges should be taken as indicative in determining where environmental effects may occur in receptors during the proposed operations.

The outputs of this modelling have been used to inform ~~analysis~~ assessments of the ~~impacts of~~ underwater noise impacts on marine mammals and fish ~~in their respective reports~~ at ABWP2 within the EIAR and NIS.

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Appendix A Additional modelling results

~~Following from the Southall et al. (2019) modelled impact piling ranges presented in section 4 of the main report, the modelling results for the non-impulsive criteria from impact piling noise at ABWP2, as discussed in section 2.3.1, are presented below.~~

A.1 First pile strike

Table A-1 to Table A-45 present the single strike ($L_{p,pk}$ and $L_{p,RMS}$) impact ranges when considering the first pile strike of the WTG and OSP foundation modelling scenarios.

A.1.1 WTG foundations

Worst Case scenario

~~Table A-1 Summary of the weighted SEL_{cum} : Unweighted $L_{p,pk}$ impact areas and ranges for marine mammals using the Southall et al. (2019) non-impulsive/impulsive criteria for the 11 m monopile first hammer strike during the WTG foundation modelling at the NW WTG installation scenario at the North (WT03) modelling location assuming a fleeing animal.~~

Southall et al. (2019) Unweighted $L_{p,pk}$		WTG foundation (first strike)			
		Area	Maximum range	Minimum range	Mean range
PTS (Impulsive)	LF (219 dB)	< 0.01 km ²	< 50 m	< 50 m	< 50 m
	HF (230 dB)	< 0.01 km ²	< 50 m	< 50 m	< 50 m
	VHF (202 dB)	0.48 km ²	400 m	380 m	390 m
	PCW (218 dB)	< 0.01 km ²	< 50 m	< 50 m	< 50 m
TTS (Impulsive)	LF (213 dB)	0.02 km ²	80 m	80 m	80 m
	HF (224 dB)	< 0.01 km ²	< 50 m	< 50 m	< 50 m
	VHF (196 dB)	2.4 km ²	930 m	840 m	880 m
	PCW (212 dB)	0.03 km ²	100 m	100 m	100 m

Table A-2: Unweighted $L_{p,RMS}$ impact areas and ranges for marine mammals using the NOAA (2005) impulsive behavioural disturbance criteria for the first hammer strike during the WTG foundation installation scenario at the North (WT03) modelling location.

NOAA (2005) Unweighted $L_{p,RMS}$		WTG foundation (first strike)			
		Area	Maximum range	Minimum range	Mean range
Level B	160 dB	370 km ²	14 km	5.1 km	11 km

Table A-3: Unweighted $L_{p,pk}$ impact areas and ranges for fish using the Popper et al. (2014) pile driving criteria for the first hammer strike during the WTG foundation installation scenario at the North (WT03) modelling location.

Popper et al. (2014) Unweighted $L_{p,pk}$		WTG foundation (first strike)			
		Area	Maximum range	Minimum range	Mean range
Pile driving	213 dB	0.02 km ²	80 m	80 m	80 m
	207 dB	0.12 km ²	200 m	190 m	190 m

Table A-4: Unweighted $L_{p,pk}$ impact areas and ranges for marine mammals using the Southall et al. (2019) impulsive criteria for the first hammer strike during the WTG foundation installation scenario at the Central (WT28) modelling location.

Southall et al. (2019) Unweighted $L_{p,pk}$		WTG foundation (first strike)			
		Area	Maximum range	Minimum range	Mean range
PTS (Impulsive)	LF (219 dB)	< 0.01 km ²	< 50 m	< 50 m	< 50 m
	HF (230 dB)	< 0.01 km ²	< 50 m	< 50 m	< 50 m
	VHF (202 dB)	0.6 km ²	460 m	410 m	440 m
	PCW (218 dB)	< 0.01 km ²	< 50 m	< 50 m	< 50 m
TTS (Impulsive)	LF (213 dB)	0.03 km ²	90 m	90 m	90 m
	HF (224 dB)	< 0.01 km ²	< 50 m	< 50 m	< 50 m

	VHF (196 dB)	3.0 km ²	1.1 km	770 m	970 m
	PCW (212 dB)	0.03 km ²	110 m	100 m	110 m

Table A-5: Unweighted $L_{p,RMS}$ impact areas and ranges for marine mammals using the NOAA (2005) impulsive behavioural disturbance criteria for the first hammer strike during the WTG foundation installation scenario at the Central (WT28) modelling location.

NOAA (2005)		WTG foundation (first strike)			
Unweighted $L_{p,RMS}$		Area	Maximum range	Minimum range	Mean range
Level B	160 dB	510 km ²	17 km	5.0 km	12 km

Table A-6: Unweighted $L_{p,pk}$ impact areas and ranges for fish using the Popper et al. (2014) pile driving criteria for the first hammer strike during the WTG foundation installation scenario at the Central (WT28) modelling location.

Popper et al. (2014)		WTG foundation (first strike)			
Unweighted $L_{p,pk}$		Area	Maximum range	Minimum range	Mean range
Pile driving	213 dB	0.03 km ²	90 m	90 m	90 m
	207 dB	0.15 km ²	220 m	210 m	220 m

Table A-7: Unweighted $L_{p,pk}$ impact areas and ranges for marine mammals using the Southall et al. (2019) impulsive criteria for the first hammer strike during the WTG foundation installation scenario at the South west (WT53) modelling location.

Southall et al. (2019)		WTG foundation (first strike)			
Unweighted $L_{p,pk}$		Area	Maximum range	Minimum range	Mean range
PTS (Impulsive)	LF (219 dB)	< 0.01 km ²	< 50 m	< 50 m	< 50 m
	HF (230 dB)	< 0.01 km ²	< 50 m	< 50 m	< 50 m
	VHF (202 dB)	0.42 km ²	400 m	340 m	370 m
	PCW (218 dB)	< 0.01 km ²	< 50 m	< 50 m	< 50 m
TTS (Impulsive)	LF (213 dB)	0.02 km ²	80 m	80 m	80 m
	HF (224 dB)	< 0.01 km ²	< 50 m	< 50 m	< 50 m
	VHF (196 dB)	2.2 km ²	930 m	730 m	840 m
	PCW (212 dB)	0.03 km ²	90 m	90 m	90 m

Table A-8: Unweighted $L_{p,RMS}$ impact areas and ranges for marine mammals using the NOAA (2005) impulsive behavioural disturbance criteria for the first hammer strike during the WTG foundation installation scenario at the South west (WT53) modelling location.

NOAA (2005)		WTG foundation (first strike)			
Unweighted $L_{p,RMS}$		Area	Maximum range	Minimum range	Mean range
Level B	160 dB	510 km ²	15 km	4.1 km	12 km

Table A-9: Unweighted $L_{p,pk}$ impact areas and ranges for fish using the Popper et al (2014) pile driving criteria for the first hammer strike during the WTG foundation installation scenario at the South west (WT53) modelling location.

Popper et al. (2014)		WTG foundation (first strike)			
Unweighted $L_{p,pk}$		Area	Maximum range	Minimum range	Mean range
Pile driving	213 dB	0.02 km ²	80 m	80 m	80 m
	207 dB	0.11 km ²	190 m	170 m	180 m

Alternative ramp-up

Table A-10: Unweighted $L_{p,pk}$ impact areas and ranges for marine mammals using the Southall et al. (2019) impulsive criteria for the first hammer strike during the WTG foundation installation (alternative ramp-up) scenario at the North (WT03) modelling location.

Southall et al. (2019) Unweighted $L_{p,pk}$		WTG foundation (first strike), alternative ramp-up			
		Area	Maximum range	Minimum range	Mean range
PTS (impulsive)	LF (219 dB)	< 0.01 km ²	< 50 m	< 50 m	< 50 m
	HF (230 dB)	< 0.01 km ²	< 50 m	< 50 m	< 50 m
	VHF (202 dB)	0.22 km ²	270 m	260 m	270 m
	PCW (218 dB)	< 0.01 km ²	< 50 m	< 50 m	< 50 m
TTS (impulsive)	LF (213 dB)	0.01 km ²	60 m	60 m	60 m
	HF (224 dB)	< 0.01 km ²	< 50 m	< 50 m	< 50 m
	VHF (196 dB)	1.2 km ²	630 m	590 m	610 m
	PCW (212 dB)	0.01 km ²	70 m	70 m	70 m

Table A-11: Unweighted $L_{p,RMS}$ impact areas and ranges for marine mammals using the NOAA (2005) impulsive behavioural disturbance criteria for the first hammer strike during the WTG foundation installation (alternative ramp-up) scenario at the North (WT03) modelling location.

NOAA (2005) Unweighted $L_{p,RMS}$		WTG foundation (first strike), alternative ramp-up			
		Area	Maximum range	Minimum range	Mean range
Level B	160 dB	240 km ²	10 km	4.7 km	8.6 km

Table A-12: Unweighted $L_{p,pk}$ impact areas and ranges for fish using the Popper et al. (2014) pile driving criteria for the first hammer strike during the WTG foundation installation (alternative ramp-up) scenario at the North (WT03) modelling location.

Popper et al. (2014) Unweighted $L_{p,pk}$		WTG foundation (first strike), alternative ramp-up			
		Area	Maximum range	Minimum range	Mean range
Pile driving	213 dB	0.01 km ²	60 m	60 m	60 m
	207 dB	0.06 km ²	130 m	130 m	130 m

Table A-13: Unweighted $L_{p,pk}$ impact areas and ranges for marine mammals using the Southall et al. (2019) impulsive criteria for the first hammer strike during the WTG foundation installation (alternative ramp-up) scenario at the Central (WT28) modelling location.

Southall et al. (2019) Unweighted $L_{p,pk}$		WTG foundation (first strike), alternative ramp-up			
		Area	Maximum range	Minimum range	Mean range
PTS (impulsive)	LF (219 dB)	< 0.01 km ²	< 50 m	< 50 m	< 50 m
	HF (230 dB)	< 0.01 km ²	< 50 m	< 50 m	< 50 m
	VHF (202 dB)	0.28 km ²	310 m	290 m	300 m
	PCW (218 dB)	< 0.01 km ²	< 50 m	< 50 m	< 50 m
TTS (impulsive)	LF (213 dB)	0.01 km ²	60 m	60 m	60 m
	HF (224 dB)	< 0.01 km ²	< 50 m	< 50 m	< 50 m
	VHF (196 dB)	1.5 km ²	740 m	600 m	680 m
	PCW (212 dB)	0.02 km ²	70 m	70 m	70 m

Table A-14: Unweighted $L_{p,RMS}$ impact areas and ranges for marine mammals using the NOAA (2005) impulsive behavioural disturbance criteria for the first hammer strike during the WTG foundation installation (alternative ramp-up) scenario at the Central (WT28) modelling location.

NOAA (2005) Unweighted $L_{p,RMS}$		WTG foundation (first strike), alternative ramp-up			
		Area	Maximum range	Minimum range	Mean range
Level B	160 dB	310 km ²	13 km	4.2 km	9.5 km

Table A-15: Unweighted $L_{p,pk}$ impact areas and ranges for fish using the Popper et al (2014) pile driving criteria for the first hammer strike during the WTG foundation installation (alternative ramp-up) scenario at the Central (WT28) modelling location.

Popper et al. (2014) Unweighted $L_{p,pk}$		WTG foundation (first strike), alternative ramp-up			
		Area	Maximum range	Minimum range	Mean range
Pile driving	213 dB	0.01 km ²	60 m	60 m	60 m
	207 dB	0.07 km ²	150 m	140 m	150 m

Table A-16: Unweighted $L_{p,pk}$ impact areas and ranges for marine mammals using the Southall et al. (2019) impulsive criteria for the first hammer strike during the WTG foundation installation (alternative ramp-up) scenario at the South west (WT53) modelling location.

Southall et al. (2019) Unweighted $L_{p,pk}$		WTG foundation (first strike), alternative ramp-up			
		Area	Maximum range	Minimum range	Mean range
PTS (Impulsive)	LF (219 dB)	< 0.01 km ²	< 50 m	< 50 m	< 50 m
	HF (230 dB)	< 0.01 km ²	< 50 m	< 50 m	< 50 m
	VHF (202 dB)	0.2 km ²	270 m	240 m	250 m
	PCW (218 dB)	< 0.01 km ²	< 50 m	< 50 m	< 50 m
TTS (Impulsive)	LF (213 dB)	< 0.01 km ²	50 m	50 m	50 m
	HF (224 dB)	< 0.01 km ²	< 50 m	< 50 m	< 50 m
	VHF (196 dB)	1.1 km ²	630 m	520 m	580 m
	PCW (212 dB)	0.01 km ²	60 m	60 m	60 m

Table A-17: Unweighted $L_{p,RMS}$ impact areas and ranges for marine mammals using the NOAA (2005) impulsive behavioural disturbance criteria for the first hammer strike during the WTG foundation installation (alternative ramp-up) scenario at the South west (WT53) modelling location.

NOAA (2005) Unweighted $L_{p,RMS}$		WTG foundation (first strike), alternative ramp-up			
		Area	Maximum range	Minimum range	Mean range
Level B	160 dB	320 km ²	11 km	3.7 km	9.9 km

Table A-18: Unweighted $L_{p,pk}$ impact areas and ranges for fish using the Popper et al. (2014) pile driving criteria for the first hammer strike during the WTG foundation installation (alternative ramp-up) scenario at the South west (WT53) modelling location.

Popper et al. (2014) Unweighted $L_{p,pk}$		WTG foundation (first strike), alternative ramp-up			
		Area	Maximum range	Minimum range	Mean range
Pile driving	213 dB	< 0.01 km ²	50 m	50 m	50 m
	207 dB	0.05 km ²	130 m	120 m	130 m

Mitigation

Table A-19: Unweighted $L_{p,pk}$ impact areas and ranges for marine mammals using the Southall et al. (2019) impulsive criteria for the first hammer strike during the mitigated WTG foundation installation scenario at the North (WT03) modelling location.

Southall et al. (2019) Unweighted $L_{p,pk}$		WTG foundation (first strike), mitigated			
Southall et al. (2019) Weighted-SEL _{cum}		Area	Maximum range	Minimum range	Mean range
PTS (Impulsive)	LF (199-219 dB)	< 0.4-0.01 km ²	< 400-50 m	< 400-50 m	< 400-50 m
	HF (198-230 dB)	< 0.4-0.01 km ²	< 400-50 m	< 400-50 m	< 400-50 m
	VHF (202 dB)	0.16 km ²	230 m	220 m	220 m
	VHF-PCW (173-218 dB)	< 0.4-0.01 km ²	< 400-50 m	< 400-50 m	< 400-50 m
TTS (Impulsive)	PCW-LF (204-213 dB)	< 0.4-0.01 km ²	< 400-50 m	< 400-50 m	< 400-50 m
	HF (224 dB)	< 0.01 km ²	< 50 m	< 50 m	< 50 m
	VHF (196 dB)	0.83 km ²	530 m	500 m	510 m

Southall et al. (2019) Unweighted $L_{p,pk}$		WTG foundation (first strike), mitigated			
Southall et al. (2019) Weighted SEL_{cum}		Area	Maximum range	Minimum range	Mean range
PCW (212 dB)		< 0.01 km ²	50 m	50 m	50 m

Table A-20: Unweighted $L_{p,RMS}$ impact areas and ranges for marine mammals using the NOAA (2005) impulsive behavioural disturbance criteria for the first hammer strike during the mitigated WTG foundation installation scenario at the North (WT03) modelling location.

NOAA (2005) Unweighted $L_{p,RMS}$		WTG foundation (first strike), mitigated			
Level B		Area	Maximum range	Minimum range	Mean range
160 dB		180 km ²	9.1 km	4.3 km	7.4 km

Table A-21: Unweighted $L_{p,pk}$ impact areas and ranges for fish using the Popper et al. (2014) pile driving criteria for the first hammer strike during the mitigated WTG foundation installation scenario at the North (WT03) modelling location.

Popper et al. (2014) Unweighted $L_{p,pk}$		WTG foundation (first strike), mitigated			
Pile driving		Area	Maximum range	Minimum range	Mean range
213 dB		< 0.01 km ²	50 m	50 m	50 m
207 dB		0.04 km ²	110 m	110 m	110 m

Table A-22: Unweighted $L_{p,pk}$ impact areas and ranges for marine mammals using the Southall et al. (2019) impulsive criteria for the first hammer strike during the mitigated WTG foundation installation scenario at the Central (WT28) modelling location.

Southall et al. (2019) Unweighted $L_{p,pk}$		WTG foundation (first strike), mitigated			
		Area	Maximum range	Minimum range	Mean range
PTS (Impulsive)	LF (219 dB)	< 0.01 km ²	< 50 m	< 50 m	< 50 m
	HF (230 dB)	< 0.01 km ²	< 50 m	< 50 m	< 50 m
	VHF (202 dB)	0.19 km ²	260 m	240 m	250 m
	PCW (218 dB)	< 0.01 km ²	< 50 m	< 50 m	< 50 m
TTS (Impulsive)	LF (213 dB)	< 0.01 km ²	50 m	50 m	50 m
	HF (224 dB)	< 0.01 km ²	< 50 m	< 50 m	< 50 m
	VHF (196 dB)	1.0 km ²	610 m	530 m	580 m
	PCW (212 dB)	0.01 km ²	60 m	60 m	60 m

Table A-23: Unweighted $L_{p,RMS}$ impact areas and ranges for marine mammals using the NOAA (2005) impulsive behavioural disturbance criteria for the first hammer strike during the mitigated WTG foundation installation scenario at the Central (WT28) modelling location.

NOAA (2005) Unweighted $L_{p,RMS}$		WTG foundation (first strike), mitigated			
Level B		Area	Maximum range	Minimum range	Mean range
160 dB		210 km ²	11 km	3.8 km	7.9 km

Table A-24: Unweighted $L_{p,pk}$ impact areas and ranges for fish using the Popper et al. (2014) pile driving criteria for the first hammer strike during the mitigated WTG foundation installation scenario at the Central (WT28) modelling location.

Popper et al. (2014) Unweighted $L_{p,pk}$		WTG foundation (first strike), mitigated			
Pile driving		Area	Maximum range	Minimum range	Mean range
213 dB		< 0.01 km ²	50 m	50 m	50 m
207 dB		0.05 km ²	120 m	120 m	120 m

Table A-25: Unweighted $L_{p,pk}$ impact areas and ranges for marine mammals using the Southall et al. (2019) impulsive criteria for the first hammer strike during the mitigated WTG foundation installation scenario at the South west (WT53) modelling location.

Southall et al. (2019) Unweighted $L_{p,pk}$		WTG foundation (first strike), mitigated			
		Area	Maximum range	Minimum range	Mean range
PTS (Impulsive)	LF (219 dB)	< 0.01 km ²	< 50 m	< 50 m	< 50 m
	HF (230 dB)	< 0.01 km ²	< 50 m	< 50 m	< 50 m
	VHF (202 dB)	0.14 km ²	220 m	200 m	210 m
	PCW (218 dB)	< 0.01 km ²	< 50 m	< 50 m	< 50 m
TTS (Impulsive)	LF (213 dB)	< 0.01 km ²	50 m	< 50 m	50 m
	HF (224 dB)	< 0.01 km ²	< 50 m	< 50 m	< 50 m
	VHF (196 dB)	0.74 km ²	530 m	440 m	480 m
	PCW (212 dB)	< 0.01 km ²	50 m	50 m	50 m

Table A-26: Unweighted $L_{p,RMS}$ impact areas and ranges for marine mammals using the NOAA (2005) impulsive behavioural disturbance criteria for the first hammer strike during the mitigated WTG foundation installation scenario at the South west (WT53) modelling location.

NOAA (2005) Unweighted $L_{p,RMS}$		WTG foundation (first strike), mitigated			
		Area	Maximum range	Minimum range	Mean range
Level B	160 dB	230 km ²	9.4 km	3.4 km	8.4 km

Table A-27: Unweighted $L_{p,pk}$ impact areas and ranges for fish using the Popper et al (2014) pile driving criteria for the first hammer strike during the mitigated WTG foundation installation scenario at the South west (WT53) modelling location.

Popper et al. (2014) Unweighted $L_{p,pk}$		WTG foundation (first strike), mitigated			
		Area	Maximum range	Minimum range	Mean range
Pile driving	213 dB	< 0.01 km ²	50 m	< 50 m	50 m
	207 dB	0.03 km ²	110 m	100 m	100 m

A.1.2 OSP foundations

Table A-28: Unweighted $L_{p,pk}$ impact areas and ranges for marine mammals using the Southall et al (2019) impulsive criteria for the first hammer strike during the OSP foundation installation scenario at the North OSP modelling location.

Southall et al. (2019) Unweighted $L_{p,pk}$		OSP foundation (first strike)			
		Area	Maximum range	Minimum range	Mean range
PTS (Impulsive)	LF (219 dB)	< 0.01 km ²	< 50 m	< 50 m	< 50 m
	HF (230 dB)	< 0.01 km ²	< 50 m	< 50 m	< 50 m
	VHF (202 dB)	0.39 km ²	360 m	340 m	350 m
	PCW (218 dB)	< 0.01 km ²	< 50 m	< 50 m	< 50 m
TTS (Impulsive)	LF (213 dB)	0.02 km ²	80 m	80 m	80 m
	HF (224 dB)	< 0.01 km ²	< 50 m	< 50 m	< 50 m
	VHF (196 dB)	1.9 km ²	810 m	730 m	780 m
	PCW (212 dB)	0.02 km ²	90 m	90 m	90 m

Table A-29: Unweighted $L_{p,RMS}$ impact areas and ranges for marine mammals using the NOAA (2005) impulsive behavioural disturbance criteria for the first hammer strike during the OSP foundation installation scenario at the North OSP modelling location.

NOAA (2005) Unweighted $L_{p,RMS}$		OSP foundation (first strike)			
		Area	Maximum range	Minimum range	Mean range
Level B	160 dB	320 km ²	12 km	4.8 km	9.9 km

Table A-30: Unweighted $L_{p,pk}$ impact areas and ranges for fish using the Popper et al. (2014) pile driving criteria for the first hammer strike during the WTG foundation installation scenario at the North OSP modelling location.

Popper et al. (2014) Unweighted $L_{p,pk}$	OSP foundation (first strike)				
	Area	Maximum range	Minimum range	Mean range	
Pile driving	213 dB	0.02 km ²	80 m	80 m	80 m
	207 dB	0.1 km ²	180 m	170 m	180 m

Table A-31: Unweighted $L_{p,pk}$ impact areas and ranges for marine mammals using the Southall et al. (2019) impulsive criteria for the first hammer strike during the WTG foundation installation scenario at the South OSP modelling location.

Southall et al. (2019) Unweighted $L_{p,pk}$	OSP foundation (first strike)				
	Area	Maximum range	Minimum range	Mean range	
PTS (Impulsive)	LF (219 dB)	< 0.01 km ²	< 50 m	< 50 m	< 50 m
	HF (230 dB)	< 0.01 km ²	< 50 m	< 50 m	< 50 m
	VHF (202 dB)	0.18 km ²	260 m	220 m	240 m
	PCW (218 dB)	< 0.01 km ²	< 50 m	< 50 m	< 50 m
TTS (Impulsive)	LF (213 dB)	0.01 km ²	60 m	60 m	60 m
	HF (224 dB)	< 0.01 km ²	< 50 m	< 50 m	< 50 m
	VHF (196 dB)	0.85 km ²	610 m	460 m	520 m
	PCW (212 dB)	0.01 km ²	70 m	60 m	70 m

Table A-32: Unweighted $L_{p,RMS}$ impact areas and ranges for marine mammals using the NOAA (2005) impulsive behavioural disturbance criteria for the first hammer strike during the OSP foundation installation scenario at the South OSP modelling location.

NOAA (2005) Unweighted $L_{p,RMS}$	OSP foundation (first strike)				
	Area	Maximum range	Minimum range	Mean range	
ITS Level B	LF (179) 160 dB)	330 260 km ²	18 11 km	1.4 km	8.4 8.9 km

Table A-33: Unweighted $L_{p,pk}$ impact areas and ranges for fish using the Popper et al (2014) pile driving criteria for the first hammer strike during the OSP foundation installation scenario at the South OSP modelling location.

Popper et al. (2014) Unweighted $L_{p,pk}$	OSP foundation (first strike)				
	Area	Maximum range	Minimum range	Mean range	
Pile driving	213 dB	0.01 km ²	60 m	60 m	60 m
	207 dB	0.05 km ²	130 m	120 m	130 m

Alternative ramp-up

Table A-34: Unweighted $L_{p,pk}$ impact areas and ranges for marine mammals using the Southall et al. (2019) impulsive criteria for the first hammer strike during the OSP foundation installation (alternative ramp-up) scenario at the North OSP modelling location.

Southall et al. (2019) Unweighted $L_{p,pk}$	OSP foundation (first strike), alternative ramp-up				
	Area	Maximum range	Minimum range	Mean range	
PTS (Impulsive)	LF (219 dB)	< 0.01 km ²	< 50 m	< 50 m	< 50 m
	HF (178 230 dB)	< 0.1 0.01 km ²	< 100 50 m	< 100 50 m	< 100 50 m
	VHF (153 202 dB)	7 0.18 km ²	7.1 km250 m	1.6 km240 m	4.4 km240 m
	PCW (181 218 dB)	0.7 < 0.01 km ²	800 < 50 m	< 100 50 m	400 < 50 m
TTS (Impulsive)	LF (213 dB)	< 0.01 km ²	50 m	50 m	50 m
	HF (224 dB)	< 0.01 km ²	< 50 m	< 50 m	< 50 m
	VHF (196 dB)	0.93 km ²	560 m	520 m	540 m
	PCW (212 dB)	0.01 km ²	60 m	60 m	60 m

Table A-35: Unweighted $L_{p,RMS}$ impact areas and ranges for marine mammals using the NOAA (2005) impulsive behavioural disturbance criteria for the first hammer strike during the OSP foundation installation (alternative ramp-up) scenario at the North OSP modelling location.

NOAA (2005) Unweighted $L_{p,RMS}$		OSP foundation (first strike), alternative ramp-up			
Level B	160 dB	Area	Maximum range	Minimum range	Mean range
		210 km ²	9.6 km	4.1 km	8.0 km

Table A-36: Unweighted $L_{p,pk}$ impact areas and ranges for fish using the Popper et al. (2014) pile driving criteria for the first hammer strike during the OSP foundation installation (alternative ramp-up) scenario at the North OSP modelling location.

Popper et al. (2014) Unweighted $L_{p,pk}$		OSP foundation (first strike), alternative ramp-up			
Pile driving	213 dB	Area	Maximum range	Minimum range	Mean range
	207 dB	< 0.01 km ²	50 m	50 m	50 m
		0.05 km ²	120 m	120 m	120 m

Table A-37: Unweighted $L_{p,pk}$ impact areas and ranges for marine mammals using the Southall et al. (2019) impulsive criteria for the first hammer strike during the OSP foundation installation (alternative ramp-up) scenario at the South OSP modelling location.

Southall et al. (2019) Unweighted $L_{p,pk}$		OSP foundation (first strike), alternative ramp-up			
		Area	Maximum range	Minimum range	Mean range
PTS (Impulsive)	LF (219 dB)	< 0.01 km ²	< 50 m	< 50 m	< 50 m
	HF (230 dB)	< 0.01 km ²	< 50 m	< 50 m	< 50 m
	VHF (202 dB)	0.09 km ²	180 m	160 m	170 m
	PCW (218 dB)	< 0.01 km ²	< 50 m	< 50 m	< 50 m
TTS (Impulsive)	LF (213 dB)	< 0.01 km ²	< 50 m	< 50 m	< 50 m
	HF (224 dB)	< 0.01 km ²	< 50 m	< 50 m	< 50 m
	VHF (196 dB)	0.42 km ²	410 m	330 m	360 m
	PCW (212 dB)	< 0.01 km ²	< 50 m	< 50 m	< 50 m

Table A-38: Unweighted $L_{p,RMS}$ impact areas and ranges for marine mammals using the NOAA (2005) impulsive behavioural disturbance criteria for the first hammer strike during the OSP foundation installation (alternative ramp-up) scenario at the South OSP modelling location.

NOAA (2005) Unweighted $L_{p,RMS}$		OSP foundation (first strike), alternative ramp-up			
Level B	160 dB	Area	Maximum range	Minimum range	Mean range
		160 km ²	8.0 km	1.4 km	7.0 km

Table A-39: Unweighted $L_{p,pk}$ impact areas and ranges for fish using the Popper et al. (2014) pile driving criteria for the first hammer strike during the OSP foundation installation (alternative ramp-up) scenario at the South OSP modelling location.

Popper et al. (2014) Unweighted $L_{p,pk}$		OSP foundation (first strike), alternative ramp-up			
Pile driving	213 dB	Area	Maximum range	Minimum range	Mean range
	207 dB	< 0.01 km ²	< 50 m	< 50 m	< 50 m
		0.02 km ²	90 m	90 m	90 m

Mitigation

~~Table A-2 Summary of the weighted SEL_{cum}-40: Unweighted L_{p,pk} impact areas and ranges for marine mammals using the Southall et al. (2019) non-impulsive criteria for the 7-m monopile first hammer strike during the mitigated OSP foundation modelling installation scenario at the NW-WTG North OSP modelling location assuming a fleeing animal.~~

Southall et al. (2019) Unweighted L _{p,pk}		OSP foundation (first strike), mitigated			
		Area	Maximum range	Minimum range	Mean range
PTS (Impulsive)	LF (219 dB)	< 0.01 km ²	< 50 m	< 50 m	< 50 m
	HF (230 dB)	< 0.01 km ²	< 50 m	< 50 m	< 50 m
	VHF (202 dB)	0.13 km ²	210 m	200 m	200 m
	PCW (218 dB)	< 0.01 km ²	< 50 m	< 50 m	< 50 m
TTS (Impulsive)	LF (213 dB)	< 0.01 km ²	< 50 m	< 50 m	< 50 m
	HF (224 dB)	< 0.01 km ²	< 50 m	< 50 m	< 50 m
	VHF (196 dB)	0.66 km ²	470 m	440 m	460 m
	PCW (212 dB)	< 0.01 km ²	50 m	50 m	50 m

Table A-41: Unweighted L_{p,RMS} impact areas and ranges for marine mammals using the NOAA (2005) impulsive behavioural disturbance criteria for the first hammer strike during the mitigated OSP foundation installation scenario at the North OSP modelling location.

NOAA (2005) Unweighted L _{p,RMS}		OSP foundation (first strike), mitigated			
		Area	Maximum range	Minimum range	Mean range
Level B	160 dB	150 km ²	8.3 km	3.7 km	6.9 km

Table A-42: Unweighted L_{p,pk} impact areas and ranges for fish using the Popper et al (2014) pile driving criteria for the first hammer strike during the mitigated OSP foundation installation scenario at the North OSP modelling location.

Popper et al. (2014) Unweighted L _{p,pk}		OSP foundation (first strike), mitigated			
		Area	Maximum range	Minimum range	Mean range
Pile driving	213 dB	< 0.01 km ²	< 50 m	< 50 m	< 50 m
	207 dB	0.03 km ²	100 m	100 m	100 m

Table A-43: Unweighted L_{p,pk} impact areas and ranges for marine mammals using the Southall et al. (2019) impulsive criteria for the first hammer strike during the mitigated OSP foundation installation scenario at the South OSP modelling location.

Southall et al. (2019) Unweighted L _{p,pk}		OSP foundation (first strike), mitigated			
		Area	Maximum range	Minimum range	Mean range
PTS (Impulsive)	LF (219 dB)	< 0.01 km ²	< 50 m	< 50 m	< 50 m
	HF (230 dB)	< 0.01 km ²	< 50 m	< 50 m	< 50 m
	VHF (202 dB)	0.06 km ²	150 m	140 m	140 m
	PCW (218 dB)	< 0.01 km ²	< 50 m	< 50 m	< 50 m
TTS (Impulsive)	LF (213 dB)	< 0.01 km ²	< 50 m	< 50 m	< 50 m
	HF (224 dB)	< 0.01 km ²	< 50 m	< 50 m	< 50 m
	VHF (196 dB)	0.3 km ²	350 m	290 m	310 m
	PCW (212 dB)	< 0.01 km ²	< 50 m	< 50 m	< 50 m

Table A-44: Unweighted L_{p,RMS} impact areas and ranges for marine mammals using the NOAA (2005) impulsive behavioural disturbance criteria for the first hammer strike during the mitigated OSP foundation installation scenario at the South OSP modelling location.

NOAA (2005) Unweighted L _{p,RMS}		OSP foundation (first strike), mitigated			
		Area	Maximum range	Minimum range	Mean range
Level B	160 dB	110 km ²	6.7 km	1.4 km	5.9 km

Table A-45: Unweighted $L_{p,pk}$ impact areas and ranges for fish using the Popper et al (2014) pile driving criteria for the first hammer strike during the mitigated OSP foundation installation scenario at the South OSP modelling location.

Popper et al. (2014) Unweighted $L_{p,pk}$		OSP foundation (first strike), mitigated			
		Area	Maximum range	Minimum range	Mean range
Pile driving	213 dB	< 0.01 km ²	< 50 m	< 50 m	< 50 m
	207 dB	0.02 km ²	80 m	70 m	70 m

A.2 Non-impulsive criteria

Following sections 2.2.1 and 2.3.1 where the principles of impulsive and non-impulsive noise are explained, Table A-46 to Table A-60 present the modelled impact piling noise in terms of the non-impulsive $L_{E,p,t}$ criteria from Southall et al. (2019). $L_{p,pk}$ criteria are not suitable for non-impulsive consideration so are not included. Note that although the impact ranges below have been calculated to the non-impulsive thresholds, in practice they should only be considered where the impact range is in excess of 5 km.

A.2.1 WTG foundations

Table A-46: Weighted $L_{E,p,24h,wtg}$ impact areas and ranges for marine mammals using the Southall et al (2019) non-impulsive criteria for the WTG installation scenario at the North (WT03) modelling location assuming fleeing receptors.

Southall et al. (2019) Weighted- SEL_{cum} / Weighted $L_{E,p,24h,wtg}$		Area Maximum range Minimum range Mean range WTG foundation			
		Area	Maximum range	Minimum range	Mean range
PTS (Non-impulsive)	LF (199 dB)	< 0.4 km ²	< 100 m	< 100 m	< 100 m
	HF (198 dB)	< 0.4 km ²	< 100 m	< 100 m	< 100 m
	VHF (173 dB)	< 0.4 km ²	< 100 m	< 100 m	< 100 m
	PCW (201 dB)	< 0.4 km ²	< 100 m	< 100 m	< 100 m
TTS (Non-impulsive)	LF (179 dB)	330.160 km ²	1813 km	1.4 km	260 m
	HF (178 dB)	< 0.4 km ²	< 100 m	< 100 m	< 100 m
	VHF (153 dB)	728.1 km ²	7.12.5 km	1.6 km	< 50 m
	PCW (181 dB)	0.7 < 0.01 km ²	800 < 50 m	< 100 m	< 50 m

Table A-47: Weighted SEL_{cum} $L_{E,p,24h,wtg}$ impact areas and ranges for marine mammals using the Southall et al- (2019) non-impulsive criteria for the 11 m monopile foundation modelling at the Central (WT28) modelling location assuming a fleeing animal receptors.

Southall et al. (2019) Weighted- SEL_{cum} / Weighted $L_{E,p,24h,wtg}$		Area Maximum range Minimum range Mean range WTG foundation			
		Area	Maximum range	Minimum range	Mean range
PTS (Non-impulsive)	LF (199 dB)	< 0.4 km ²	< 100 m	< 100 m	< 100 m
	HF (198 dB)	< 0.4 km ²	< 100 m	< 100 m	< 100 m
	VHF (173 dB)	< 0.4 km ²	< 100 m	< 100 m	< 100 m
	PCW (201 dB)	< 0.4 km ²	< 100 m	< 100 m	< 100 m
TTS	LF (179 dB)	1,300.350 km ²	3217 km	5.1 km	330 m

(Non-impulsive)	HF (178 dB)	< 0.4 0.01 km ²	< 100 50 m	< 100 50 m	< 100 50 m
	VHF (153 dB)	280 20 km ²	13 4.1 km	4.7 km < 50 m	9.0 1.9 km
	PCW (181 dB)	28 < 0.01 km ²	3.9 km < 70 m	1.7 km < 50 m	2.9 km < 50 m

Table A-4 Summary of the -48: Weighted $SEL_{cum}L_{E,p,24h,wt}$ impact areas and ranges for marine mammals using the Southall et al- (2019) non-impulsive criteria for the 7-m monopile foundation modelling at the C WTG installation scenario at the South west (WT53) modelling location assuming a-fleeing animal receptors.

Southall et al. (2019) Weighted- SEL_{cum} Weighted $L_{E,p,24h,wt}$		Area Maximum range Minimum range Mean range WTG foundation			
		Area	Maximum range	Minimum range	Mean range
PTS (Non-impulsive)	LF (199 dB)	< 0.4 km ²	< 100 m	< 100 m	< 100 m
	HF (198 dB)	< 0.4 km ²	< 100 m	< 100 m	< 100 m
	VHF (173 dB)	< 0.4 km ²	< 100 m	< 100 m	< 100 m
	PCW (201 dB)	< 0.4 km ²	< 100 m	< 100 m	< 100 m
TTS (Non-impulsive)	LF (179 dB)	1,300 km ²	290 km	3215 m	5.1 km
	HF (178 dB)	< 0.4 km ²	< 100 m	< 100 m	< 100 m
	VHF (153 dB)	280 km ²	22 km	133.8 km	4.7 km
	PCW (181 dB)	28 < 0.01 km ²	3.9 km	50 m	1.7 km

Alternative ramp-up

Table A-49: Weighted $L_{E,p,24h,wt}$ impact areas and ranges for marine mammals using the Southall et al (2019) non-impulsive criteria for the WTG installation (alternative ramp-up) scenario at the North (WT03) modelling location assuming fleeing receptors.

Southall et al. (2019) Weighted $L_{E,p,24h,wt}$		WTG foundation (alternative ramp-up)			
		Area	Maximum range	Minimum range	Mean range
PTS (Non-impulsive)	LF (199 dB)	< 0.01 km ²	< 50 m	< 50 m	< 50 m
	HF (198 dB)	< 0.01 km ²	< 50 m	< 50 m	< 50 m
	VHF (173 dB)	< 0.01 km ²	< 50 m	< 50 m	< 50 m
	PCW (201 dB)	< 0.01 km ²	< 50 m	< 50 m	< 50 m
TTS (Non-impulsive)	LF (179 dB)	22 km ²	5.4 km	130 m	1.7 km
	HF (178 dB)	< 0.01 km ²	< 50 m	< 50 m	< 50 m
	VHF (153 dB)	< 0.01 km ²	< 50 m	< 50 m	< 50 m
	PCW (181 dB)	< 0.01 km ²	< 50 m	< 50 m	< 50 m

Table A-5 Summary of the -50: Weighted $SEL_{cum}L_{E,p,24h,wt}$ impact areas and ranges for marine mammals using the Southall et al- (2019) non-impulsive criteria for the 11-m monopile foundation modelling at the SW-WTG installation scenario (alternative ramp-up) at the Central (WT28) modelling location assuming a-fleeing animal receptors.

Southall et al. (2019) Weighted- SEL_{cum} Weighted $L_{E,p,24h,wt}$		Area Maximum range Minimum range Mean range WTG foundation (alternative ramp-up)			
		Area	Maximum range	Minimum range	Mean range
PTS (Non-impulsive)	LF (199 dB)	< 0.4 km ²	< 100 m	< 100 m	< 100 m
	HF (198 dB)	< 0.4 km ²	< 100 m	< 100 m	< 100 m
	VHF (173 dB)	< 0.4 km ²	< 100 m	< 100 m	< 100 m
	PCW (201 dB)	< 0.4 km ²	< 100 m	< 100 m	< 100 m
TTS (Non-impulsive)	LF (179 dB)	1,600 km ²	97 km	349.7 km	5.7 km
	HF (178 dB)	< 0.4 km ²	< 100 m	< 100 m	< 100 m

	VHF (153 dB)	360 0.21 km ²	15 km 570 m	5.2 km < 50 m	10 km 160 m
	PCW (181 dB)	29 < 0.01 km ²	4.0 km < 50 m	1.7 km < 50 m	2.9 km < 50 m

Table A-6 Summary of the -51: Weighted $SEL_{cumLE,p,24h,wt}$ impact areas and ranges for marine mammals using the Southall et al- (2019) non-impulsive criteria for the 7-m monopile foundation modelling at the SW WTG installation scenario (alternative ramp-up) at the South west (WT53) modelling location assuming a fleeing animal receptors.

Southall et al. (2019) Weighted SEL_{cum} / Weighted $L_{E,p,24h,wt}$		Area Maximum range Minimum range Mean range WTG foundation (alternative ramp-up)			
		Area	Maximum range	Minimum range	Mean range
PTS (Non-impulsive)	LF (199 dB)	< 0.4 km ²	< 400 m	< 400 m	< 400 m
	HF (198 dB)	< 0.4 km ²	< 400 m	< 400 m	< 400 m
	VHF (173 dB)	< 0.4 km ²	< 400 m	< 400 m	< 400 m
	PCW (201 dB)	< 0.4 km ²	< 400 m	< 400 m	< 400 m
TTS (Non-impulsive)	LF (179 dB)	1,500 km ²	348.0 km	5.6 km	120 m
	HF (178 dB)	< 0.4 km ²	< 400 m	< 400 m	< 400 m
	VHF (153 dB)	350 km ²	15 km	5.2 km	< 50 m
	PCW (181 dB)	28 < 0.01 km ²	4.0 km	1.7 km	< 50 m

Mitigation

Table A-7 Summary of the -52: Weighted $SEL_{cumLE,p,24h,wt}$ impact areas and ranges for marine mammals using the Southall et al- (2019) non-impulsive criteria for the 14-m monopile foundation modelling mitigated WTG installation scenario at the North OSP (WT03) modelling location assuming a fleeing animal receptors.

Southall et al. (2019) Weighted SEL_{cum} / Weighted $L_{E,p,24h,wt}$		Area Maximum range Minimum range Mean range WTG foundation, mitigated			
		Area	Maximum range	Minimum range	Mean range
PTS (Non-impulsive)	LF (199 dB)	< 0.4 km ²	< 400 m	< 400 m	< 400 m
	HF (198 dB)	< 0.4 km ²	< 400 m	< 400 m	< 400 m
	VHF (173 dB)	< 0.4 km ²	< 400 m	< 400 m	< 400 m
	PCW (201 dB)	< 0.4 km ²	< 400 m	< 400 m	< 400 m
TTS (Non-impulsive)	LF (179 dB)	490 km ²	214.1 km	2.0 km	100 m
	HF (178 dB)	< 0.4 km ²	< 400 m	< 400 m	< 400 m
	VHF (153 dB)	100 < 0.01 km ²	8.2 km	2.1 km	< 50 m
	PCW (181 dB)	3.4 < 0.01 km ²	1.4 km	200	< 50 m

Table A-53: Weighted $LE_{p,24h,wt}$ impact areas and ranges for marine mammals using the Southall et al (2019) non-impulsive criteria for the mitigated WTG installation scenario at the Central (WT28) modelling location assuming fleeing receptors.

Southall et al. (2019) Weighted $L_{E,p,24h,wt}$		WTG foundation, mitigated			
		Area	Maximum range	Minimum range	Mean range
PTS (Non-impulsive)	LF (199 dB)	< 0.01 km ²	< 50 m	< 50 m	< 50 m
	HF (198 dB)	< 0.01 km ²	< 50 m	< 50 m	< 50 m
	VHF (173 dB)	< 0.01 km ²	< 50 m	< 50 m	< 50 m
	PCW (201 dB)	< 0.01 km ²	< 50 m	< 50 m	< 50 m
TTS (Non-impulsive)	LF (179 dB)	57 km ²	7.1 km	110 m	3.0 km
	HF (178 dB)	< 0.01 km ²	< 50 m	< 50 m	< 50 m
	VHF (153 dB)	0.01 km ²	120 m	< 50 m	50 m
	PCW (181 dB)	< 0.01 km ²	< 50 m	< 50 m	< 50 m

Table A-54: Weighted $L_{E,p,24h,wtd}$ impact areas and ranges for marine mammals using the Southall et al (2019) non-impulsive criteria for the mitigated WTG installation scenario at the South west (WT53) modelling location assuming fleeing receptors.

Southall et al. (2019) Weighted $L_{E,p,24h,wtd}$		WTG foundation, mitigated			
		Area	Maximum range	Minimum range	Mean range
PTS (Non-impulsive)	LF (199 dB)	< 0.01 km ²	< 50 m	< 50 m	< 50 m
	HF (198 dB)	< 0.01 km ²	< 50 m	< 50 m	< 50 m
	VHF (173 dB)	< 0.01 km ²	< 50 m	< 50 m	< 50 m
	PCW (201 dB)	< 0.01 km ²	< 50 m	< 50 m	< 50 m
TTS (Non-impulsive)	LF (179 dB)	37 km ²	5.8 km	90 m	2.8 km
	HF (178 dB)	< 0.01 km ²	< 50 m	< 50 m	< 50 m
	VHF (153 dB)	< 0.01 km ²	70 m	< 50 m	< 50 m
	PCW (181 dB)	< 0.01 km ²	< 50 m	< 50 m	< 50 m

A.2.2 OSP foundations

Table A-8 Summary of the -55: Weighted $SEL_{cum}L_{E,p,24h,wtd}$ impact areas and ranges for marine mammals using the Southall et al- (2019) non-impulsive criteria for the 7-m monopile foundation modelling OSP installation scenario at the North OSP modelling location assuming a fleeing animal receptors.

Southall et al. (2019) Weighted SEL_{cum} Weighted $L_{E,p,24h,wtd}$		Area Maximum range Minimum range Mean range OSP foundation			
		Area	Maximum range	Minimum range	Mean range
PTS (Non-impulsive)	LF (199 dB)	< 0.4 0.01 km ²	< 100 50 m	< 100 50 m	< 100 50 m
	HF (198 dB)	< 0.4 0.01 km ²	< 100 50 m	< 100 50 m	< 100 50 m
	VHF (173 dB)	< 0.4 0.01 km ²	< 100 50 m	< 100 50 m	< 100 50 m
	PCW (201 dB)	< 0.4 0.01 km ²	< 100 50 m	< 100 50 m	< 100 50 m
TTS (Non-impulsive)	LF (179 dB)	480 110 km ²	21 11 km	1.9 km 230 m	10 4.6 km
	HF (178 dB)	< 0.4 0.01 km ²	< 100 50 m	< 100 50 m	< 100 50 m
	VHF (153 dB)	100 5.5 km ²	8.2 2.2 km	2.1 km < 50 m	5.4 1.1 km
	PCW (181 dB)	3.3 < 0.01 km ²	1.4 km < 50 m	200 < 50 m	900 < 50 m

Table A-9 Summary of the -56: Weighted $SEL_{cum}L_{E,p,24h,wtd}$ impact areas and ranges for marine mammals using the Southall et al- (2019) non-impulsive criteria for the 14-m monopile foundation modelling OSP installation scenario at the South OSP modelling location assuming a fleeing animal receptors.

Southall et al. (2019) Weighted SEL_{cum} Weighted $L_{E,p,24h,wtd}$		Area Maximum range Minimum range Mean range OSP foundation			
		Area	Maximum range	Minimum range	Mean range
PTS (Non-impulsive)	LF (199 dB)	< 0.4 0.01 km ²	< 100 50 m	< 100 50 m	< 100 50 m
	HF (198 dB)	< 0.4 0.01 km ²	< 100 50 m	< 100 50 m	< 100 50 m
	VHF (173 dB)	< 0.4 0.01 km ²	< 100 50 m	< 100 50 m	< 100 50 m
	PCW (201 dB)	< 0.4 0.01 km ²	< 100 50 m	< 100 50 m	< 100 50 m
TTS (Non-impulsive)	LF (179 dB)	1,500 47 km ²	337.0 km	5.6 km 120 m	193.1 km
	HF (178 dB)	< 0.4 0.01 km ²	< 100 50 m	< 100 50 m	< 100 50 m
	VHF (153 dB)	340 5.6 km ²	142.2 km	5.1 km < 50 m	9.9 1.2 km
	PCW (181 dB)	23 < 0.01 km ²	3.8 km < 50 m	1.5 km < 50 m	2.6 km < 50 m

[Alternative ramp-up](#)

Table A-10 Summary of the 57: Weighted $SEL_{cum} L_{E,p,24h,wtd}$ impact areas and ranges for marine mammals using the Southall et al. (2019) non-impulsive criteria for the 7-m monopile foundation modelling at the South OSP installation (alternative ramp-up) scenario at the North OSP modelling location assuming a fleeing animal receptors.

Southall et al. (2019) Weighted SEL_{cum} Weighted $L_{E,p,24h,wtd}$		Area Maximum range Minimum range Mean range OSP foundation (alternative ramp-up)			
		Area	Maximum range	Minimum range	Mean range
PTS (Non-impulsive)	LF (199 dB)	< 0.4 0.01 km ²	< 400 50 m	< 400 50 m	< 400 50 m
	HF (198 dB)	< 0.4 0.01 km ²	< 400 50 m	< 400 50 m	< 400 50 m
	VHF (173 dB)	< 0.4 0.01 km ²	< 400 50 m	< 400 50 m	< 400 50 m
	PCW (201 dB)	< 0.4 0.01 km ²	< 400 50 m	< 400 50 m	< 400 50 m
TTS (Non-impulsive)	LF (179 dB)	1,500 8.6 km ²	333.8 km	5.6 km 120 m	181.0 km
	HF (178 dB)	< 0.4 0.01 km ²	< 400 50 m	< 400 50 m	< 400 50 m
	VHF (153 dB)	340 < 0.01 km ²	14 km < 50 m	5.1 km < 50 m	9.9 km < 50 m
	PCW (181 dB)	23 < 0.01 km ²	3.7 km < 50 m	1.5 km < 50 m	2.6 km < 50 m

Table A-58: Weighted $L_{E,p,24h,wtd}$ impact areas and ranges for marine mammals using the Southall et al (2019) non-impulsive criteria for the OSP installation scenario (alternative ramp-up) at the South OSP modelling location assuming fleeing receptors.

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P352R0103					
P352R0104	VHF (173 dB)	< 0.01 km ²	< 50 m	< 50 m	< 50 m
P352R015	PCW (201 dB)	< 0.01 km ²	< 50 m	< 50 m	< 50 m
P352R0106	-				
TIS (Non-impulsive)	LF (179 dB)	0.29 km ²	840 m	80 m	240 m
	HF (178 dB)	< 0.01 km ²	< 50 m	< 50 m	< 50 m
	VHF (153 dB)	< 0.01 km ²	< 50 m	< 50 m	< 50 m
	PCW (181 dB)	< 0.01 km ²	< 50 m	< 50 m	< 50 m

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Title classification	Unclassified
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Abstract	
Abstract classification	Unclassified; Unlimited distribution

Mitigation

Table A-59: Weighted $L_{E,p,24h,wtd}$ impact areas and ranges for marine mammals using the Southall et al (2019) non-impulsive criteria for the mitigated OSP installation scenario at the North OSP modelling location assuming fleeing receptors.

Southall et al. (2019) Weighted $L_{E,p,24h,wtd}$		OSP foundation, mitigated			
		Area	Maximum range	Minimum range	Mean range
PTS (Non-impulsive)	LF (199 dB)	< 0.01 km ²	< 50 m	< 50 m	< 50 m
	HF (198 dB)	< 0.01 km ²	< 50 m	< 50 m	< 50 m
	VHF (173 dB)	< 0.01 km ²	< 50 m	< 50 m	< 50 m
	PCW (201 dB)	< 0.01 km ²	< 50 m	< 50 m	< 50 m
TTS (Non-impulsive)	LF (179 dB)	7.0 km ²	3.0 km	100 m	1.1 km
	HF (178 dB)	< 0.01 km ²	< 50 m	< 50 m	< 50 m
	VHF (153 dB)	< 0.01 km ²	< 50 m	< 50 m	< 50 m
	PCW (181 dB)	< 0.01 km ²	< 50 m	< 50 m	< 50 m

Table A-60: Weighted $L_{E,p,24h,wtg}$ impact areas and ranges for marine mammals using the Southall et al (2019) non-impulsive criteria for the mitigated OSP installation scenario at the South OSP modelling location assuming fleeing receptors.

Southall et al. (2019) Weighted $L_{E,p,24h,wtg}$		OSP foundation, mitigated			
		Area	Maximum range	Minimum range	Mean range
PTS (Non-impulsive)	LF (199 dB)	< 0.01 km ²	< 50 m	< 50 m	< 50 m
	HF (198 dB)	< 0.01 km ²	< 50 m	< 50 m	< 50 m
	VHF (173 dB)	< 0.01 km ²	< 50 m	< 50 m	< 50 m
	PCW (201 dB)	< 0.01 km ²	< 50 m	< 50 m	< 50 m
TTS (Non-impulsive)	LF (179 dB)	0.28 km ²	790 m	70 m	240 m
	HF (178 dB)	< 0.01 km ²	< 50 m	< 50 m	< 50 m
	VHF (153 dB)	< 0.01 km ²	< 50 m	< 50 m	< 50 m
	PCW (181 dB)	< 0.01 km ²	< 50 m	< 50 m	< 50 m